



Impact of wave exposure on bleaching of large benthic foraminifera

Claire E. Reymond^{1,2,4} · Caroline Romo⁴ · Gabija Posiunaite⁴ · Maria Byrne^{1,2,3} · Jody M. Webster⁴

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Abstract During the 2024 global mass bleaching event, a rapid bleaching assessment was conducted on two large benthic foraminifera populations on One Tree Reef, southern Great Barrier Reef. In tropical reef ecosystems, large benthic foraminifera are major carbonate sediment producers and function as important ecological engineers. We documented the thermal stress of *Marginopora vertebralis* and *Baculogypsina sphaerulata* from two high-density populations along the leeward and windward side of One Tree Reef in March 2024 following an 8-degree heating week with local on-reef temperatures exceeding 30 °C. Bleaching was more prevalent at the lower-energy leeward site (81.4% of *M. vertebralis* and 80.2% of *B. sphaerulata* individuals bleached, respectively) than the windward site (31.1% of *M. vertebralis* and 40.8% of *B. sphaerulata* bleached), suggesting localised hydrodynamic exposure has a significant impact on the health and bleaching susceptibility of benthic foraminifera. This study provides a rapid bleaching guide and emphasises the necessity of ongoing monitoring of benthic foraminifera

to understand local and regional impacts of climate-induced stressors on reef carbonate production.

Keywords Marine heat waves · Bleaching · Hydrodynamics · Reef monitoring

Introduction

Marine heatwaves are becoming more frequent and prolonged due to anthropogenic greenhouse gas emissions (Hobday et al. 2018; Schmidt et al. 2011). The global mass coral bleaching event in early 2024 marked the Great Barrier Reef's (GBR) seventh event since 1998 and the fifth since 2016. Satellite and aerial surveys in 2024 showed extreme ocean temperatures and extensive coral bleaching across all regions (north, central, and south) of the GBR (Cantin et al. 2024). The southern region, including the highly protected One Tree Reef (OTR), suffered severe coral bleaching and mortality following prolonged ocean warming exceeding 30 °C and 8-degree heating week (DHW) (Byrne et al. 2025). Since the major coral bleaching events in 2016 and 2017, the Southern Great Barrier Reef, including OTR, has not experienced significant bleaching, with only low-level bleaching (< 10% of colonies) observed in 2021 (Emslie 2021; Cantin et al. 2024).

More research is needed to assess the implications of marine heat waves on other non-coral marine carbonate producers, such as large benthic foraminifera (LBF). These organisms contribute significantly to the carbonate sediment budget within reef systems (Hallock 1981; Langer et al. 1997; Hohenegger 2006; Doo et al. 2017) and function as key ecological engineers (Dawson et al. 2014). The sand produced by LBF shapes coastal shorelines and habitats, and provides coastal stabilisation (Fellows et al 2017). In the southern

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✉ Claire E. Reymond
claire.reymond@sydney.edu.au

- ¹ School of Life and Environmental Sciences, The University of Sydney, Sydney, NSW 2006, Australia
- ² Marine Studies Institute, The University of Sydney, Sydney, NSW 2006, Australia
- ³ Marine Biology and Aquaculture, The College of Science and Engineering, James Cook University Townsville, Townsville, QLD 4811, Australia
- ⁴ Geocoastal Research Group, School of Geosciences, The University of Sydney, Sydney, NSW 2006, Australia

Great Barrier Reef, LBF contribute approximately $3 \text{ kg CaCO}_3 \text{ m}^{-2} \text{ y}^{-1}$ and 32% of the reef sediments in the sand aprons in OTR (Doo et al. 2014; Fellowes et al. 2017). Populations of *Marginopora* in the subtropical northwest Pacific contribute approximately $5 \text{ kg CaCO}_3 \text{ m}^{-2} \text{ yr}^{-1}$ to the reef sediments (Fujita et al. 2000), which is higher than the estimated global average of $0.03\text{--}1 \text{ kg CaCO}_3 \text{ m}^{-2} \text{ y}^{-1}$ (Langer et al. 1997). Other studies, such as Hallock (1981) and Doo et al. (2017), have provided similar estimates for different species of foraminifera, highlighting the significant contribution of these organisms to carbonate production in reef ecosystems. In the western Pacific, studies on LBF have reported an average lifespan of 17 months and the presence of multiple modes of reproduction (Fujita et al. 2000; Hohenegger 2006; Hosono et al. 2013; Hohenegger et al. 2019). Despite their importance, bleaching patterns of LBF are less understood than other photosymbiotic reef organisms (Narayan et al. 2022). This highlights the need to include LBF in long-term and continuous reef health initiatives.

The capacity of LBF to prolifically produce calcium carbonate exoskeleton (CaCO_3 tests) is in part due to their obligatory symbioses with photosynthetic microalgae, such as diatoms, rhodophytes, chlorophytes, and/or zooxanthellae-dinoflagellates (Lee and Hallock 1987; Lee 2006). Within a single LBF species, these photosymbionts have high genetic

diversity (Momigliano and Uthicke 2013; Prazeres et al. 2021; Stuhr et al. 2021), enhancing their potential adaptability to environmental stressors (Stuhr et al. 2017, 2018).

Bleaching in foraminifera was initially documented in *Amphistegina gibbosa* from the Florida Keys (Hallock et al. 1993; Hallock 2000), and here, we present the first report of bleaching in LBF in the GBR. Experimental studies suggest that the thermal threshold for optimal physiological function in LBF is $30 \text{ }^\circ\text{C}$ (Talge and Hallock 1995, 2003; Fujita et al. 2014; Schmidt et al. 2014; Stuhr et al. 2017, 2018), as indicated by significant declines in calcification (Schmidt et al. 2011, 2014; Doo et al. 2012; Pinko et al. 2020) and reproduction (Hohenegger et al. 2019; Hallock and Reymond 2022; Reymond et al. 2022). The response to thermal stress can vary among LBF species, likely due to differences in local physicochemical conditions and the ability to acclimatise in different cross-shelf habitats (Doo et al. 2014; Fujita et al. 2014; Prazeres et al. 2016; Kenigsberg et al. 2022) or through evolutionary adaptation (Schmidt et al. 2016; Stuhr et al. 2021).

With marine heatwaves becoming more frequent and prolonged (Hobday et al. 2018), there is an increasing need to understand the impact of thermal stress on LBF populations. We documented the thermal stress response of two prominent LBF species, *Marginopora vertebralis* (a dinoflagellate endosymbiont-bearing porcelaneous foraminifera)

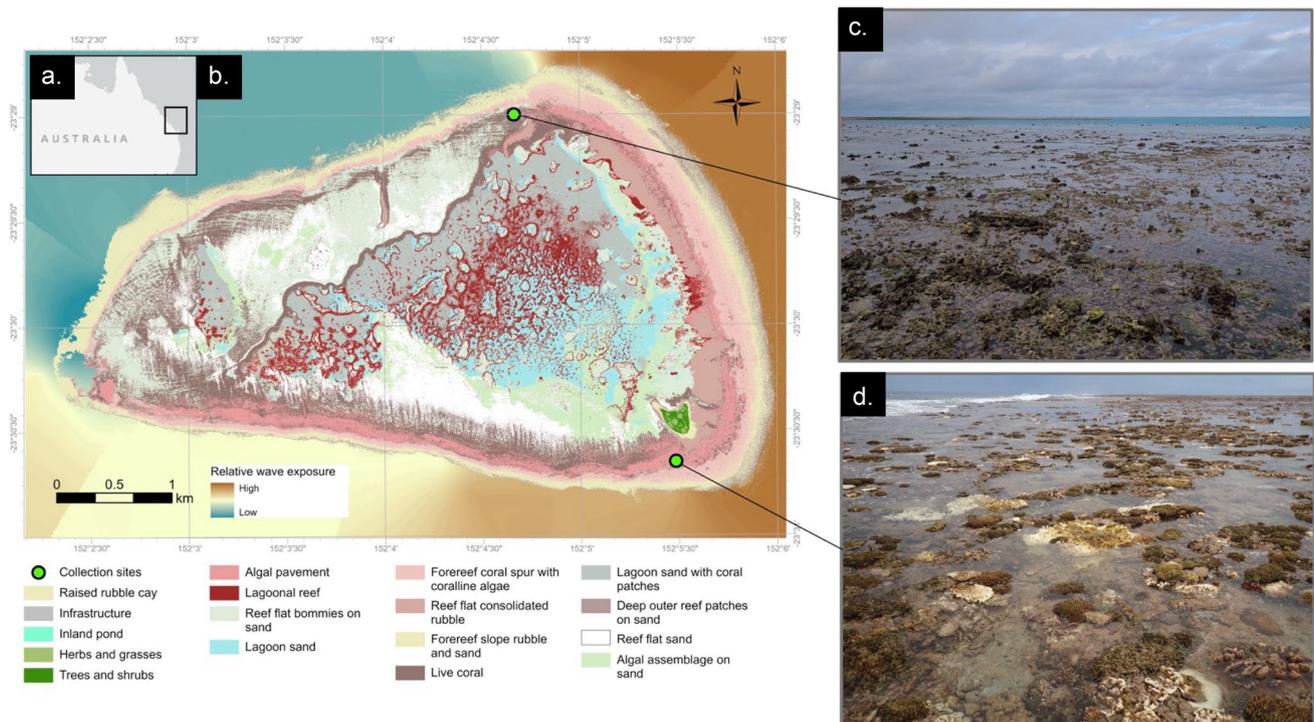


Fig. 1 Map of **a** Capricorn Bunker Group 100 km off the Australian mainland in the Great Barrier Reef. **b** Benthic-type map of One Tree Reef (OTR), adapted from Hamylton et al. (2013), and relative wave exposure (GREMO) (Pepper and Puotinen 2009). Collection sites are

indicated with in situ images of relative microhabitat. Samples were collected from two coral–algal fore reef locations: **c** the northern leeward zone and **d** the southern windward zone

and *Baculogypsina sphaerulata* (a diatom endosymbiont-bearing hyaline foraminifer), across One Tree Reef's leeward and windward coral–algal reef flats during the 2024 marine heatwave that caused mass bleaching in corals (Cantin et al. 2024).

Methods and materials

One Tree Reef (OTR) is located 100 km off the Australian mainland in the Capricorn Bunker Group. For 40 years, it has had the highest level of protection (Scientific Zone) in the GBR World Heritage Area. OTR is a tide-dependent system, in which the leeward (northern) and windward (southern) sides of the reef experience different hydrodynamic

exposures (Pepper and Puotinen 2009; Duce et al. 2020). Samples of LBF were collected from two coral reef sand aprons (Fig. 1): the northern leeward zone (23°29'00" S, 152°04'40" E) and the southern windward zone (23°30'39" S, 152°05'29" E). These sites, accessible at low tide, serve as the primary sources of carbonate sediment for the lagoon (Doo et al. 2012; Vila-Concejo et al. 2022; Bauder et al. 2023) and have up to one-third foraminifera content (Fellows et al. 2017).

One Tree Lagoon reached Alert Level 1, equivalent to 4-DHW on January 24 and Alert Level 2 on February 8 (8-DHW), with local on-reef temperature sensors recording a peak of 30.55 °C (Byrne et al. 2025). Edging closer to extreme heatwaves earlier than anticipated, sea surface temperature (SST) in the upper range of 30 °C and 32 °C

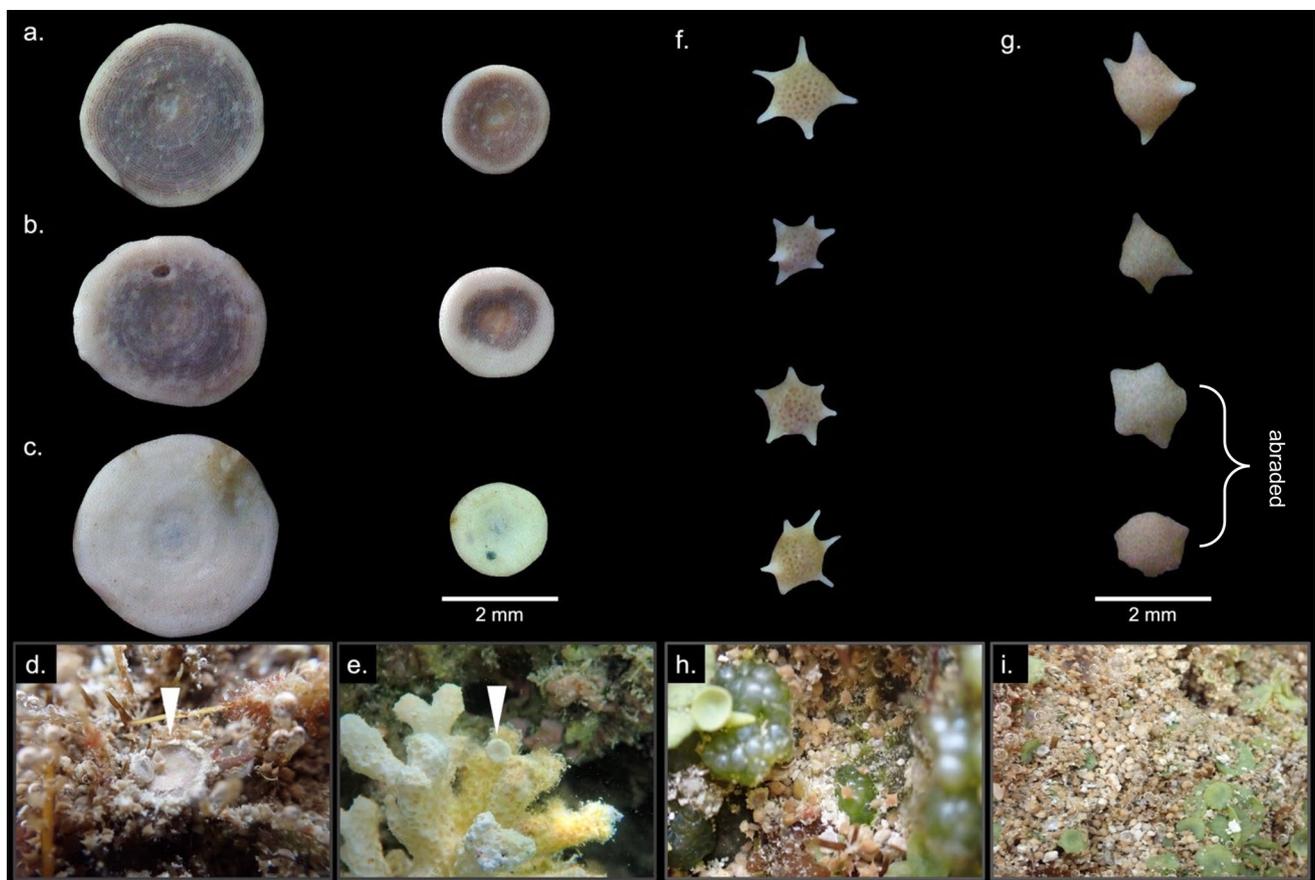


Fig. 2 Thermal stress chart of *M. vertebralis* and *B. sphaerulata*, the 2 mm scale bar is consistent for all individual images of the foraminifera. **a–e** *M. vertebralis* sorted into size classes left, > 2 mm; right, ≤ 2 mm. **a** Healthy: active pseudopodia with retention of endosymbionts noted by the robust pigmentation of brown/purple hues and a thin white ring around the aperture of the test. **b** Mottled: weakly retained endosymbiont, possible abrasion of test and patchy pigmentation. **c** Bleached: complete loss of endosymbiont, no visible pigmentation little to no abrasion. **d** Healthy *M. vertebralis* attached to turfing algae. **e** In situ bleached *M. vertebralis* attached

to bleached coral. Thermal stress chart of *B. sphaerulata* (**f–i**) sorted by **f** Healthy: active pseudopodia, intact radial spines, robust brown/orange pigmentation, and visibly punctuated chambers. **g** Bleached; non-active pseudopodia radial spines in tack, loss of endosymbiont indicated by reduced pigmentation, including examples of abraded *B. sphaerulata*, which were excluded from the analysis. **h** Healthy individuals located within the turf algal pavement compared to **i** bleached and abraded tests. Individuals from both species with extensive abrasion and/or covered in epiphytes were excluded from the analysis because to timing of the death likely occurred prior to the heatwave

is in line with the IPCC upper projections (A1F) for 2100 (Meehl et al. 2007).

Samples of *Marginopora vertebralis* and *Baculogypsina sphaerulata* were systematically collected from forty 50 cm² quadrats (twenty samples of each species along each reef site). Collection sites were located on the crustose coral-line (*Porolithion onkodes*) dominated coral–algal reef pavements, which were also covered with various calcifying, filamentous, and turfing algae, e.g. *Halimeda* spp., *Laurencia* spp., *Caulerpa racemosa* (c.f. Borowitzka and Larkum 1986). In each quadrat, the first 100 individuals of each species encountered were collected. However, some quadrats on the leeward reef flat contained fewer than 100 *M. vertebralis* individuals; in these cases, the maximum number available were collected, in total 3,430 individuals were assessed. Samples were collected and rapidly assessed at low tide over four consecutive days, ensuring consistent tidal conditions and no unusual wave action. Grids were carefully selected in areas with stable water coverage. To minimise disturbance, foraminifera were collected in situ using a fine brush to gently remove them from turf algae without damaging the substrate, ensuring the unbiased collection of both recently dead and living samples. Specimens were carefully sorted, with movement of living individuals observed, and photographed

in both the field and post-sorting under low light conditions. Samples were collected into sealed containers with natural seawater and transported back to the water laboratory at the research station in an insulated bag to reduce light exposure and maintain stable ambient temperatures. Samples were stored in separate vials to examine pseudopodia activity, to distinguish between living and dead individuals, and scored based on the extent of visible pigment loss. Precautions were also taken to minimise handling and light exposure during pigment checks to reduce potential bleaching effects. Taxonomic identifications were conducted following Hayward et al. (2025) and Uthicke and Nobes (2008) using a digital Zeiss STEMI 305 microscope with an attached camera.

A health chart for *M. vertebralis* and *B. sphaerulata* was developed and used to assign individuals into specific categories (Fig. 2). Healthy specimens of *M. vertebralis* exhibited an even brown/purple hue with a thin white ring along the aperture, while healthy *B. sphaerulata* displayed an orange/brown hue with intact radial spines. Individuals of *M. vertebralis* with slight purple or brown hues and thick white bands or patches were categorised as mottled. Mobility was also observed among individuals that were healthy or mottled. Pale individuals with pigment loss (photosymbiont damage) and no abraded individuals

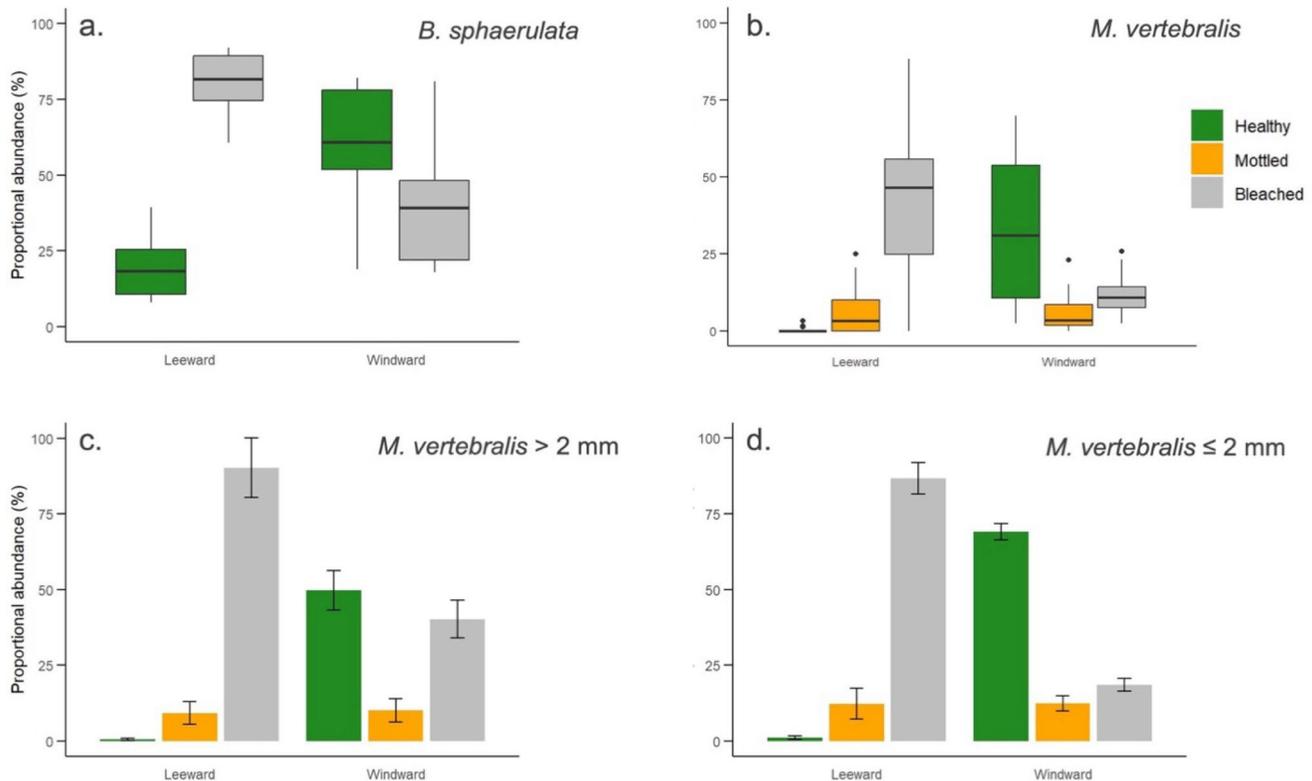


Fig. 3 Response of *M. vertebralis* and *B. sphaerulata* from One Tree Reef, indicating there is a greater thermal stress in the northern leeward coral–algal reef flat. Overall proportional abundance of bleach-

ing in *M. vertebralis* and *B. sphaerulata* (a, b). Bleach levels of c large (> 2 mm) and d small (≤ 2 mm) *M. vertebralis* are categorised by size to reflect different life cycle stages

Table 1 Chi-squared tests contingency table assessing the effects of location on the health status of *M. vertebralis* and *B. sphaerulata* populations. The health status distribution significantly differs (p -value < 0.001) between the two sites, suggesting that location is an important factor influencing the thermal stress response of *M. vertebralis* and *B. sphaerulata*

		Marginopora vertebralis															
		<i>Baculogypsina sphaerulata</i>				Percentage (ALL)				Percentage (Small)				Percentage (Large)			
		Healthy	Bleached	Mottled	Count	Healthy	Bleached	Mottled	Count	Healthy	Bleached	Mottled	Count	Healthy	Bleached	Mottled	Count
Leeward	Windward	19.75	80.25	11.28	0.94	87.78	12.29	86.59	1.12	12.45	18.53	0.57	90.23	9.20	10.05	40.21	
		59.17	40.83	11.94	64.96	23.10	69.02	18.53	69.02	12.45	18.53	49.74	40.21	10.05	10.05	40.21	
		Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	Count	
Leeward	Windward	182	769	60	5	467	4	310	4	44	310	1	157	16	19	76	
		592	459	107	582	207	488	131	488	88	131	94	76	19	19	76	
Chi-squared test of independence																	
		30.90			99.98		106.79		106.79				17.37				
		1			2		2		2				2				
		<0.001			<0.001		<0.001		<0.001				<0.001				

from both species were considered recently dead and classified as bleached. Encrusted and highly abraded individuals were excluded from the analysis as the time of death likely occurred prior to the heatwave. Test abrasion was determined by radial spine or prolocular alteration. The size class distribution for *M. vertebralis* was divided into two categories based on test diameter: those with a diameter of ≤ 2 mm and those with a diameter of > 2 mm. This classification was used to capture two distinct cohorts and ensure that age was considered. Conversely, the test size for *B. sphaerulata* was uniform, so no size-based classification was necessary.

The benthic cover types were mapped in ArcGIS Pro 3.3 by adapting the reef zonation classifications from Hamylton et al. (2013). In combination, GREMO (Generic Model for Estimating Relative Wave Exposure) was overlaid as a visual representation of wave exposure, highlighting the differences between the leeward and windward reef zones at OTR (Pepper and Puotinen 2009). While we recognise that these data may not fully capture the nuances of intertidal hydrodynamics, they represent the most accurate and comprehensive datasets currently available. Since the dataset is categorical in nature, the chi-squared test of independence, analysed in R Studio, was used to test the health status of each species from two wave exposure locations. Moreover, the residuals, diagnostic, and plots were made using the statistical package (lme4) as a GLMM general linear model were examined to ensure model fit and normality (Fig. S1).

Results and discussion

After experiencing an 8-DWH with local on-reef temperatures exceeding 30 °C (Byrne et al. 2025), it is evident both the northern and southern coral–algal reef zones on OTR were seriously impacted by the prolonged marine heatwave on OTR in 2024. Even with their morphological and physiological differences (Fig. 2), *M. vertebralis* and *B. sphaerulata* both exhibited similar thermal stress patterns in the shallow coral–algal reef flat of OTR. Nonetheless, populations of *M. vertebralis* and *B. sphaerulata* from the northern leeward coral–algal reef flat on OTR showed greater thermal stress (Fig. 3; Table 1). Approximately 80% of all *B. sphaerulata* was bleached in the leeward site compared to ~40% of individuals in the windward site. Even though high intensity bleaching was observed at each site, the chi-squared of independence suggests a higher bleaching trend associated with lower wave exposure ($p < 0.01$) for both *M. vertebralis* and *B. sphaerulata* (Table 1). The removal and transport of dead epiphytic foraminiferal shells by wave energy is a plausible mechanism influencing observed bleaching patterns, particularly over longer timescales (years to decades), as evidenced by test abrasion

in high hydrodynamic environments (Fellowes et al. 2017). These findings suggest the windward (southern) reef flat, which experiences greater hydrodynamic exposure (Pepper and Puotinen 2009; Duce et al. 2020), has the potential to create and support thermal refuges as observed with internal waves at depths in coral reefs (Wyatt et al. 2020).

While this is a rapid assessment of the response of LBF to heatwave conditions on one reef, the implications for carbonate sediment production and biodiversity on a larger regional or global scale are serious. Previous studies have shown the impacts of temperature on *M. vertebralis* growth from OTR were evident at 2 °C above ambient conditions (28 °C) with a 130% decline in CaCO₃ production at 32 °C (Doo et al. 2012). Similarly, studies along a natural temperature and nutrient gradient in the GBR indicated the average summer temperature of 28 °C is near the upper limit before the onset of thermal stress for *M. vertebralis* (Reymond et al. 2011).

Shifting habitats and range potentials of LBF due to climate change will significantly impact biodiversity and carbonate budgets. Future studies will benefit from additional monitoring of the transport and fate of detached shells in different hydrodynamic conditions to further refine our understanding of the influence of hydrodynamics on the bleaching intensity of LBF. Even though the unique conditions at OTR influenced our decision to use a north–south sampling design, an east–west sampling approach could also capture differences between coasts with varying hydrodynamic regimes. This rapid in-field bleaching guide (Fig. 2) can be applied globally to understand reef population dynamics and sediment budgets of foraminifera. This guide can be applied in other monitoring sites to gauge the effect of marine heat waves along with additional parameters such as PAM fluorometry, chlorophyll-a concentration, and proteomic analyses (e.g. Talge and Hallock 2003; Schmidt et al. 2011; Stuhr et al. 2021) to understand the overall health and potential climate adaptation of LBF.

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Author contributions C.E.R. and C.R. conceived the study and designed the methodology. C.R. and G.P. conducted fieldwork and data analysis. C.E.R. contributed to statistical validation and visualisation. All authors contributed to writing and editing the manuscript. Supervision and funding acquisition were led by C.E.R., J.M.W., and M.B.

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Data availability Data is provided within the manuscript or supplementary information files.

Declarations

Conflict of interest The authors declare no competing interests.

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