




No effects of abiotic and anthropogenic factors on reef-associated neonate shark abundance within a shark nursery-area system

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ABSTRACT

Context. Coastal habitats function as shark nursery areas; however, coastal habitats can experience extreme variation in abiotic conditions and are susceptible to human disturbances. **Aims.** Drivers of abundance were tested within a shark nursery-area system in two populations of reef-associated neonate sharks, namely, blacktip reef sharks (*Carcharhinus melanopterus*) and sicklefin lemon sharks (*Negaprion acutidens*). **Methods.** Catch data from a fisheries-independent gill-net survey ($n = 90$ sets from October 2018 to March 2019) at 10 sites around Moorea, French Polynesia, were used to test for associations between shark abundance and abiotic conditions (temperature, oxygen, pH, salinity, lunar phase and depth). Historical levels of fin-fish fishing effort, trampling (i.e. human movement through habitat), and coastal artificialisation (i.e. walls and embankments) estimated for each site were used to test for anthropogenic effects on shark abundance. **Key results.** There were no effects of any abiotic or anthropogenic factor on abundance of either species. **Conclusions.** Previous work corroborates our findings by demonstrating neonate sharks' physiological tolerance to extreme abiotic conditions and high survival in response to anthropogenic stressors. Alternatively, populations are already degraded from decades of coastal development. **Implications.** These data can aid in predicting the use of coastal habitats as shark nursery areas.

Keywords: blacktip reef shark, catch-per-unit-effort, coastal development, fishing pressure, human disturbance, sicklefin lemon shark, temperature, young-of-the-year.

Introduction

Coastal habitats can play important roles for sharks. Some species spend their entire life among coastal habitats, whereas others use them repeatedly over varying temporal scales (e.g. for foraging, reproduction, or refuge from seasonally extreme abiotic conditions), and some just rely on them during early ontogeny (Knip *et al.* 2010). For example, neonates and juveniles may use coastal habitats as nursery areas (Heupel *et al.* 2019), which are thought to improve the recruitment of reproducing adults (Beck *et al.* 2001; Fodrie *et al.* 2009). Because of the potential significance of shark nursery areas for fitness in shark populations (Heupel *et al.* 2007; Kinney and Simpfendorfer 2009), there is general interest in identifying such habitats (e.g. Simpfendorfer and Milward 1993; DeAngelis *et al.* 2008; Froeschke *et al.* 2010a) and defining drivers of habitat use within those potential nursery areas as population management or conservation tools (e.g. Carrier and Pratt 1998; Heupel and Simpfendorfer 2005; Oh *et al.* 2017a). Known drivers of shark abundance include local abiotic conditions and anthropogenic influences (Schlaff *et al.* 2014; MacNeil *et al.* 2020), and these may also explain why sharks use, or do not use, habitats as nursery areas.

Within coastal areas, changes in abiotic conditions affect the distribution, abundance and patterns of habitat use in young sharks. Associations among various abiotic conditions (e.g. temperature, salinity, turbidity, oxygen, pH, tide, photoperiod) have been observed in sharks across life-history stages and have been discussed in a comprehensive review by Schlaff *et al.* (2014). Temperature can affect the abundance of neonate and juvenile reef sharks (Yates *et al.* 2015; Oh *et al.* 2017a; Vierus *et al.* 2018). For example, juvenile bull

sharks (*Carcharhinus leucas*) were documented to shift their nursery area poleward in response to climate-driven ocean warming (Bangley *et al.* 2018a). Oxygen also has variable effects on young shark abundance (Froeschke *et al.* 2010b; Oh *et al.* 2017a; Bangley *et al.* 2018b), which is thought to reflect the inter-specific differences in oxygen-supply capacity (Dubuc *et al.* 2021). Salinity is another important driver of the abundance of young sharks (Froeschke *et al.* 2010b; Yates *et al.* 2015), including euryhaline species (Dwyer *et al.* 2020). There is interest in identifying abiotic drivers of abundance in young sharks because of the importance of predicting important habitats for shark management and conservation (Oh *et al.* 2017a), and for adaptive management of shark populations in the face of climate change (Bouyoucos and Rummer 2021).

Globally, reef shark populations are experiencing declines owing to environmental change and anthropogenic pressures (Sherman *et al.* 2023). Anthropogenic influences also demonstrate potent effects on the distribution, abundance and habitat-use patterns in sharks. Fishing pressure (i.e. general fishing activity that may interact with predators via capture as bycatch, loss of prey species, etc.) and the accessibility of marine habitats by humans has been shown to negatively affect the abundance of predators, including sharks (Nadon *et al.* 2012; Cinner *et al.* 2018; Frisch and Rizzari 2019; Clementi *et al.* 2021; Lester *et al.* 2022). However, this can be reversed by implementing science-based fisheries management (e.g. reducing catch limits, fishing bans, or designating marine protected areas, MPAs); indeed, these strategies have been found to support high shark abundance (Ward-Paige and Worm 2017; MacNeil *et al.* 2020). Anthropogenic-driven habitat loss or degradation can also affect the survival and abundance of sharks. For example, the abundance of juvenile lemon sharks (*Negaprion brevirostris*) within a Bahamian nursery area decreased in response to coastal development, likely because of the associated habitat loss (Jennings *et al.* 2008). Similarly, studies using environmental DNA also suggested that shark species richness declines with increasing levels of anthropogenic disturbances (Bakker *et al.* 2017). As such, sources of anthropogenic disturbance related to fishing pressure, direct human impacts on shark nursery areas, and coastal development should be considered when assessing drivers of shark abundance.

The purpose of this study was to test the effects of abiotic conditions and the level of anthropogenic impact on the abundance of two populations of reef-associated neonate sharks within a shark nursery-area system. Neonate blacktip reef (*Carcharhinus melanopterus*) and sicklefin lemon (*Negaprion acutidens*) sharks were studied at 10 sites around the island of Moorea, French Polynesia. This is a system that is known to contain shark nursery areas for both species (Bouyoucos *et al.* 2022), where shark fishing has been prohibited since 2012. The first study objective was to test for differences in neonate *C. melanopterus* and *N. acutidens* abundance and abiotic conditions (i.e. temperature, oxygen,

pH, salinity and depth) among sites around Moorea. Then, we tested for associations between neonate abundance and abiotic conditions (i.e. temperature, oxygen, pH, salinity, tide, depth and lunar phase) across all sites. The second study objective was to determine the historical levels of anthropogenic impact (i.e. non-targeted fishing effort, direct human impact and coastal artificialisation by construction of walls and embankments) for each site around the island and then test for associations between neonate abundance and the level of anthropogenic impact across all sites. Together, these data are important for understanding the effects of environmental change and anthropogenic effects on the abundance of reef-associated neonate sharks. More broadly, our results can aid in predicting the use of habitats as shark nursery areas, information that is critical for spatiotemporal planning and management of important shark habitats.

Materials and methods

Study site and species

This study was conducted on the island of Moorea in French Polynesia (17°30'S, 149°50'W). French Polynesia was designated a 'shark sanctuary' when the commercial harvest of all shark species was banned in 2012 (Ward-Paige 2017). The lagoon surrounding the island of Moorea has resident populations of neonatal *C. melanopterus* and *N. acutidens* that have been studied extensively by the Centre de Recherches Insulaires et Observatoire de l'Environnement (CRIOBE) in collaboration with the Physioshark Project on Moorea since 2007 (Mourier *et al.* 2013a; Debaere *et al.* 2023). Ten sites (i.e. Apaura, Haapiti, Maharepa, Paorea, Papetoai, Pihaena, Tiki, Vaiane, Vaiare and Valorie) that encompass the island's ~60-km coastline are surveyed annually during the wet season (October–March) with gill-nets to sample shark neonates (Fig. 1). The areas of these sites that are being surveyed are very shallow (i.e. <2 m deep), relatively small (i.e. <1 km²), and are delineated by deep channels that occur naturally or have been dredged. Channels are typically within 50 m of shore and are thought to limit the distribution of shark neonates (Bouyoucos *et al.* 2020a). Apaura and Vaiane are nursery areas for *N. acutidens* (Bouyoucos *et al.* 2022), meaning that these sites exhibit high shark abundance relative to other sites and persistent use of these habitats by individuals over time (Heupel *et al.* 2007). Abundance of neonatal *C. melanopterus* appears to fluctuate between Maharepa and Paorea on an annual basis, suggesting that these sites may serve as nursery areas over longer time scales (Bouyoucos *et al.* 2022). Yet, there are some sites around Moorea where both neonates of both species can co-occur (Matich *et al.* 2017).

The abundance of neonates of both species around Moorea increases annually around October. The parturition season for *C. melanopterus* is thought to occur annually from September through February, with the most births occurring in November

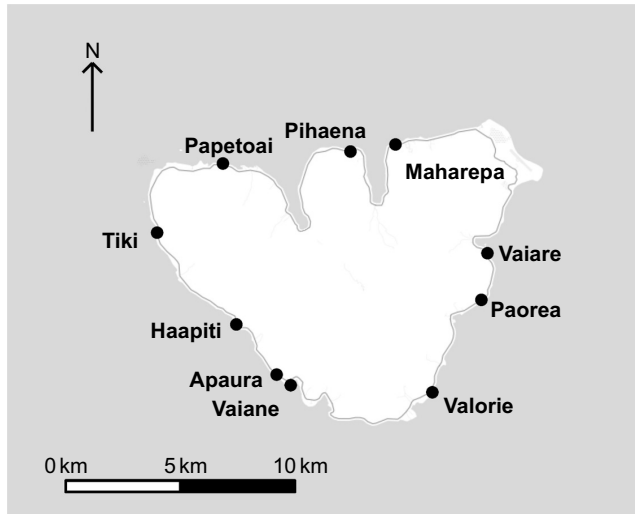


Fig. 1. Locations of 10 reef-associated neonate shark sampling sites all around the island of Moorea, French Polynesia (17°30'S, 149°50'W).

(Debaere *et al.* 2023), whereas the parturition season for *N. acutidens* occurs biennially from July through November (Porcher 2005; Mourier *et al.* 2013b). Then, the abundance of neonates decreases around January, as parturition ends and neonates emigrate from the sampling sites or experience mortality (Mourier and Planes 2013; Mourier *et al.* 2013b). During the parturition season, monthly temperatures at the above-mentioned sites range between 27 and 31°C (Eustache *et al.* 2024), with daily temperatures ranging from 26 to 36°C and most sites experiencing oxygen supersaturation (Bouyoucos *et al.* 2021). Neonates of both species are reported to exhibit site fidelity around Moorea, as evidenced by recaptures of individuals at a single site and not multiple sites during the months following parturition (Bouyoucos *et al.* 2022); indeed, this reflects what has been documented for these species elsewhere (Lea *et al.* 2016; Oh *et al.* 2017b).

Gill-net survey

Sampling occurred between 1 October 2018 and 12 February 2019. One site was fished per night, and all 10 sites were fished at least once per fortnight. Sampling effort was standardised to control for possible confounding effects of time of day by fishing for a 3-h period, usually from 17:00 to 20:00 hours. Neonatal *C. melanopterus* and *N. acutidens* were captured using monofilament gill-net (50 m long by 1.5 m tall with 5-cm mesh; 10-cm stretch) set perpendicularly from shore. In some sites (e.g. Apaura, Maharepa, Pihaena, Valorie), the gill-net reached from shore all the way to the nearest deep channel delineating the site. Sharks were removed from gill-nets within minutes of capture to minimise the possibility of capture-induced mortality (Dapp *et al.* 2017; Bouyoucos *et al.* 2018). Once removed, sharks were photographed and tagged for identification with passive

integrated transponder (PIT) tags, measured (fork and total length, mm; weight, kg; and girth, mm), sexed and then released. The duration of each gill-net set and the number of *C. melanopterus* and *N. acutidens* individuals were recorded. Set number was coded (i.e. 1, 2... *n*) for inclusion in models to test for temporal effects. Catch-per-unit effort (CPUE, sharks h⁻¹) was calculated for each species for each set as the total number of sharks caught per hour of fishing.

Measuring abiotic conditions

To test whether shark abundance was correlated with abiotic conditions, environmental conditions were measured and recorded at the beginning of each gill-net set. Single temperature (°C), dissolved oxygen (mg L⁻¹), pH (National Bureau of Standards scale) and salinity readings were taken at the mid-point of the net at the surface of the water. Temperature and oxygen were measured with a YSI Pro 20 (YSI Inc., Yellow Springs, OH, USA), pH was measured with a Seven2Go Pro (Mettler-Toledo GmbH, Greifensee, Switzerland), and salinity was measured with a hand-held refractometer. Depth (cm) was measured at the mid-point of the net. Tidal phase (i.e. ebbing, flooding) and lunar phase (percentage of full moon) were recorded from local online tables.

Defining anthropogenic impact

The following three types of anthropogenic impacts were investigated: fishing effort, direct human impact (i.e. trampling) and coastal artificialisation. Fishing effort was obtained from survey and indirect mapping from Thiault *et al.* (2017), by combining fishing capacity on the basis of the density of households and their dependency on fishing for livelihood, with fishing suitability, based on fishers' preference for fishing grounds (Thiault *et al.* 2017). These data, therefore, provide an indication of relative fin-fish fishing effort for a given area. Indeed, fishing effort does not reflect targeted shark fishing, which was banned in French Polynesia (c. 2012); rather, fishing effort describes the potential for shark bycatch, alongside catch of target, non-shark species that may be predators or prey of shark neonates. Indeed, fin-fish fishing, predominantly by hook-and-line or spear, is not banned at any of the study sites; however, several of the sites (e.g. Paorea, Pihaena) occur on the border of MPAs (Thiault *et al.* 2019). Direct human impact, which represents the volume of human movement through an area, was assessed for areas <1.8 m deep by applying a kernel-density function with a 200-m radius around all households (weighted by the number of people per household) and high-density tourism sites (weighted by daily visitation rates), as described in Loiseau *et al.* (2021). These data were collected in 2012 (household size) and 2017 (tourism visitation) and were originally presented in Loiseau *et al.* (2021). Finally, coastal artificialisation was modelled on the basis of a high-resolution shoreline categorisation undertaken around the entire island in

2018, by using previously published data (Madi Moussa *et al.* 2019). A coastal artificialisation index was obtained by extrapolating the ‘walls and embankments’ category onto the lagoon by using a 100-m radius (Loiseau *et al.* 2021). Each metric was mapped throughout Moorea’s lagoon at a 5- × 5-m resolution, and values were extracted at each of the 10 study sites by taking the average value within a 100-m radius of the centre of the net.

Statistical analysis

Catch-per-unit effort was compared among sites by using Kruskal–Wallis rank sum tests, by using the *R* package *Stats* (ver. 4.4.0, see <https://stat.ethz.ch/R-manual/R-devel/library/stats/html/00Index.html>; R Foundation for Statistical Computing, Vienna, Austria, see <https://www.r-project.org/>). *Post hoc* multiple comparisons were examined using the *R* package *pgirmess* (ver. 2.0.3, P. Giraudoux, see <https://CRAN.R-project.org/package=pgirmess>). Similarly, differences in temperature, oxygen, pH, salinity and depth among sites were tested by using Kruskal–Wallis rank sum tests.

Correlation and collinearity between continuous explanatory variables (i.e. temperature, oxygen, pH, salinity, depth, log ratio transformed lunar phase and chronological set number) was first assessed with Pearson’s correlation coefficients (*r*) and variance inflation factors (VIF) respectively. Pearson’s *r* and VIF were calculated using the *R* packages *Stats* and *car* packages (ver. 3.1-2, see <https://CRAN.R-project.org/package=car>; Fox and Weisberg 2019). Criteria for correlation and collinearity were $r \geq 0.80$ and $VIF \geq 3$ respectively (Zuur *et al.* 2009). Parameters that did not meet these criteria were excluded from the final models.

Effects of abiotic conditions on shark abundance were assessed using generalised linear mixed models (GLMM). Shark abundance was assumed to follow a negative binomial distribution (Yates *et al.* 2015), and visual inspection of histograms of *C. melanopterus* and *N. acutidens* abundance suggested that shark abundance data were zero-inflated for both species. To account for variation in sampling effort among sets, the logarithm of sampling effort in minutes was included as an offset variable (Yates *et al.* 2015; Vierus *et al.* 2018). Site was included as a random effect to account for repeated sampling within sites and to investigate trends across *C. melanopterus* and *N. acutidens* neonate subpopulations around Moorea. Within models for each species, sites were excluded if no sharks were caught during the entire study. Zero-inflated negative binomial GLMMs were implemented for each species by using the *R* package *glmmTMB* (ver. 1.1.9, see <https://cran.r-project.org/package=glmmTMB>; Brooks *et al.* 2017). Model validation included visual inspection of scatterplots of model residuals against fitted values to assess homogeneity, and model residuals against explanatory variables to assess independence (Zuur *et al.* 2009). Statistical significance was determined using 95% confidence intervals (CIs) of fixed-effect term estimates

generated from 1000 posterior distributions (Hasler *et al.* 2016).

Because of differences in the temporal scale of abiotic and anthropogenic variables, these were modelled separately. Effects of anthropogenic impacts on CPUE were tested using linear models. Associations were investigated within each species between the natural logarithm of anthropogenic impact score (fishing effort, trampling and artificialisation) and mean CPUE at each site. Owing to differences in variation in CPUE among sites, CPUE was weighted by its standard error for each site. Model validation was conducted as described previously. Analyses were conducted using the *R* package *Stats*.

Ethical approval

All methods described herein were approved by the James Cook University Animal Ethics Committee (A2394). Permission to conduct research on sharks was granted by the French Polynesian Ministère de la Promotion des Langues, de la Culture, de la Communication, et de l’Environnement (Arrêté Number 5129).

Results

Site-specific differences

Over the course of the 2018–2019 parturition season, 90 gill-net sets resulted in the capture of 110 *C. melanopterus* and 116 *N. acutidens* individuals (Table 1). Accounting for recaptures, 94 *C. melanopterus* individuals and 80 *N. acutidens* individuals were caught in total. Shark abundance, measured as catch-per-unit effort (CPUE; sharks h^{-1}), differed among sites for *C. melanopterus* (Kruskal–Wallis rank sum test, $\chi^2 = 42.65$, d.f. = 9, $P < 0.001$; Fig. 2) and *N. acutidens* ($\chi^2 = 37.74$, d.f. = 6, $P < 0.001$; Fig. 2). At Haapiti, CPUE of

Table 1. Sampling effort of reef-associated neonate sharks around Moorea, French Polynesia.

Site	Gill-net sets (n)	Sampling effort (h)	<i>Carcharhinus melanopterus</i> (n)	<i>Negaprion acutidens</i> (n)
Apaura	8	21.0	2	13
Haapiti	12	25.7	33	0
Maharepa	10	25.6	3	2
Paorea	9	22.1	13	0
Papetoai	10	25.7	24	1
Pihaena	9	19.25	1	9
Tiki	9	21.9	4	24
Vaiane	7	15.8	4	62
Vaiare	9	23.4	12	0
Valorie	7	16.3	14	5

C. melanopterus was higher than that at Apaura, Maharepa, Pihaena, Tiki, and Vaiane, whereas CPUE of *N. acutidens* at Vaiane and Tiki was higher than that at Maharepa, Papetoai and Valorie. All *C. melanopterus* and most *N. acutidens* individuals were neonates. At least four *N. acutidens* individuals were 1- or 2-year-old juveniles; three of these sharks were recaptured at Vaiane and Apaura from the previous survey year (i.e. 2017–2018), and one individual that was captured at Valorie was 1100 mm in total length. *C. melanopterus* was present at all sites, whereas *N. acutidens* was absent from Haapiti, Paorea and Vaiare.

There were significant differences in dissolved oxygen (Kruskal–Wallis rank sum test, $\chi^2 = 41.72$, d.f. = 9, $P < 0.001$), pH ($\chi^2 = 48.19$, d.f. = 9, $P < 0.001$) and depth among sites ($\chi^2 = 40.76$, d.f. = 9, $P < 0.001$). Temperature and salinity did not differ significantly among sites (Fig. 3). No variable, including depth, differed between tide states (i.e. ebbing or flooding; data not shown).

Abiotic conditions

Oxygen was found to be correlated with pH ($r = 0.801$) and exhibited collinearity ($VIF \geq 3$) with temperature, salinity, depth, lunar phase. pH exhibited collinearity with set number and lunar phase. Therefore, statistically informed selection criteria (i.e. Pearson's r and VIF) justified the *a priori*

decision to exclude oxygen and pH as explanatory variables of shark abundance in further models. All other explanatory variables (i.e. temperature, salinity, depth, lunar phase and set number) were not correlated ($r = -0.433$ – 0.441), did not exhibit collinearity ($VIF < 2$), and were, therefore, included in final models. Although abiotic conditions did differ among sites (Fig. 3), there were no significant effects of any variable (i.e. temperature, salinity, tide, depth, lunar phase and set number) on the abundance of *C. melanopterus* or *N. acutidens* (Table 2).

Anthropogenic impact

Anthropogenic-impact scores varied considerably among sites, as demonstrated by coefficients of variation for fishing effort (32%), artificialisation index (82%) and direct human impact (58%). However, linear regressions did not show statistically significant relationships between CPUE and any anthropogenic impact score (Fig. 4, Table 3).

Discussion

The purpose of this study was to identify potential abiotic and anthropogenic drivers of abundance in neonatal reef sharks. Data collected at dusk during the 2018–2019 parturition season around the island of Moorea, French Polynesia,

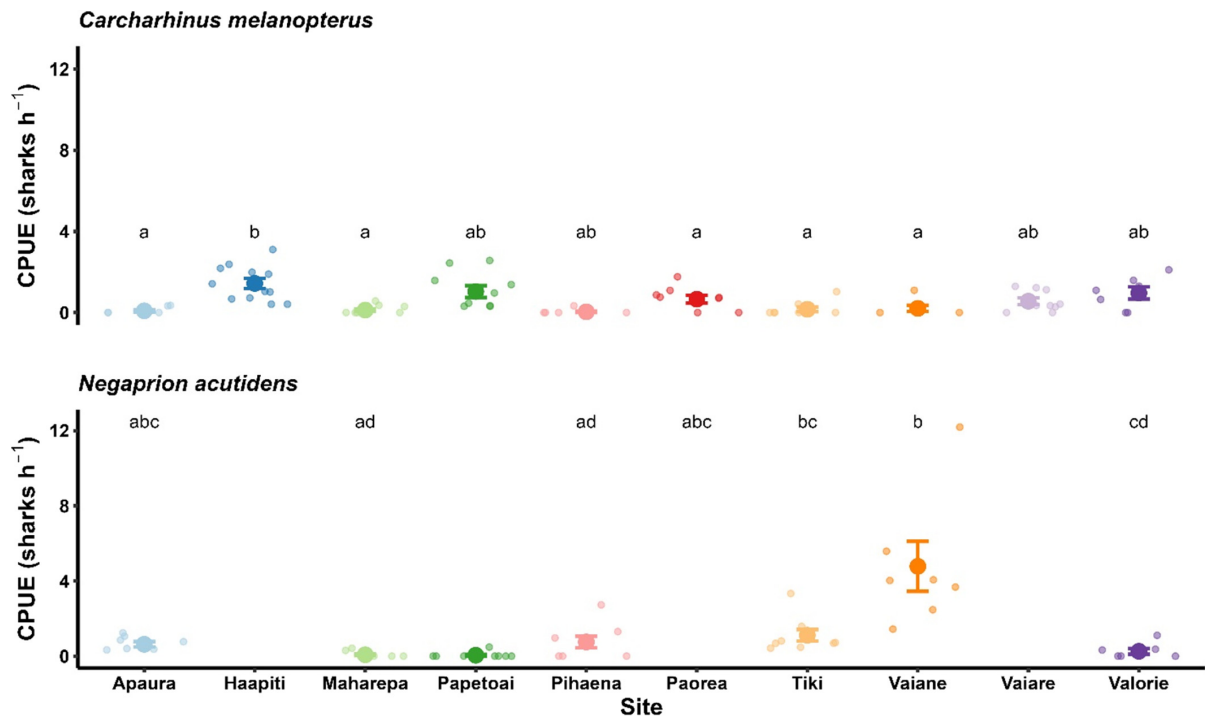


Fig. 2. Catch-per-unit effort (CPUE) of reef-associated neonate sharks from 10 sites around Moorea, French Polynesia. Data are presented as means with standard errors of the mean. Individual observations ($n = 7$ – 12) represent CPUE for *C. melanopterus* and *N. acutidens* calculated for individual sampling events at each site. Different letters represent significant differences in CPUE between sites within each species (Kruskal–Wallis test, $P < 0.05$). Note that *N. acutidens* was not encountered at the Haapiti, Paorea and Vaiare sites.

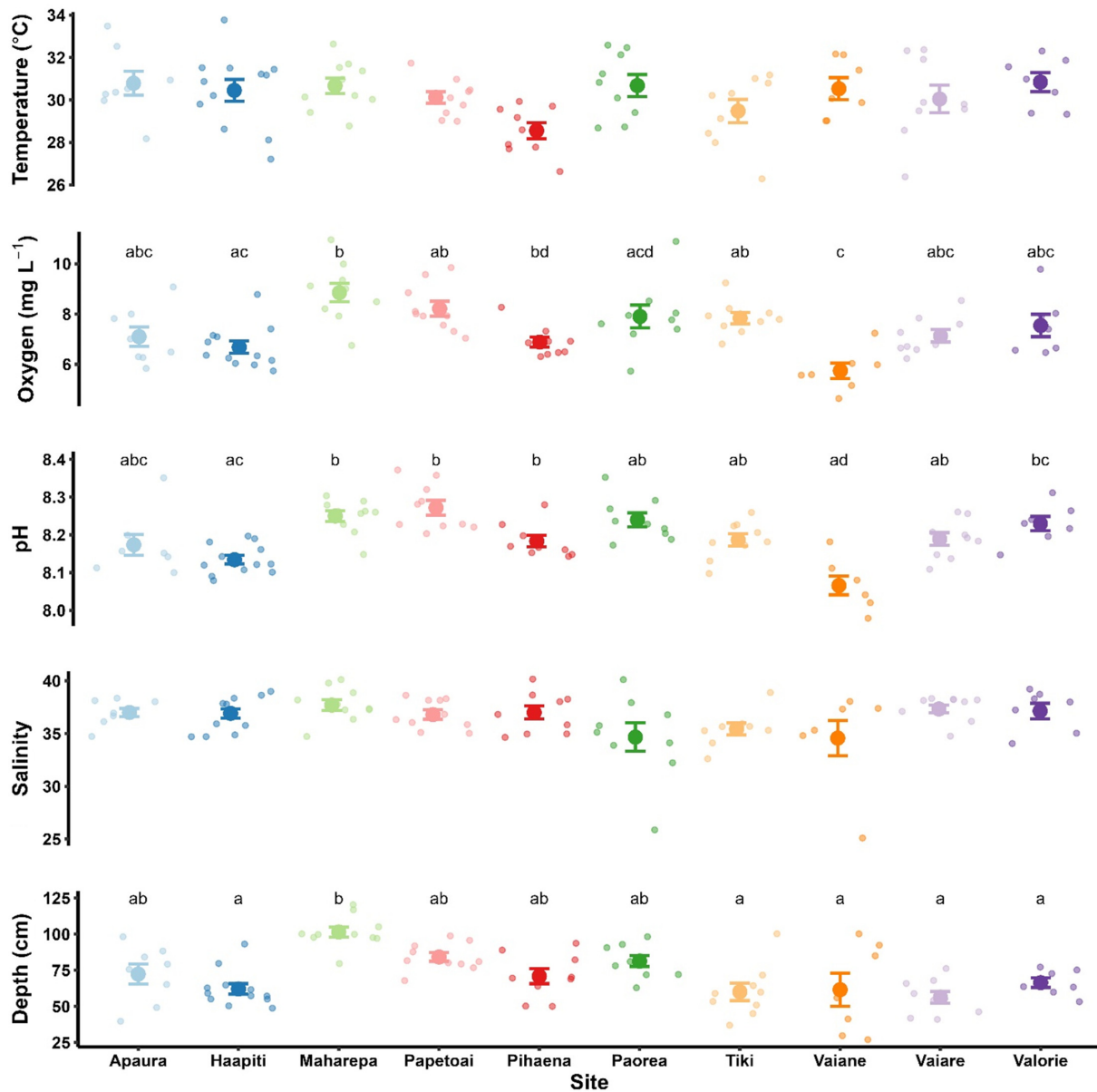


Fig. 3. Abiotic conditions of 10 reef-associated neonate shark sampling sites around Moorea, French Polynesia. Data are presented as means with standard errors of the mean. Individual observations ($n = 7$ – 12) represent values for abiotic conditions recorded at individual sampling events within each site. Different letters represent significant differences in CPUE between sites within each species (Kruskal–Wallis test, $P < 0.05$).

demonstrated that abiotic conditions were not associated with abundance or catch-per-unit effort (CPUE) of *C. melanopterus* and *N. acutidens* neonates. Our findings contradicted other studies on the same species that demonstrated that abiotic conditions, some of the same measured here, do affect abundance and habitat use (Oh *et al.* 2017b; Schlaff *et al.* 2017). Level of anthropogenic impact also did not explain among-site differences in CPUE of either species. As our data suggested, site-attached neonates must have some capacity to tolerate environmental and anthropogenic stressors within the parturition season and within their first months of life

(Bouyoucos *et al.* 2018, 2022). It will be necessary to evaluate long-term trends to determine whether inter-annual variability in abiotic conditions or continued anthropogenic disturbances affect populations (Eustache *et al.* 2024). This type of information will be key to understanding whether additional management or protections (i.e. beyond shark-sanctuary status), especially on temporal (e.g. parturition season) and spatial (e.g. potential nurseries) scales, would affect the abundance of either population.

Abundance of neonates was not influenced by temperature. Although temperature acts directly on the physiology

Table 2. Generalised linear mixed-model outputs describing effects of abiotic conditions on the abundance of neonatal *Carcharhinus melanopterus* and *Negaprion acutidens* around Moorea, French Polynesia.

Response	Parameter	Effect size	95% CI
Abundance of <i>C. melanopterus</i>	Intercept	1.459	−4.88 to 7.79
	Temperature	−0.084	−0.244 to 0.075
	Salinity	−0.019	−0.122 to 0.082
	Depth	−0.006	−0.022 to 0.009
	Lunar phase	0.011	−0.053 to 0.075
	Tide	0.428	−0.031 to 0.89
	Set number	−0.004	−0.014 to 0.004
Abundance of <i>N. acutidens</i>	Intercept	−7.781	−15.471 to −0.092
	Temperature	0.188	−0.029 to 0.406
	Salinity	0.007	−0.081 to 0.095
	Depth	0.002	−0.011 to 0.015
	Lunar phase	0.029	−0.064 to 0.122
	Tide	−0.245	−0.742 to 0.253
	Set number	−0.008	−0.018 to 0.002

Shark abundance data were zero-inflated and analysed assuming a negative binomial distribution. Outputs are presented as mean effect sizes and 95% confidence intervals (CI) of fixed-effect estimates generated from 1000 posterior simulations.

of aquatic ectotherms (e.g. Little *et al.* 2020) and exhibits variable effects on the abundance of reef-associated neonate sharks (Yates *et al.* 2015; Oh *et al.* 2017a; Vierus *et al.* 2018), we did not find that to be the case for *C. melanopterus* and *N. acutidens* around Moorea. Temperature also had only weak effects on habitat use (i.e. residence and kernel area) in other populations of juvenile *C. melanopterus* (Schlaff *et al.* 2017) and *N. acutidens* (Oh *et al.* 2017b) in Australia. Although we sampled only at dusk, the range of water temperatures at which sharks were caught in this study (26.3–33.8°C) matches the range of body temperatures (26.1–34.1°C) reported for free-ranging *C. melanopterus* neonates around Moorea (Bouyoucos *et al.* 2020b). Moreover, thermal tolerance estimates (i.e. critical thermal maximum; CT_{Max}) for *C. melanopterus* ($CT_{Max} = 37^{\circ}C$; Bouyoucos *et al.* 2020b) and *N. acutidens* ($CT_{Max} = 36^{\circ}C$; Bouyoucos *et al.* 2021) exceed the maximum recorded temperatures (i.e. 33.8°C) in this study. Importantly, the abundance of *C. melanopterus* neonates around Moorea was not shown to be affected by seasonal temperatures over 7 years of monitoring (Eustache *et al.* 2024). Together, species-specific CT_{Max} and body-temperature data corroborate the finding of this study of no effect of temperature on shark abundance.

Neonate shark abundance was also not associated with salinity. Rainfall and drought can affect the salinity of seawater in shallow, coastal environments, such that changes in salinity may affect the abundance of young sharks and rays (Chin *et al.* 2010; Knip *et al.* 2010). Indeed, decreases in salinity

have been associated with decreases in abundance of reef-associated neonate sharks (Yates *et al.* 2015; Oh *et al.* 2017a; Vierus *et al.* 2018), including changes in patterns of habitat use in juvenile *C. melanopterus* (Schlaff *et al.* 2017). Rainfall, which is associated with reductions in salinity in shallow, coastal environments, also affected the habitat use (i.e. kernel area) of *N. acutidens* (Oh *et al.* 2017b). Conversely, *C. melanopterus* exhibits no changes in habitat use during tropical storms (Udyawer *et al.* 2013). During this study, sampling was conducted during all weather conditions to achieve a range of salinity (25–40); however, salinity only marginally dropped below that of full-strength seawater (~34) at two sites. However, of note is that the single greatest catch of *N. acutidens* in this study (12.2 sharks h^{-1}) occurred at the Vaiane site immediately after rainfall, which resulted in a salinity of 25. Increases in shark neonate abundance with a decreasing salinity have been previously documented, where Vierus and colleagues observed an increase in the abundance of juvenile blacktip sharks (*Carcharhinus limbatus*) with a decreasing salinity (Vierus *et al.* 2018). Indeed, it has been suggested for young (euryhaline) sharks that salinity preference may differ from that of adults (Simpfendorfer *et al.* 2005; Heupel and Simpfendorfer 2008), and this may be associated with sufficient salinity tolerance to occupy habitats the variable salinities of which may exclude predators, such as estuaries (Morash *et al.* 2016). Although salinity tolerances of neonatal *C. melanopterus* and *N. acutidens* are unknown, salinity did not appear to consistently affect the abundance of neonatal *C. melanopterus* and *N. acutidens* around Moorea.

Dissolved oxygen and pH were both correlated with depth, which was also not a significant predictor of shark abundance despite past studies having found that shark abundance decreases with a decreasing dissolved oxygen (Schlaff *et al.* 2014). However, around Moorea, oxygen concentrations rarely fell below 100% saturation (i.e. environmental hyperoxia), and dissolved oxygen fell only below hypoxia tolerance thresholds (i.e. critical oxygen partial pressures; p_{crit}) in two sampling sets (Bouyoucos *et al.* 2021). Therefore, more severe hypoxia than what is typical of this system may have been necessary to elicit declines in abundance. We have previously demonstrated that neither *C. melanopterus* nor *N. acutidens* are hypoxia tolerant ($p_{crit} = 85$ and 78% saturation respectively; Bouyoucos *et al.* 2021); however, these species may undergo various physiological and behavioural responses to declining dissolved oxygen that could allow them to remain within their habitat, as has been documented for other shark and ray species (Parsons and Carlson 1998; Carlson and Parsons 2003; Dabruzzi and Bennett 2013). Thus, environmental hyperoxia is likely to have masked any potential effects of dissolved oxygen on neonate abundance here, which could also be the case for other species that are characteristically hypoxia intolerant.

The abundance of neonate sharks was also not affected by environmental cues, such as tidal state and lunar phase.

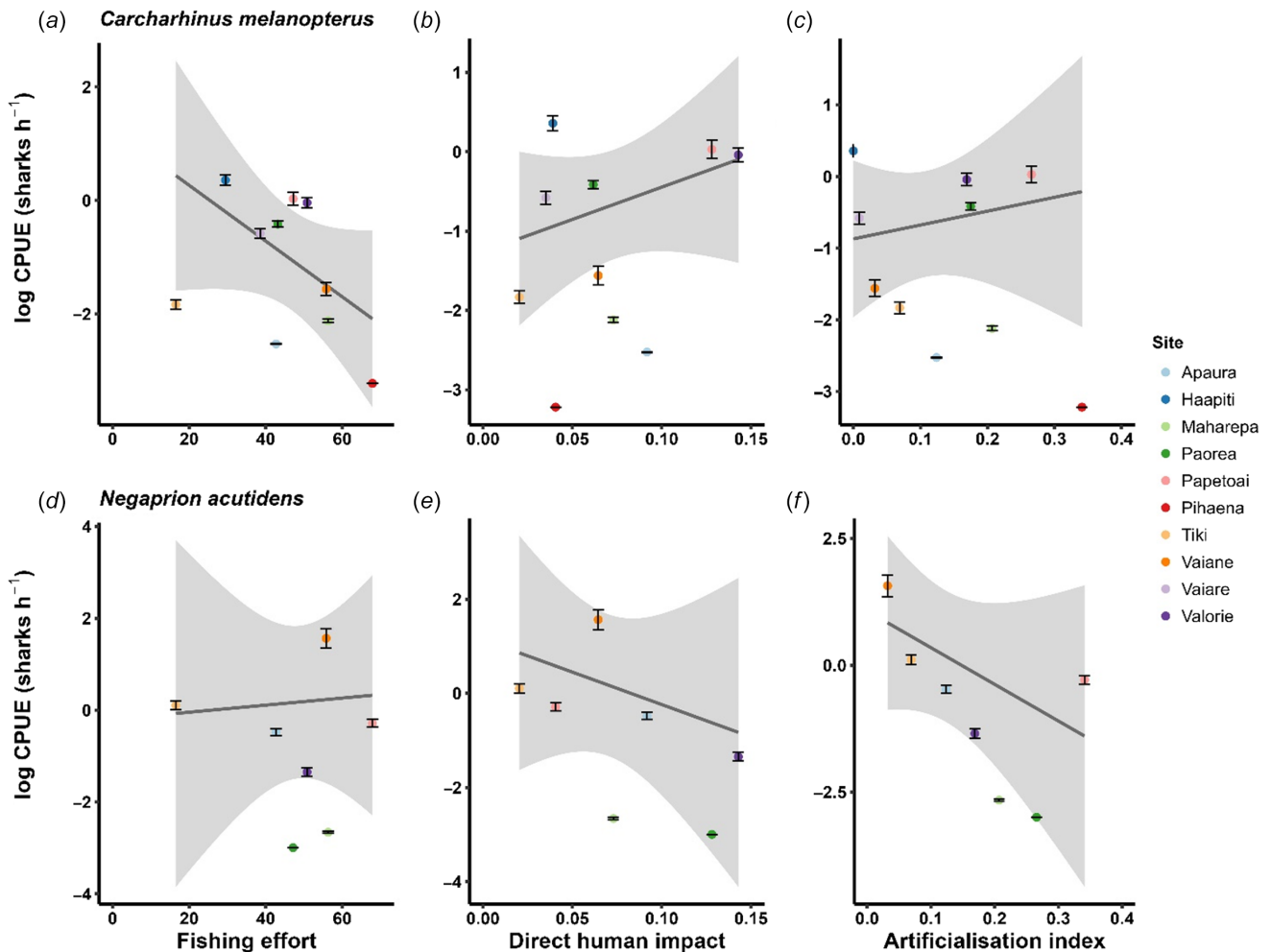


Fig. 4. Associations between various anthropogenic impacts and reef-associated neonate shark abundance at different sites around Moorea, French Polynesia. Individual observations (*Carcharhinus melanopterus* $n = 10$; *Negaprion acutidens* $n = 7$) represent the mean impact value estimated for individual sites and mean catch-per-unit effort (CPUE) for (a–c) *C. melanopterus* and (d–f) *N. acutidens* at each site. Error bars represent standard errors of the mean of CPUE at each site. Shading around regression lines represents 95% confidence intervals. Regression parameters are presented in Table 3.

Although these variables are generally associated with abundance and habitat use in sharks (Schlaff *et al.* 2014), effects can be quite variable, even within habitats (Hansell *et al.* 2018). Sharks in this study were fished on both ebbing and flooding tides and from new to full moon; yet, tidal state was not associated with water depth, which is unsurprising given Moorea's geographic location in the South Pacific where tidal change is quite minimal (Andréfouët and Adjeroud 2019). Elsewhere, neonatal and juvenile *C. melanopterus* and *N. acutidens* individuals follow tidal cues to make use of mangroves for shelter (Oh *et al.* 2017b; George *et al.* 2019). Around Moorea, mangrove cover for sharks to exploit is unsubstantial (Madi Moussa *et al.* 2019), and minimal variation in tide height does not affect the availability of other refugia. Thus, a lack of tidal effects may relate to minimal tidal variability; however, an explanation for the lack of lunar-phase effects remains unclear. Shark abundance and lunar

phase are sometimes associated through foraging (Lowry *et al.* 2007) and prey behaviour (Mourier *et al.* 2016); however, such trends may not be evident in shallow, coastal environments (Hansell *et al.* 2018), or be applicable to neonates (Vierus *et al.* 2018).

Although fishing effort, human impact and coastal artificialisation varied considerably among sites, these parameters did not predict neonate shark abundance around Moorea during the 2018–2019 parturition season. Reef shark populations in French Polynesia appear to be among the most abundant relative to reef shark populations elsewhere (MacNeil *et al.* 2020); however, benefits of local MPAs around Moorea have not been evaluated for sharks (Thiault *et al.* 2019). In the event that sharks are caught as bycatch, delayed mortality is low if *C. melanopterus* and *N. acutidens* are quickly released (Dapp *et al.* 2017; Bouyoucos *et al.* 2018). Neonates with retained fishing hooks are observed around

Table 3. Linear model output describing effects of anthropogenic impact scores on the catch-per-unit effort (CPUE) of neonatal *Carcharhinus melanopterus* and *Negaprion acutidens* around Moorea, French Polynesia.

Response	Parameter	Estimate	Error	t-value	P-value
CPUE of <i>C. melanopterus</i>	Intercept	1.248	1.282	0.973	0.358
	Fishing effort	-0.049	0.026	-1.871	0.098
	Intercept	-1.261	0.576	-2.188	0.065
	Direct human impact	8.139	6.747	1.206	0.267
	Artificialisation index	1.934	3.160	0.612	0.560
CPUE of <i>N. acutidens</i>	Intercept	-0.199	1.944	-0.103	0.923
	Fishing effort	0.008	0.038	0.203	0.849
	Intercept	1.140	1.143	0.998	0.375
	Direct human impact	-13.785	14.393	-0.958	0.392
	Artificialisation index	-7.227	4.427	-1.632	0.178

Moorea (~4 sharks per parturition season; J. Rummer, pers. comm.), which suggests that some fishers may cut the line if a shark is hooked; indeed, this approach has been shown to improve survival in catch-and-release angling scenarios (Fobert *et al.* 2009). Furthermore, low post-release mortality rates suggest that fishing effort might have little effect on neonatal *C. melanopterus* and *N. acutidens* populations (Bouyoucos *et al.* 2018). Importantly, fishing effort, as defined in this study, includes other methods than hook-and-line or net techniques that can catch sharks as bycatch. In adult *C. melanopterus*, there is evidence of learned avoidance of fishing gear (Mourier *et al.* 2017). No studies have directly tested for an effect of trampling as a metric of human disturbance on shark abundance; however, trampling can lead to the destruction of mangrove habitat and coral reefs (Hardiman and Burgin 2010), which reef-associated neonate sharks depend on. Coastal artificialisation also did not directly affect shark abundance in this study; although, coastal artificialisation has been occurring for decades around Moorea (Madi Moussa *et al.* 2019). It may be that coastal artificialisation is not severe enough to affect the ecosystem to affect shark abundance. Conversely, at the time of the present study, shark populations around Moorea may have already been in a degraded state resulting from coastal artificialisation that has been increasing since the 1970s (Madi Moussa *et al.* 2019).

Beyond abiotic and anthropogenic stressors, an important driver of neonate abundance is fidelity to the location where parturition occurs. We have demonstrated through mark-recapture that *C. melanopterus* and *N. acutidens* neonates are not caught at more than one site (Bouyoucos *et al.* 2022; Debaere *et al.* 2023). Further, we have demonstrated that *C. melanopterus* neonates exhibit remarkably small home ranges (Bouyoucos *et al.* 2020a). Thus, site fidelity among

reef-associated neonate sharks around Moorea may limit individuals from moving among sites despite variable abiotic conditions and levels of anthropogenic disturbance. Site fidelity in neonates is then potentially driven by reproductive philopatry by mothers, or the tendency of females to give birth at the same location over multiple years, which has been demonstrated for adult female *C. melanopterus* around Moorea (Mourier and Planes 2013; Eustache *et al.* 2023). Drivers of reproductive philopatry in *C. melanopterus* females are not understood; therefore, the extent to which variable abiotic conditions and levels of anthropogenic disturbance influence reproductive philopatry or explain why some female *C. melanopterus* individuals have been shown to reproduce at single or multiple sites is not clear (Eustache *et al.* 2023). Beyond reproductive philopatry, it is also pertinent to consider the effects of survival and dispersal from parturition sites over the course of each parturition season. Abundance of *C. melanopterus* generally declines over time during parturition seasons around Moorea (Eustache *et al.* 2024), which is suggestive of mortality within this population. Regarding dispersal, there is a notable absence of 1+ year-old reef sharks from abundance surveys of Moorea (Mourier *et al.* 2013a), which suggests that neonates will eventually leave their parturition site and is further suggestive of high mortality among neonates. Indeed, characteristics of the reef shark populations around Moorea (e.g. site fidelity, reproductive philopatry, natural mortality, etc.) are among myriad possible drivers of the abundance of neonates.

In conclusion, the abundance of neonate subpopulations of two reef shark species during the 2018–2019 parturition season was unaffected by abiotic conditions and historical levels of anthropogenic impact. Regarding anthropogenic impact, it is not clear from the present study whether shark neonate populations are resilient to the stressors that were investigated, or whether the populations are already in a degraded state owing to decades of impact (e.g. coastal artificialisation). Thus, a key area for future research will be establishing paired time series of anthropogenic impacts and shark abundance to examine temporal trends over larger scales. Ongoing monitoring of anthropogenic impacts on shark populations within shark nursery areas will be particularly important because the restricted distribution of the populations requires that they are tolerant of the sometimes-extreme abiotic conditions of their habitat. Indeed, animals cannot adapt to habitat loss, which is a strong driver of declines in reef shark and ray populations (Sherman *et al.* 2023), and a mechanism through which climate change is threatening sharks and rays (Dulvy *et al.* 2021). Because coastal and reef-associated shark species are thought to be among the most vulnerable to environmental change (Chin *et al.* 2010), efforts to mitigate compounding anthropogenic stressors in these environments will be important. As such, the present study has potential broader impacts on shark management within French Polynesia, which accounts for 25%

of the area of all MPAs for sharks (i.e. 'shark sanctuaries'), and similar systems through the implementation of quantitative techniques to assess nursery-area use in two widely distributed Indo-Pacific reef shark species.

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Data availability. Data presented in this paper are available at the Research Data Repository (Tropical Data Hub) at James Cook University. <https://doi.org/10.25903/25FB-QX79>.

Conflicts of interest. C. A. Simpfendorfer is an Associate Editor for *Marine and Freshwater Research*. Despite this relationship, he did not at any stage have editor-level access to this manuscript while in peer review, as is the standard practice when handling manuscripts submitted by an editor to this journal. *Marine and Freshwater Research* encourages its editors to publish in the journal and they are kept totally separate from the decision-making process for their manuscripts. The authors have no further conflicts of interest to declare.

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