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1 *Local environmental context structures animal-habitat*
2 *associations across biogeographic regions*

3 **Short title: Fish-habitat context**

4
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35 manuscript, MB led the field sampling, video analysis, statistical analysis and writing.

36

37 **Abstract**

38 The mechanisms that govern fauna-habitat associations across multiple spatial scales
39 remain largely undefined. Can environmental factors structure fauna-habitat associations
40 over both local and global spatial scales, alongside biogeographical processes and patterns?
41 We compare the extent to which the use of mangroves by fishes is consistent within and
42 between biogeographic locations, and whether any similarities and differences can be
43 attributed to the environmental context of those forests, such as the physical environment,
44 seascape composition, and constraints on access by fishes. We focus on three important
45 proxies of these structuring forces for fish – salinity, distance to reefs, and tidal amplitude.
46 Using directly comparable remote underwater visual census from a range of diverse
47 environmental contexts in the Central and Eastern Indo-Pacific, we examine similarity in the
48 family-level taxonomic composition of fish assemblages in mangrove forests. Local
49 environmental context appears to explain similarities and differences in mangrove
50 association by fishes at both regional and local scales across the Indo-Pacific. There were
51 strong consistencies in taxonomic composition in similar environmental contexts despite
52 geographic separation. Tidal amplitude was a powerful explanatory factor that interacted
53 with both distance to reef and salinity in partitioning variation in fish assemblage structure.
54 Substantial differences in the use of mangroves between regions appear to be independent
55 of historical biogeography, relating instead to local context. Our findings suggest that the
56 effects of local context on habitat suitability can play out over biogeographical scales, and
57 global similarities in fauna-habitat associations may be partially explained by comparable
58 environmental contexts, with important management implications.

59

60 **Keywords:** habitat, spatial scale, nursery, context, setting, coastal ecosystem, seascape,
61 juvenile fish.

62

63 **Highlights:**

- 64 • Local environmental context can determine fauna-habitat associations
- 65 • Fauna-habitat relationships are consistent in similar environmental settings
- 66 • Differences in fauna relate strongly to tidal, seascape and physical factors

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82 **Introduction**

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84 Contrasting ecosystem functioning and values can arise from geographic variation in fauna-
85 habitat relationships. This variability can result in diverging notions of the importance of
86 different habitats, and the impact of their removal or modification on species survival and
87 community maintenance (Randin and others 2006; Zanini and others 2009). The
88 mechanisms that govern similarities in fauna-habitat relationships across multiple spatial
89 scales remain largely undefined (Bradley and others 2020). Historic phylogeographic
90 mechanisms (i.e. the available species pool) determine large-scale patterns in fauna-habitat
91 relationships across the globe, while environmental and ecological mechanisms (e.g.
92 ecophysiological tolerances, predator-prey dynamics) determine habitat use at smaller
93 spatial scales. However, large-scale patterns must also reflect the outcome of fine-scale
94 interactions. Thus large-scale patterns in habitat use can also be driven by environmental
95 conditions which vary over multiple spatial scales (Wiens and Donoghue 2004), such as
96 temperature and rainfall (Whitehead and others 1992).

97

98 Variability in fauna-habitat relationships can often be caused by environmental context-
99 dependence, whereby a suite of interacting environmental factors (physical, ecological,
100 spatial and temporal) provide the context for the use of certain habitats, driving different
101 habitat use under different environmental contexts (Bradley and others 2020). The physical
102 environmental context can interact with abiotic constraints of fauna, leading to changes in
103 habitat use that play out over multiple spatial scales. For instance, in terrestrial ecosystems,
104 the use of vegetative structure can depend on climatic gradients, due to the

105 thermoregulatory requirements of fauna. Amphibian and reptile fauna modify their use of
106 tree habitat, shifting from ground-dwelling to arboreal depending on climatic regimes
107 (Adolph 1990; Scheffers and others 2013). These patterns, observed at local scales within
108 forests, also explain variation in habitat associations over larger biogeographic scales
109 (Scheffers and others 2013). Availability of relevant habitat features in the landscape, and
110 constraints on access, will also modify local habitat use. For instance, in Africa's savanna
111 ecosystems, fauna require access to drinking water, with distance to waterholes a key
112 determinant of habitat use (Redfern and others 2003; Roever and others 2012). These
113 access requirements can interact with variation in the physical environment – for example,
114 under wetter conditions, elephants are no longer constrained by distance to waterholes
115 (Roever and others 2012). Additionally, the use of habitat can be determined by predation
116 risk (Brown 1999), and all of these physical, ecological and landscape factors of local
117 environmental context come together to determine habitat use by African herbivores in the
118 presence of lions (Valeix and others 2009). In some ecosystems, local environmental forces
119 appear to exert such a great influence in determining patterns of fauna-habitat associations
120 (Igulu and others 2014) that some large-scale, global differences in habitat use may be
121 partially explained by differences in local environmental context.

122

123 In this study, we ask whether environmental context can be an important driver of habitat
124 use by animals at both local and regional scales. The association between fish fauna and
125 mangrove habitat provides a useful model system for testing this idea. Mangroves occupy a
126 wide range of environmental settings (Worthington and others 2020) – from sandy reef flats
127 to freshwater swamps, and are distributed throughout the tropics and subtropics worldwide
128 (Nagelkerken and others 2008). The fish taxa that use these forests are known to vary

129 considerably, both at finer scales such as between bays and islands in a single region
130 (Kimirei and others 2011), and at regional scales such as between nearby regions (Thollot
131 1992), and between biogeographic realms (Hemingson and Bellwood 2018). As a result, the
132 ecological role of mangrove forests for fish has long been the subject of international
133 debate (Nagelkerken 2009; Sheaves 2017). This variability complicates our understanding of
134 mangrove-fish relationships, and our ability to employ these relationships, both scientifically
135 (Faunce and Layman 2009) and in conservation and environmental management (Sanchirico
136 and Mumby 2009).

137

138 Both environmental context and historic phylogeography will shape the taxa that use
139 mangrove forests, but their relative importance will depend on spatial scale and taxonomic
140 resolution. While phylogeography determines the available species pool across the globe,
141 at the broader family level, phylogeography is less likely to have substantial impact within
142 biogeographic realms with a shared evolutionary history, such as the Indo-Pacific (Cowman
143 and Bellwood 2013). In these situations, it might be possible to tease out the effects of
144 environmental context without the orthogonal influence of phylogeographic history.

145

146 The physical environment, landscape-scale habitat requirements, and constraints on access
147 by fauna, all strongly influence the use of coastal habitats by fishes in general, and appear
148 particularly important in determining the use of mangroves. Firstly, salinity can structure
149 fish assemblages throughout the coastal zone due to differing salinity tolerances (Harrison
150 and Whitfield 2006; Whitfield and others 2006). Salinity can determine the fish assemblages
151 found in mangroves at local scales (Ley and others 1999), and appears to be a major
152 determinant of mangrove habitat function globally (Igulu and others 2014). Secondly, the

153 surrounding seascape can have a profound influence on how fish use coastal habitats, with
154 connectivity or distance between mangroves and reefs of particular importance
155 (Nagelkerken and others 2012; Olds and others 2012) due to the ontogenetic use of
156 mangroves by reef fish (Dorenbosch and others 2007; Jones and others 2010; Pittman and
157 others 2007). Finally, the tidal amplitude experienced in a location can profoundly shape
158 how intertidal habitats like mangroves are used by fish (Sheaves 2005). In areas with small
159 tidal amplitudes, fish are able to use mangrove habitat continuously (Dorenbosch and
160 others 2007), whereas in areas of large tidal amplitudes, mangroves are exposed at low tide,
161 and the duration of exposure can vary substantially among locations (Baker and others
162 2015). To use these intertidal habitats, fish are forced to perform potentially risky intertidal
163 migrations (Dorenbosch and others 2004; Hammerschlag and others 2010; Unsworth and
164 others 2007). The magnitude of inundation can shape the use of intertidal habitat within
165 locations (Castellanos-Galindo and Krumme 2013; Castellanos-Galindo and Krumme 2015;
166 Minello and others 2012; Rozas 1995), and is thought to be a major driver of the use of
167 mangroves by fishes globally (Igulu and others 2014). While each of these three drivers have
168 received considerable attention individually, it has been difficult to study how their
169 interplay shapes fish assemblages. This is because field studies have tended to standardise
170 variation in one or more of these variables through site selection, and because the array of
171 different sampling techniques used in different locations precludes a robust understanding
172 across different environmental contexts.

173

174 Little is known about how environmental factors interact to determine fauna-habitat
175 associations over multiple spatial scales. Until now, there has been no evaluation of the
176 combined influence of salinity, distance from reef and tidal regime on patterns in coastal

177 fauna-habitat associations across regions using directly comparable data. Therefore, the aim
178 of this study was to understand the role of environmental context in defining habitat
179 associations at multiple spatial scales in the Central and Eastern Indo-Pacific. While there
180 are differences in species distributions across this realm driven by biogeographic forces, at
181 the family level, most taxa are present throughout the realm, due to a shared evolutionary
182 history (Cowman and Bellwood 2013). By examining patterns in family level use of
183 mangrove forests, we aimed to examine broad similarities and differences in mangrove use
184 that could not be attributed solely to biogeography.

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186

187 **Methods**

188

189 **Study sites:**

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191

192 Replicate study sites across five locations (Fig. 1) were selected to approximate the range of
193 contextual variation experienced in mangrove habitats across the Indo-Pacific (Table 1).
194 These locations covered almost the entire spread of tidal amplitudes in the Indo-Pacific,
195 from 0.5 m to 11 m. Within each location, efforts were made to cover the full range of
196 variation present in distance between mangroves and reef, and in salinity. Some locations
197 have larger variation in distance to reef than others due to the underlying geomorphological
198 composition of the seascape. Similarly, some locations had larger variation in salinity than

199 others, due to differences in rainfall. The amphidrome point of the South Pacific lies far
200 from any major landmasses (Luther and Wunsch 1975), meaning that areas with the
201 smallest tidal regimes lack extensive areas of mangroves far from reefs, so the extremes of
202 both variables were partially confounded in this analysis. In addition, salinity is partially
203 correlated with distance to reef, as reef building coral are associated with marine salinities.
204 These imbalances in the dataset (see supplementary material, Fig. S1) are representative of
205 natural variation, and do not violate any underlying assumptions in the exploratory analyses
206 that we employed. All results were interpreted with these imbalances in mind.

207

208 **Fish Surveys**

209

210 We studied fish assemblages using remote unbaited video census. Video surveys were
211 conducted throughout the day and tidal cycle, in an attempt to adequately represent
212 natural variability. Cameras were deployed inside mangrove habitat structure, within the
213 first 2 metres of the seaward edge of the mangrove forest, facing mangrove structure at an
214 approximate distance of 1 m. This standardised the part of forest habitat sampled (edge),
215 and the effective sampling distance of each video recording (~1 m). Cameras were placed at
216 a minimum distance of 20 m, but usually much farther (>50 m) from each other to ensure
217 sample independence. Cameras were left undisturbed to record while the vessel maintained
218 a distance of at least 100 m. In-water visibility distance was determined using a patterned
219 image 0.5 m from the lens, and videos below the 0.5 m threshold were discarded. This
220 produced 389 video samples for analysis. A 15 minute section of video was analysed from
221 each deployment, which provides an optimal and consistent comparison of species

222 assemblages between deployments (Piggott and others 2020). Videos were viewed to
223 record the taxa present, and fish identifications were reviewed by relevant experts. Only
224 presence/absence of families was used in the following analyses. Presence data are more
225 robust to variability in visibility, and the fact that differential presence of schooling and
226 solitary species within families could make abundance data non-comparable between
227 locations.

228

229 **Explanatory variables**

230

231 Salinity

232

233 Salinity surveys were carried out at each location during each sampling trip using either a
234 refractometer or salinometer, measured in psu. These surveys were conducted concurrently
235 with video sampling, but measurements were not taken for each video recording
236 individually. Therefore, salinity values represent the shallow water readings associated with
237 the local area (within 100 m), rather than individual position of each video site. As sampling
238 occurred outside the monsoon season for each location, we are making the assumption,
239 based on previous research carried out in these locations (Aharon 1991; Baker and others
240 2018; Langer and Lipps 2006; Williams and others 2006; Wolanski and others 1990;
241 Wolanski and Spagnol 2003) that waters were reasonably well mixed and salinity was not
242 fluctuating dramatically during the sampling period, making these values adequately
243 representative.

244

245 Distance to reef

246

247 Distance to reef was measured using the shortest path through water between the video
248 site, as recorded by GPS position, and the nearest visibly identifiable patch of biogenic reef
249 from satellite imagery. A variety of satellite sources were used to avoid cloud cover and
250 view areas under low wind and low tide conditions to maximise the possibility of reef
251 detection, including Google earth, ARC GIS and DigitalGlobe (<http://www.digitalglobe.com>
252 accessed August 2018). Ground sample distance (i.e. pixel resolution) ranged between 65
253 cm and 39 cm. While this method may miss minor patches and deeper reefs, it provides a
254 good indication of the distance from a point to the nearest substantial photic-zone reef.

255

256 Tidal amplitude

257

258 The maximum tidal amplitude of the location in which the video sample was taken was used
259 as an explanatory variable. Values were derived from tidal gauge data, obtained from
260 Australia's National Tidal Centre and University of Hawaii's Sea Level Centre (Caldwell and
261 others 2015).

262

263 **Multivariate analyses**

264

265 Family-level assemblage composition (presence/absence) was calculated for each site.

266 While there are differences in the species pool in the different locations sampled, all broadly
267 contain the same families of near-shore fishes (Sheaves 2012). Family-level taxonomic

268 assemblage composition is a meaningful ecological measure, because the family level
269 taxonomy of tropical fishes is broadly related to functional morphology and ecology. Broad
270 trophic ecology (e.g. herbivory vs carnivory) and some broad ecosystem associations (e.g.
271 reef vs estuary) are often persevered at the family level. Therefore, we used family-level
272 taxonomic assemblage to obtain a comparable metric by which to assess differences in the
273 use of mangrove forests by fish throughout the broad geographic area encompassed by our
274 study locations. Not all taxa were included in analysis at the family level; fishes from the
275 families Gobiidae and Blenniidae could not be reliably distinguished in many video samples,
276 therefore the order Gobiiformes was used. Similarly, fishes from the families Atherinidae,
277 Clupeidae and Engraulidae could not be reliably distinguished, and the order Clupeiformes
278 was used. The distinction between Scarinae and other labrid fishes was maintained due to
279 their broadly divergent functional ecologies.

280

281 Multidimensional scaling ordination

282

283 To display variation in fish assemblage composition, and identify the taxa that characterised
284 this variation, non-metric multidimensional scaling (nMDS) ordination was employed. Non-
285 metric ordination was chosen because it uses the internal order of the entire dataset to
286 display similarities and differences between samples without any presupposed
287 stratification. While information about the magnitude of distances is lost, rank-based
288 methods are generally more robust to data which do not have an identifiable distribution,
289 and also avoids the assumption of linear patterns inherent in metric ordinations, for which
290 there is no particular basis in this study. Multivariate data that is dominated by absences
291 can present problems during multivariate analysis due to the undue influence of rare taxa

292 and empty sites (Clarke and others 2006). Therefore, rare taxa are often excluded from
293 multivariate analyses. However, taxa that are rare at the level of the entire dataset can still
294 be important, as they might be common in one location and absent in others. To ensure
295 these potentially important contributions were included, taxonomic assemblage data were
296 treated in two different ways. In the first more conventional treatment, an arbitrary
297 dataset-wide cut-off was used to determine the inclusion of taxa in the analysis (e.g. Davis
298 and others 2014). All taxa present in less than 10% of total sites were excluded, which
299 removed all taxa with low occurrences regardless of location-specific occurrence rates,
300 providing a dataset with minimal influence of rare taxa. In the second, more inclusive
301 treatment, taxa present in at least 10% of sites from any one location were retained in the
302 analyses. This only removed taxa that were rare throughout all locations, and provided a
303 dataset where all non-trivial differences between locations would be maintained. This
304 treatment resulted in substantial outliers in subsequent analysis. These outliers were
305 excluded from the dataset, providing a third and final data treatment.

306

307 To examine the variation in taxonomic assemblage composition in all treatments, Jaccard's
308 coefficient similarity matrices were calculated on the binary presence or absence of each
309 taxonomic group in each site. nMDS applied to these similarity matrices was used to display
310 patterns of multivariate variation graphically. This was performed using the 'metaMDS'
311 function in package Vegan (Oksanen and others 2013) in R. The conventional treatment
312 produced a two-dimensional MDS solution with an acceptable (<0.2) level of stress (see
313 supplementary material, Fig. S2). The inclusive treatments resulted in high stress two-
314 dimensional MDS solutions (see supplementary material, Fig. S3 and S4), therefore lower
315 stress three-dimensional solutions were produced, and the first two dimensions (which

316 capture the bulk of variation in the data) were examined. Due to general agreement
317 between the different data treatments, only the results of the inclusive treatment excluding
318 outliers (i.e. the third data treatment described above) is presented in the main text. The
319 results of the other two treatments are provided in the supplementary material (Fig. S5
320 through to S8).

321

322 Multivariate Regression Tree Analysis

323

324 To examine whether the broad variation in taxonomic assemblage composition could be
325 statistically explained by environmental context, a multivariate regression tree (MRT)
326 analysis was performed with the 'mvpart' package in R (De'Ath 2007; Ouellette and
327 Legendre 2012). This technique recursively partitions the dataset into homogenous
328 subgroups using explanatory variables (De'Ath 2002), in this case the three descriptors of
329 environmental context: salinity, distance to reef and tidal amplitude, as well as
330 biogeographic province (central vs eastern Indo-Pacific) following Bowen and others (2016).
331 For each partition, the machine learning analysis considers all four explanatory variables,
332 and selects the variable that maximises the reduction in group heterogeneity. In the final
333 tree, the variables that were selected, and how the dataset was partitioned, shows how
334 these variables explain assemblage structure when considered together. The importance of
335 each contextual variable in explaining variability in the tree model was also calculated based
336 on the reduction in mean squared error by each candidate variable at each split, using the
337 'caret' package in R (Kuhn 2012). Regression tree approaches are robust to the co-variation
338 in explanatory variables as described above (Table 1, supplementary material, Fig. S1). The
339 tree examined was based on the inclusive data treatment excluding outliers, as described

340 above, in order to preserve locational differences and exclude the undue influence of
341 outliers. The analysis was based on the same multivariate taxonomic similarity matrix used
342 in the nMDS analysis above.

343

344 Surface fitting

345

346 To examine how each environmental contextual variable related to the variation in
347 taxonomic assemblage composition observed, surface fitting of explanatory variables was
348 applied to the nMDS ordinations. This technique uses generalized additive models to fit a
349 smooth surface, with the degree of smoothing determined by generalised cross validation,
350 and was carried out using the 'ordisurf' function in package Vegan (Oksanen and others
351 2013) in R. This technique is appropriate for examining relationships between
352 environmental variables and assemblage composition when relationships are not
353 necessarily linear, as is assumed in vector fitting.

354

355 **Results**

356

357 In total, 45 different family-level fish taxa occurred in mangrove habitats throughout our
358 study locations. Most taxa were only present in mangroves in certain locations, with the
359 exception of Carangidae, Gobiiformes and Lutjanidae, which were present in mangroves
360 throughout the Indo-Pacific (see supplementary material, Table S1). Few of these families
361 were commonly encountered, with only 9 families occurring in greater than 10% of total
362 sites.

363

364 **Variation in mangrove fish assemblages**

365

366 Overall, there was substantial variation in family-level composition of sites both within and
367 between locations. There was a broad separation of sites into two location-based clusters –
368 sites from French Polynesia and Papua New Guinea vs sites from Australian locations. Sites
369 from French Polynesia and Papua New Guinea were characterised by Scarinae, Labridae,
370 Acanthuridae, Chaetodontidae, Pomacentridae, Ostraciidae, Zanclidae, Mullidae,
371 Nemipteridae and Apogonidae. These were separated from the majority of sites from
372 Northern Australian locations, which were characterised by Gobiiformes, Gerreidae and
373 Sparidae. Differences between locations appear to be an important contributor to the
374 overall variation of mangrove fish assemblages across the Indo-Pacific. There was also clear
375 overlap of the locations within each clusters, and some overlap of these two clusters (Fig. 2).
376 This indicates that a substantial portion of the variation in fish assemblages is not
377 explainable in terms of regional difference, and occurs within each location individually.
378 Many sites were more similar to sites in other locations than to the majority of the sites
379 from their own location, demonstrating that mangrove assemblages can be quite similar
380 between locations. The magnitude of variation between locations is similar to the
381 magnitude of variation between sites within locations. North Eastern Australia had the
382 widest spread of sites, covering the entire ordination space. This demonstrates that
383 mangrove assemblages can be highly variable within a location, and implies that this
384 variation can equal the entire span of regional variation. Only the final multivariate data
385 treatment is presented in detail (Fig. 2). This treatment included taxa that were common

386 within locations, but excluded outliers. Both the conventional data treatment (see
387 supplementary material, Fig. S5) and the inclusive treatment prior to removal of outliers
388 (see supplementary material, Fig. S6) captured these same general patterns within and
389 between locations.

390

391 **Potential contextual drivers of variation in mangrove fish** 392 **assemblages**

393

394 All three potential contextual drivers were selected by multivariate regression tree analysis
395 as factors that reduced heterogeneity in the dataset, i.e. explained the structure of the fish
396 assemblage. The resulting tree structure (Fig. 3) indicates the primacy of tidal amplitude in
397 defining mangrove fish assemblages. The major division in the dataset was between small
398 (<2 m, microtidal) and large (>2 m, meso- and macro-tidal) tidal amplitudes. Within sites
399 from small tidal regimes, variability was best explained by the tide again, effectively
400 distinguishing between sites from Polynesia (eastern Indo-Pacific, 0.5 m maximum tidal
401 range) and New Britain (central Indo-Pacific, 0.9 m maximum tidal range). As a location-
402 scale variable, tidal amplitude is confounded with differences in biogeography. However,
403 this analysis indicates that assemblage structure is consistent with the magnitude of tidal
404 amplitude, across biogeographic provinces (central vs eastern Indo-Pacific). Within sites
405 from large tidal amplitudes, the biggest difference between assemblage composition was
406 between sites close to reefs (<560 m) and sites far from reefs (>560 m). Sites in large tidal
407 amplitude locations close to reefs varied most according to tidal amplitude again. Sites from

408 large tidal amplitudes far from reefs diverged most according to salinity, with sites in
409 salinities below 28 (i.e. brackish waters) distinct from those in salinities above 28 (i.e. near-
410 marine and marine waters). Variable importance values (Table 2) indicated that tidal
411 amplitude was the most important factor in explaining variability in the dataset. Salinity and
412 distance to reef were both equally capable of partitioning variation, and biogeographic
413 province less so. These results indicate that tidal amplitude can explain a large amount of
414 variation in assemblage composition between sites, and that the effects of distance to reef
415 and salinity are highly interactive with tidal amplitude.

416

417 Surface fitting was used to model the gradient of change for each of the three explanatory
418 context variables across the ordination space (Fig. 4, 5 and 6). Salinity, distance to reef and
419 tidal amplitude were all significantly correlated with the ordination ($p < 0.001$), meaning that
420 the distribution of sites in the ordination was in some way concordant with changes in
421 values of each of these variables. The variation explained by each variable differed
422 markedly. Patterns were consistent across all three different data treatments for the
423 exclusion of rare taxa (see supplementary material, Fig. S7 and S8).

424

425 Tidal amplitude explained 61% of the variation in the ordination (Fig. 4). The gradient in tidal
426 amplitude is generally linearly correlated with dimension 1, mirroring the broad differences
427 in locations captured across the ordination space (Fig. 4a). The strong gradient apparent on
428 the surface plot between micro-tidal (Polynesia and New Britain) and macro-tidal regimes
429 (Northern Australia), indicates that tidal amplitude explains differences in this part of the
430 space, but the lack of a gradient over Northern Australian sites indicates that other factors
431 are important in this part of the space. This is consistent with the multivariate tree (Fig. 3).

432 Scarinae, Labridae, Acanthuridae, Chaetodontidae, Pomacentridae, Ostraciidae, Zanclidae,
433 Mullidae, Nemipteridae and Apogonidae appear to be associated with small tidal
434 amplitudes and Gobiiformes, Gerreidae and Sparidae appear to be associated with large
435 tidal amplitudes (Fig. 4b).

436

437 Distance to reef explained 47% of the variation in the ordination (Fig. 5). The surface
438 appears to explain variation both within and between locations, with the gradient running
439 diagonally across the space, varying substantially over both dimensions 1 and 2 (Fig. 5a).

440 Gobiformes, Sparidae, Gerreidae, Tetraodontidae, Ambassidae, Toxotidae, Carangidae,
441 Terapontidae and Scatophagidae appear to be associated with large distances from reefs,
442 and Scarinae, Labridae, Acanthuridae, Chaetodontidae, Pomacentridae, Ostraciidae,
443 Zanclidae, appear to be associated with small distances from reefs (Fig. 5b).

444

445 Salinity explained 14% of the variation in the ordination space (Fig. 6). Variation across
446 much of the space is likely driven by tidal range and distance to reef, as indicated by the
447 multivariate regression tree (Fig. 3), where salinity was the most important discriminator

448 only within sites far from reef in large tidal amplitude locations. Salinity appears to
449 predominantly describe variation that occurs within locations rather than between them

450 (Fig. 6a), with the gradient mainly occurring along dimension 2. Sparidae, Gerreidae,

451 Tetraodontidae, Ambassidae, Toxotidae, Carangidae, Terapontidae and Scatophagidae

452 appear to be associated with brackish salinities, and Gobiformes, Lethrinidae, Scarinae,

453 Labridae, Acanthuridae, Chaetodontidae, Pomacentridae, Ostraciidae, Zanclidae, Mullidae

454 and Nemipteridae associated with higher salinities (Fig. 6b).

455

456 Despite natural imbalances in the dataset, each variable appears to explain a somewhat
457 different component of variation in the ordination space. Gradients occurred along different
458 axes, reflecting the separation between different sets of species. However, both distance to
459 reef and tidal range explained variation occurring along dimension 1, indicating correctly
460 that they co-varied across our study locations (see supplementary material, Fig. S1). As a
461 result, these three variables explain more than 100% of the variation in the ordination
462 space, i.e. variation is not strictly partitioned between them. Due to the natural structure of
463 variability in our study systems, sites from micro-tidal Polynesia were all close to reef (<500
464 m), whereas the majority of sites from macro-tidal North Central and North West Australia
465 were far from reefs (>500 m). Thus, differences between locations (and thus tidal regimes)
466 cannot be clearly distinguished from differences in distance to reef in the overall analysis.
467 However, by examining the sites that are close to reefs (<500 m) across all locations (Fig. 7),
468 the potential effects of locational differences in tidal range can be observed independently
469 of the effects of distance from reef. In the ordination, the arrangement of points is
470 consistent with the directional gradient of tidal range across the space.

471

472 **Discussion**

473

474 Environmental context explained variation in the habitat associations of fishes at both
475 regional and local scales across the locations sampled in the central and eastern Indo-
476 Pacific. Faunal composition at the family level was clearly not location specific, and locations
477 showed considerable overlap in assemblage composition. Instead, the response of fish
478 fauna was largely consistent with environmental variation. Locations with broadly similar

479 tidal amplitudes showed considerable overlap in fish assemblage composition. While the
480 natural structuring of environmental contexts in the Indo-Pacific meant that tidal range
481 covaried with proportion of sites close to reef in the dataset, the fish assemblage of sites
482 close to reefs at each location were different, and these differences largely aligned with
483 tidal magnitude. While this is not conclusive evidence that tidal mechanisms are responsible
484 for these interregional patterns, tidal amplitude has been identified as a potential global
485 determinant of the use of mangroves by fish (Faunce and Layman 2009; Igulu and others
486 2014), and results from our study supports this hypothesis. The variation seen within
487 locations with broadly similar tidal regimes was explained by both salinity gradients and
488 seascape structure. This is consistent with the findings of other mangrove studies from
489 across the Indo-Pacific, which together show that geographically distant mangrove forests in
490 similar environments can share similarities in fish taxa. In coastal marine environments close
491 to reefs, mangroves tend to be characterised by reef associated taxa, including families that
492 occur in mangroves as juveniles, such as Nemipteridae, Scarinae and Labridae, and those
493 that more typically occur as adults such as Chaetodontidae (Barnes and others 2012;
494 Unsworth and others 2009, Nagelkerken and Van der Velde 2004). In brackish environments
495 far from reefs, mangroves tend to be characterised by coastal and estuarine taxa such as
496 Ambassidae and Toxotidae (Blaber and others 1989, Sheaves and others 2016). This is not
497 the only important axis of variation in these two factors. In large tidal amplitude locations,
498 in mangroves far from reefs, there was an important difference in fish fauna based on
499 salinity alone. This is consistent with the idea that salinity plays a key role in structuring fish
500 fauna in coastal and estuarine environments (Blaber 2008; Bradley and others 2019; Ley et
501 al 1999; Weinstein and others 1980). These results are consistent with the idea that
502 variation in habitat use is driven by specific environmental conditions, and demonstrate that

503 in the Central and Eastern Indo-Pacific, abiotic environmental context can largely explain
504 differences in the use of coastal habitats by fauna.

505

506 The substantial differences in family-level mangrove fish assemblage composition observed
507 at the locations sampled in the Central and Eastern Indo-Pacific appear to be broadly
508 independent of historical biogeography. The major biogeographic division in our dataset
509 was between the biogeographic provinces of the East Pacific (represented by Polynesia) and
510 Central Indo-Pacific (all other locations), as defined by Bowen and others (2016).

511 Assemblages in mangrove forests did not diverge according to biogeographical province in
512 either analysis, but instead converged according to similarities in tidal amplitude, distance
513 to reef, and salinity. While the use of family-level taxonomic composition intentionally
514 avoids much of the differences between locations caused by species distributions across the
515 Indo-Pacific (Sheaves 2012), some of the families that distinguished our North Central
516 Australian and North Western Australian sites are entirely absent from French Polynesia –
517 namely Sparidae and Gerreidae (Siu and others 2017) – so their absence from mangroves
518 there can be directly attributed to their absence from the species pool at these locations
519 due to biogeographic history. However, these families are common in New Britain (Froese
520 and Pauly 2018), and their absence from mangroves there serves to distinguish mangrove
521 sites in this location from those with larger tidal ranges. Importantly, none of the families
522 that distinguished small tidal regime locations (Polynesia and New Britain) were unique to
523 these regions – they are all common reef taxa found across biogeographic provinces in all
524 locations sampled (Froese and Pauly 2018), yet they were absent from mangroves in all
525 larger tidal regimes (northern Australian locations). This demonstrates that there are strong
526 commonalities in mangrove assemblages found in low tidal amplitudes close to reefs, in

527 high tidal amplitudes close to reefs, and high tidal amplitudes far from reefs, that are
528 independent of historical biogeography.

529

530 The mechanisms by which salinity, seascape structure and tidal range might interactively
531 determine the use of coastal habitats by fish are not well understood. It is possible that the
532 direct ecological costs and benefits of inhabiting mangroves shape fish assemblages in
533 regions where mangrove roots are routinely exposed during the tidal cycle. Regular
534 migration between mangroves and suitable low tide habitat are likely to involve increased
535 risk of predation (Gilliam and Fraser 2001), and incur a substantial energetic cost (Alexander
536 2002, Bernatchez and Dodson 1987). The duration, frequency and depth of inundation of
537 intertidal habitats all vary with tidal amplitude (Baker and others 2015; Minello and others
538 2012), which will directly affect the specific values of mangrove habitat for fish (Ellis and Bell
539 2004). Fish that use mangroves must respond to these dynamics with strategies for coping
540 with temporal variation in habitat suitability, including inter-habitat migrations (Reis-Filho
541 and others 2016; Krumme 2009, Sheaves 2005). In tidally-influenced areas, mangroves
542 should be predominantly inhabited by taxa that have developed adaptations to these
543 challenges (Castellanos-Galindo and Krumme 2015). Under these conditions, the
544 surrounding seascape, particularly the kind of habitat that is available at low tide, will
545 strongly influence the set of fish that are able to use mangrove habitat (Pittman and others
546 2007). Some of the variability observed in the present study could relate to qualities of the
547 surrounding seascape that were not examined, such as distance to subtidal habitats other
548 than reef, e.g. seagrass (Gilby and others 2016), or level of predation risk across different
549 habitats across the intertidal zone (Grol and others 2011; Kimirei and others 2013; Kimirei
550 and others 2015). As the results of the present study suggest, tidal amplitude and seascape

551 structure are likely to interact to determine the use of intertidal habitat by fish. Overall,
552 regional differences in mangrove fish communities across the Indo-Pacific are likely due to
553 the pervasive, ecosystem-wide effects of differences in climate, geomorphology and tidal
554 regime.

555

556 While there is certainly taxonomic variation in fauna-habitat associations among regions
557 due to biogeographic history, a large proportion may be predictable based on
558 environmental context. The presence or absence of entire family level taxa can illustrate
559 only broad differences in the use of mangroves, and is likely to conceal important species-
560 level differences among locations. That such large differences in habitat use are apparent,
561 and are partly explainable by environmental context, highlights the primacy of these drivers
562 for coastal ecosystem function. There were consistencies in mangrove use in similar
563 environmental contexts despite geographic separation, suggesting that divergent, context-
564 specific notions of habitat function are valid and necessary. Habitat associations underpin
565 our understanding of the requirements of fauna, and inform the way we manage natural
566 ecosystems. Due to the complex interplay of factors found in this study, it is important to
567 consider the breadth of factors that define environmental context together, in order to
568 understand habitat function. If the context-dependence of a habitat is properly understood,
569 notions of its function for animals may be properly informed, enhancing our ability to make
570 robust environmental decisions.

571

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574

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585

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799 Table 1. Study location details, including environmental parameters and associated meta-
800 data. Biogeographic region follows the biogeographic provinces in Bowen and others (2016).
801 Climate data from www.climate-data.org, accessed October 2018. Tidal pattern data from
802 data.shom.fr, accessed October 2018.

Location	Biogeographic region	Max. tidal range	Tidal pattern	Sampling period	Yearly average rainfall, Climate	Salinity: range sampled	Distance to reef: range sampled	Coordinates
Polynesia	East Indo-Pacific	0.5 m	Semi-diurnal with diurnal inequality	October 2017	1800 mm, Tropical monsoon climate	32-34	30 m - 360 m	17° S 149° W
New Britain	Central Indo-Pacific	0.9 m	Mixed	Sept. – Nov. 2015 & 2016	3700 mm, Tropical monsoon climate	0-35	30 m - 14 km	4° S 151° E
North eastern Aust.	Central Indo-Pacific	3.9 m	Semi-diurnal with diurnal inequality	June – Dec. 2014 & 2015	2237 mm, Tropical monsoon climate	0-35	30 m - 25 km	18° S 146° E
North central Aust.	Central Indo-Pacific	7.8 m	Semi-diurnal with diurnal inequality	September 2015	1694 mm, Tropical savanna climate	29-35	1 km - 27 km	12° S 130° E
North western Aust.	Central Indo-Pacific	10.8 m	Semi-diurnal	June 2017	562 mm, Tropical semi-arid climate	34-35	230 m - 8 km	16° S 123° E

803

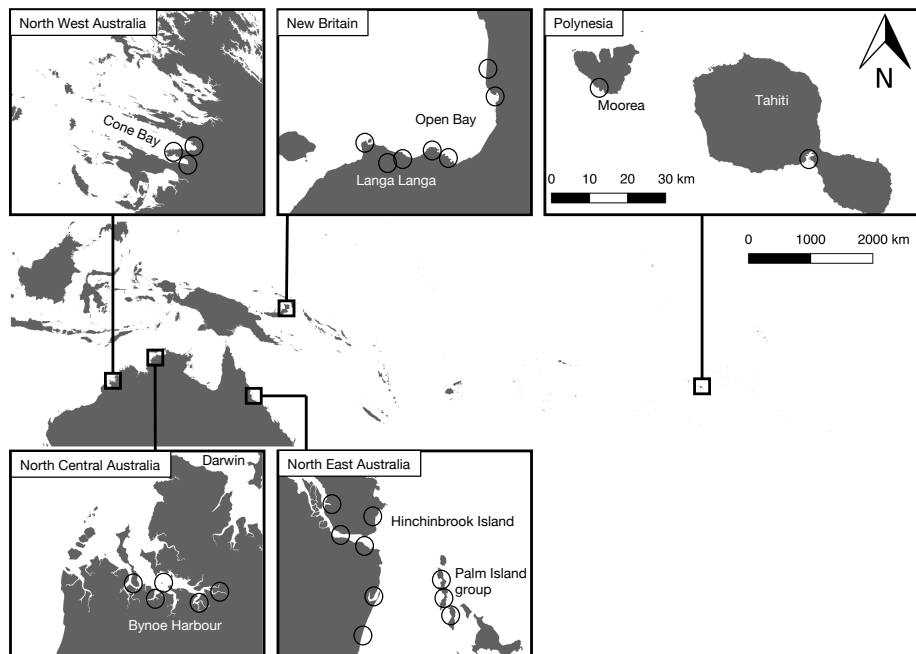
804 Table 2. Variable importance for the three contextual variables used in multivariate
805 regression tree analysis.

Predictor	Rank	Reduction in MSE	Proportional reduction in MSE relative to best predictor
Tidal amplitude	1	0.56	1
Salinity	2	0.37	0.66
Distance to reef	3	0.35	0.63
Biogeographic region	4	0.21	0.36

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807 **Figures**

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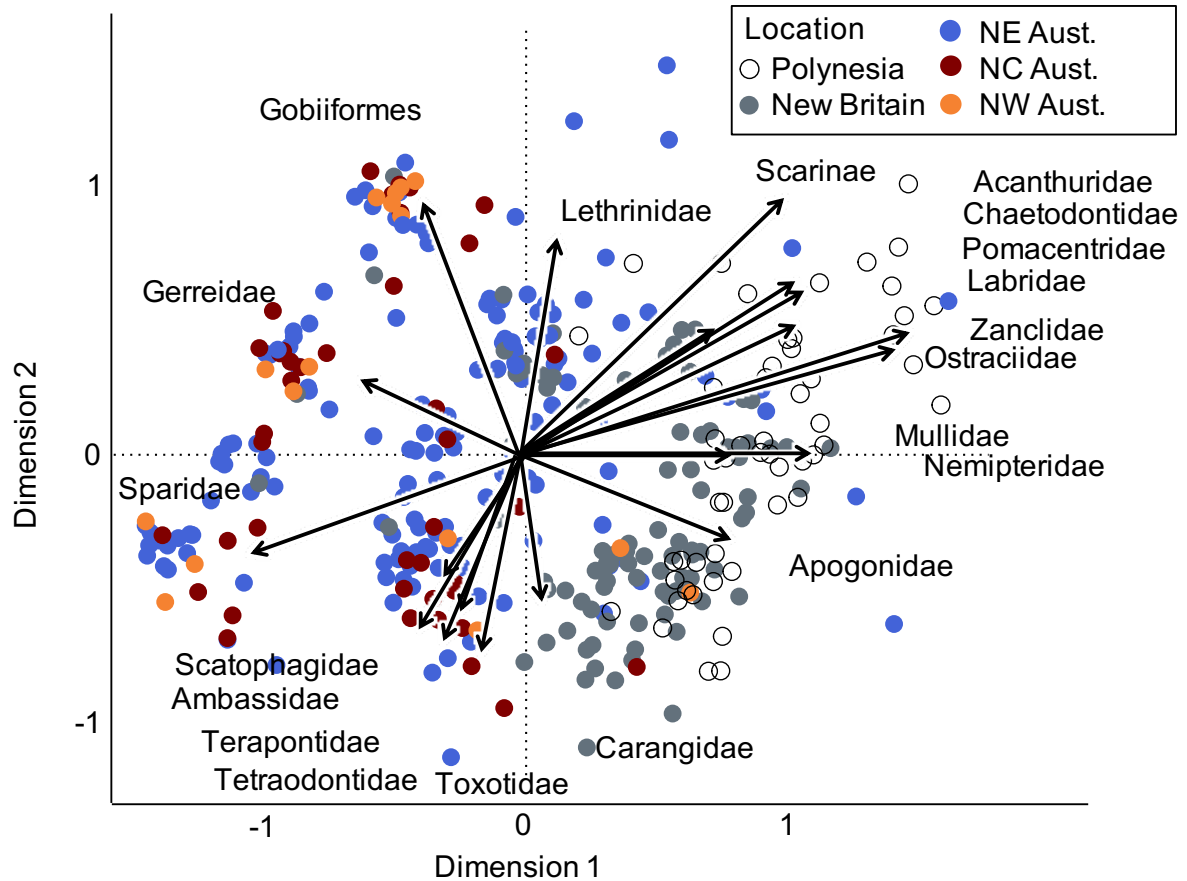
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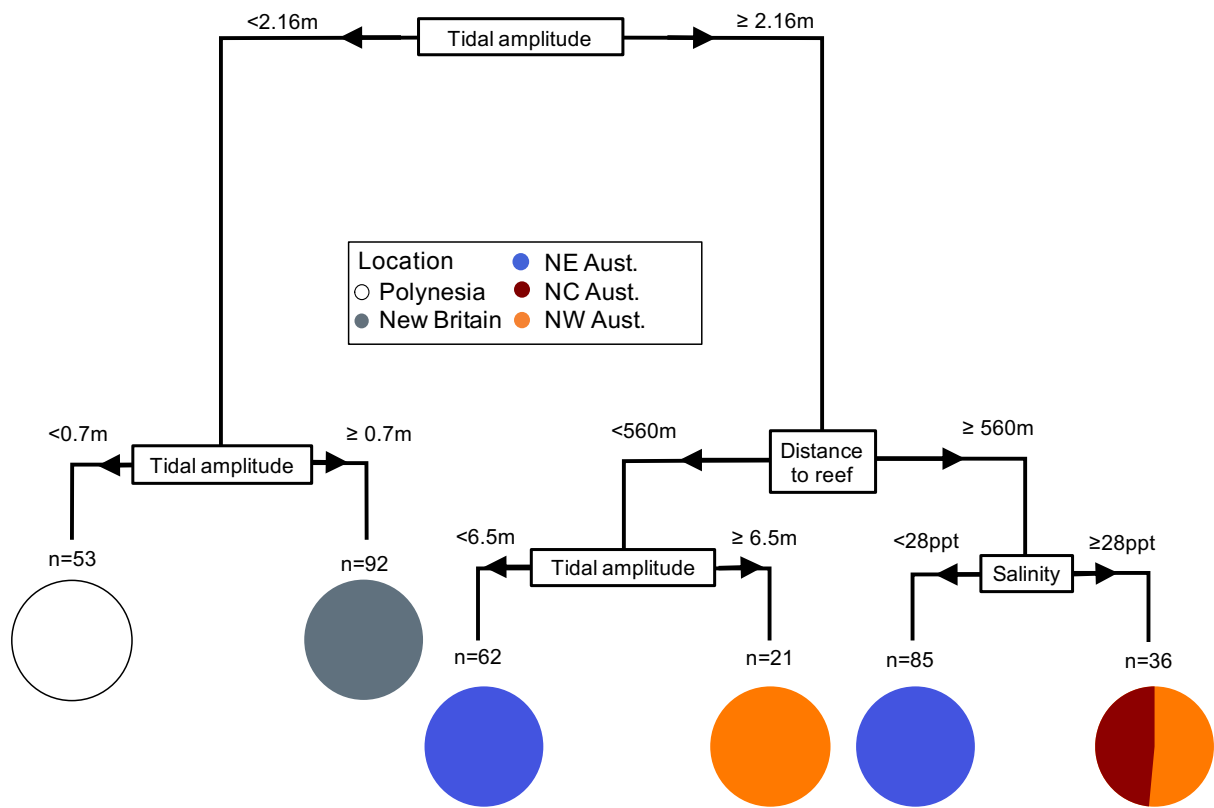
Figure 1. Distribution of sampling effort. Central map shows the Central and Eastern Indo-Pacific, and the location of our study locations. Each study region map is identically scaled (scale bar located in French Polynesia map). Black circles show video sampling sites.



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815 Figure 2. nMDS ordination displaying the first two dimensions of a 3D solution (stress:
 816 0.152) capturing the differences between sites (individual data points) based on an inclusive
 817 data treatment (excluding outliers) of family level taxonomic assemblage composition ($n =$
 818 349). Dimensional values are scaled such that a distance of one unit represents a halving of
 819 assemblage similarity between sites. Taxonomic vectors represent the direction of positive
 820 correlation with the ordination space. Vector terminal position represents a taxa's centre of
 821 occurrence in the ordination space, calculated using the weighted average of site
 822 abundances. Only taxa that are far (>0.5 dimensional units) from the centre of the
 823 ordination space (i.e. taxa that differ strongly across the ordination space) are shown. Points
 824 have been jittered (at a scale of 0.1 on each dimensional axis) to reveal quantities of points
 825 occurring at identical positions in the ordination.

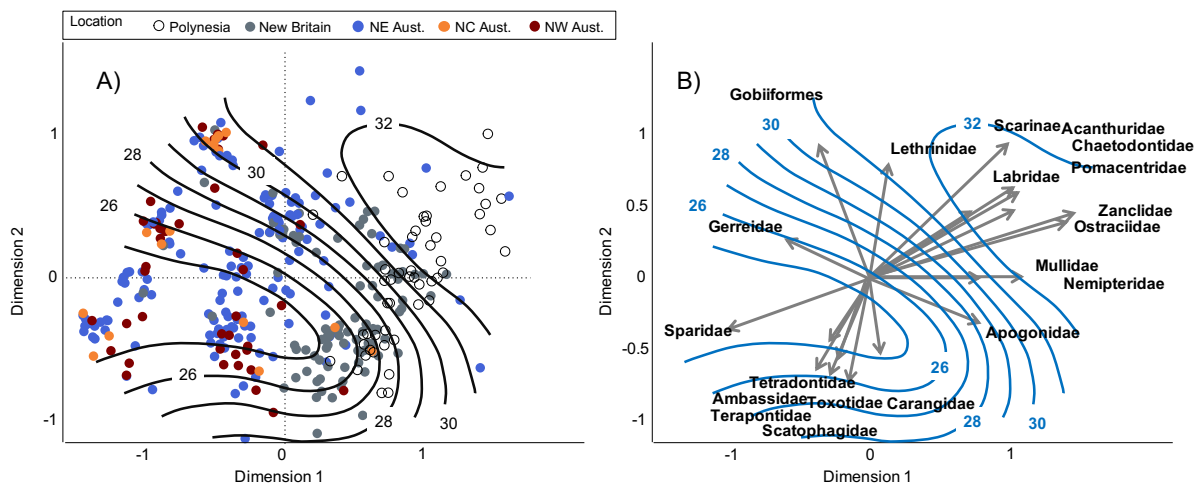
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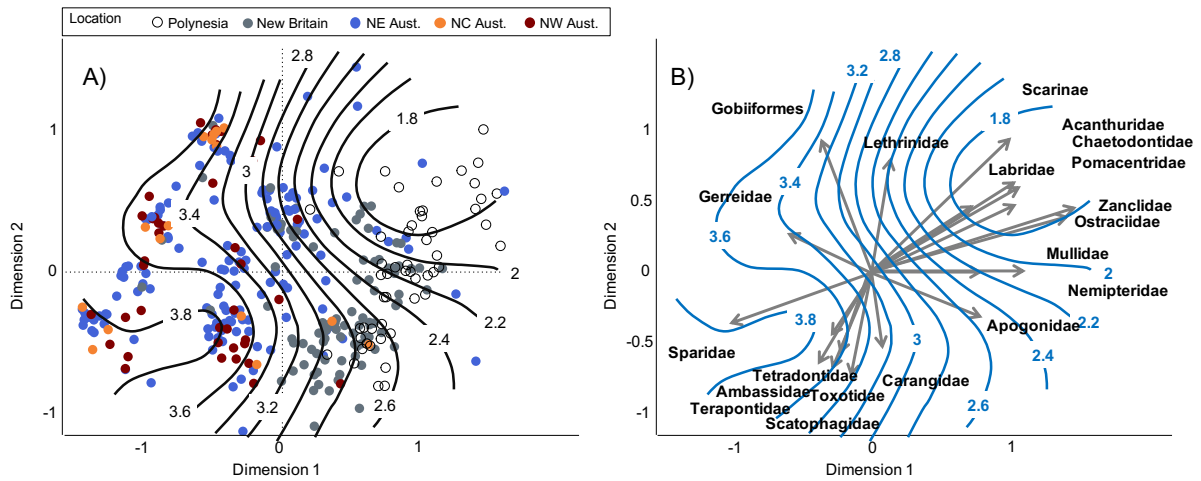
828 Figure 3. Multivariate regression tree describing the major divisions in fish assemblage
 829 composition of mangrove habitat based on environmental context, throughout the Indo-
 830 Pacific (n = 349 sites). Each division is labelled by the contextual factor that is best able to
 831 split the data into homogenous groups, and the values that best distinguish them. The
 832 difference between group community structure is proportional to vertical distance in the
 833 dendrogram. Circles at each final node show the regional composition of samples in each
 834 node – only the far right node contains a mix of samples from different regions.

835

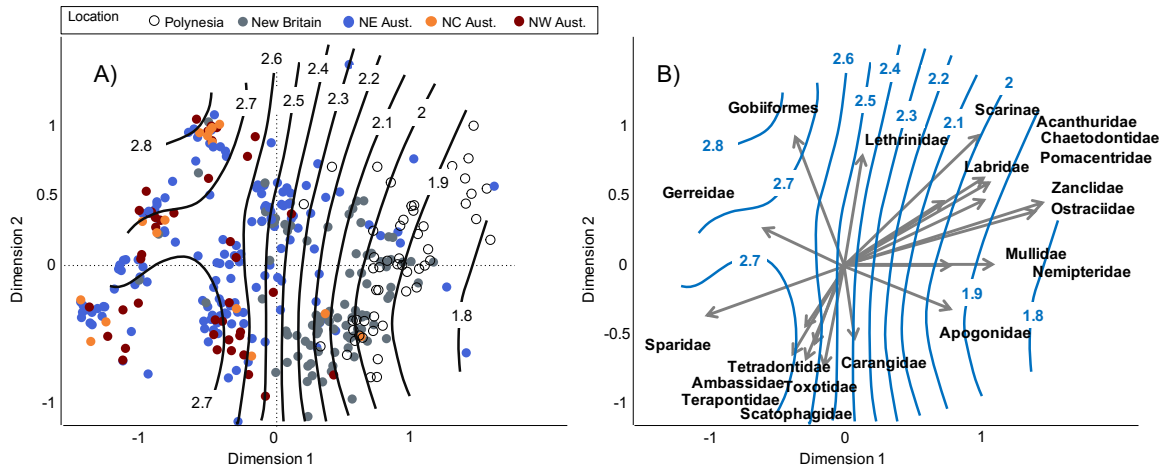


836 Figure 4. Surface fitting of tidal amplitude as an explanatory variable, depicting how
 837 modelled gradients in tidal amplitude relate to fish families and locations. Panel A shows
 838 surface with site points for reference, Panel B shows surface with species vectors for
 839

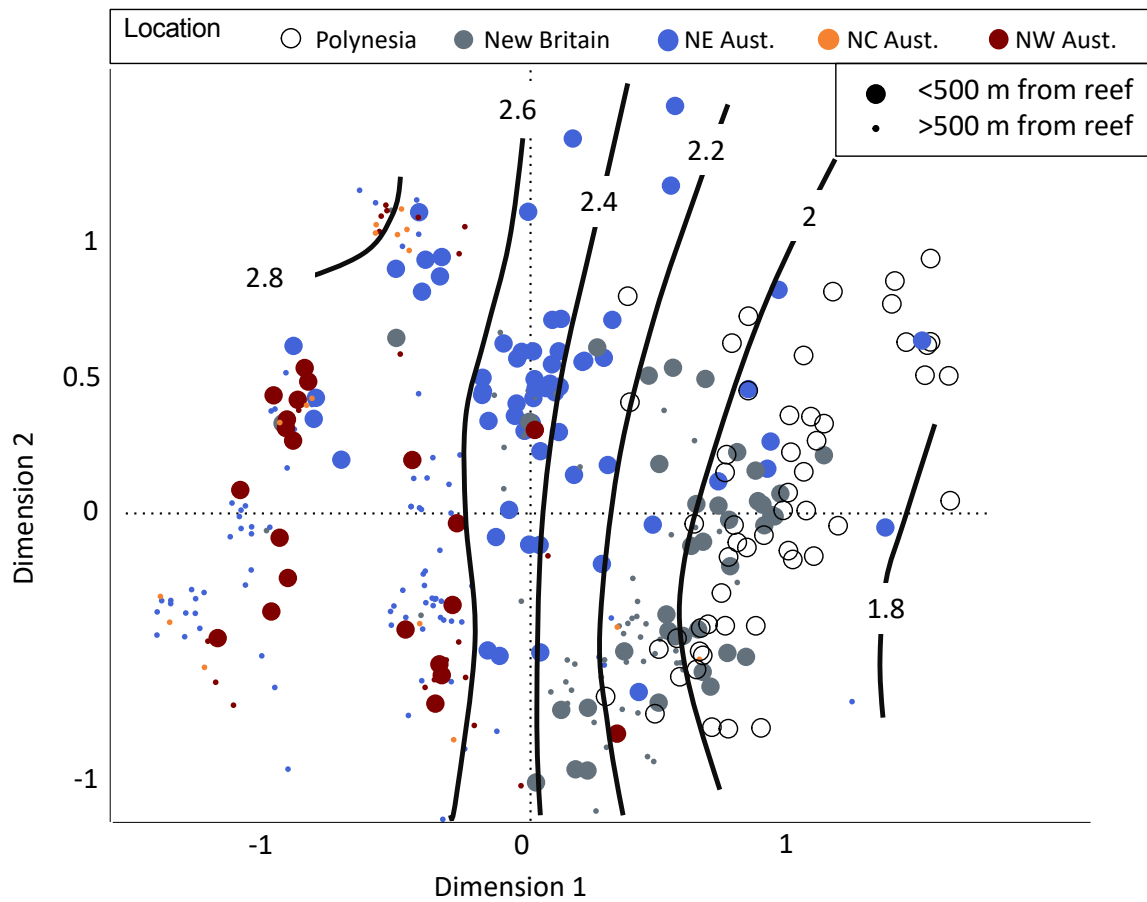
840 reference. Lines describe the topography of the gradient, and units are log (cm). Deviance
 841 explained: 60.9%, indicating to what extent changes in assemblage composition across the
 842 space mirror changes in the values of tidal amplitude.
 843



844
 845 Figure 5. Surface fitting of distance to reef as an explanatory variable, depicting how
 846 modelled gradients in distance to reef relate to fish families and locations. Panel A shows
 847 surface with site points for reference, Panel B shows surface with species vectors for
 848 reference. Lines describe the topography of the gradient, and units are in log (metres).
 849 Deviance explained: 46.7%, indicating to what extent changes in assemblage composition
 850 across the space mirror changes in the values of tidal amplitude.
 851



852
 853 Figure 6. Surface fitting of salinity as an explanatory variable, depicting how modelled
 854 gradients in salinity relate to fish families and locations. Panel A shows surface with site
 855 points for reference, Panel B shows surface with species vectors for reference. Lines
 856 describe the topography of the gradient, and units are in psu. Deviance explained: 14.5%,
 857 indicating to what extent changes in assemblage composition across the space mirror
 858 changes in the values of tidal amplitude.
 859



860
861

862 Figure 7. nMDS ordination showing the relationship between family-level taxonomic
863 assemblage and tidal amplitude. The ordination is a revisualisation of the ordination
864 displayed in Figures 2-6, where sites close to reefs (<500 m) are visible and sites far from
865 reefs (>500 m) have been reduced in size. Surface fitting of tidal amplitude as an
866 explanatory variable is displayed as in Figure 4, with reduced contours for visual clarity.
867 Lines describe the topography of the gradient, and units are log-transformed (cm).