



Comparing Genomic Signatures of Selection Between the Abbassa Strain and Eight Wild Populations of Nile Tilapia (*Oreochromis niloticus*) in Egypt

Maria G. Nayfa^{1,2*}, David B. Jones^{1,2*}, John A. H. Benzie^{3,4}, Dean R. Jerry^{1,2,5} and Kyall R. Zenger^{1,2}

¹ Centre for Sustainable Tropical Fisheries and Aquaculture, College of Science and Engineering, James Cook University, Townsville, QLD, Australia, ² Centre for Tropical Bioinformatics and Molecular Biology, College of Science and Engineering, James Cook University, Townsville, QLD, Australia, ³ WorldFish, Penang, Malaysia, ⁴ School of Biological, Earth and Environmental Sciences, University College Cork, Cork, Ireland, ⁵ Tropical Futures Institute, James Cook University, Singapore, Singapore

OPEN ACCESS

Edited by: Lior David, Hebrew University of Jerusalem, Israel

Reviewed by:

Charles Masembe, Makerere University, Uganda Muniyandi Nagarajan, Central University of Kerala, India

*Correspondence:

Maria G. Nayfa mnayfa12@gmail.com David B. Jones david.jones051986@gmail.com

Specialty section:

This article was submitted to Evolutionary and Population Genetics, a section of the journal Frontiers in Genetics

> Received: 31 May 2020 Accepted: 31 August 2020 Published: 15 October 2020

Citation:

Nayfa MG, Jones DB, Benzie JAH, Jerry DR and Zenger KR (2020) Comparing Genomic Signatures of Selection Between the Abbassa Strain and Eight Wild Populations of Nile Tilapia (Oreochromis niloticus) in Egypt. Front. Genet. 11:567969. doi: 10.3389/fgene.2020.567969 Domestication to captive rearing conditions, along with targeted selective breeding have genetic consequences that vary from those in wild environments. Nile tilapia (Oreochromis niloticus) is one of the most translocated and farmed aquaculture species globally, farmed throughout Asia, North and South America, and its African native range. In Egypt, a breeding program established the Abbassa Strain of Nile tilapia (AS) in 2002 based on local broodstock sourced from the Nile River. The AS has been intensively selected for growth and has gone through genetic bottlenecks which have likely shifted levels and composition of genetic diversity within the strain. Consequently, there are questions on the possible genetic impact AS escapees may have on endemic populations of Nile tilapia. However, to date there have been no genetic studies comparing genetic changes in the domesticated AS to local wild populations. This study used 9,827 genome-wide SNPs to investigate population genetic structure and signatures of selection in the AS (generations 9-11) and eight wild Nile tilapia populations from Egypt. SNP analyses identified two major genetic clusters (captive and wild populations), with wild populations showing evidence of isolation-by-distance among the Nile Delta and upstream riverine populations. Between genetic clusters, approximately 6.9% of SNPs were identified as outliers with outliers identified on all 22 O. niloticus chromosomes. A lack of localized outlier clustering on the genome suggests that no genes of major effect were presently detected. The AS has retained high levels of genetic diversity (H_o _{All} = 0.21 \pm 0.01; H_e _{All} = 0.23 \pm 0.01) when compared to wild populations (H_{o All} = 0.18 \pm 0.01; H_{e All} = 0.17 \pm 0.01) after 11 years of domestication and selective breeding. Additionally, 565 SNPs were unique within the AS line. While these private SNPs may be due to domestication signals or founder effects, it is suspected that introgression with blue tilapia (Oreochromis aureus) has occurred. This study highlights the importance of understanding the effects of domestication in addition

1

to wild population structure to inform future management and dissemination decisions. Furthermore, by conducting a baseline genetic study of wild populations prior to the dissemination of a domestic line, the effects of aquaculture on these populations can be monitored over time.

Keywords: domestication, natural population, population genetics, population structure, outlier analysis, selection, farm management, aquaculture

INTRODUCTION

As aquaculture production increases, so does the number of species undergoing domestication (currently estimated at 598 species; FAO, 2018), where domestication is defined here as the adaptation of an organism from the wild to a captive environment (Price, 1984). These adaptations can be a combination of genetic changes that occur over generations through selective breeding for desirable traits (Argue et al., 2002; Hossain et al., 2011; Moss et al., 2012), but also include adjustments to a captive environment such as reduced antipredator behaviors and aggression (Johnsson et al., 1996; Robinson and Hayes, 2008).

The four main genetic processes that affect animals during domestication are founder effects, selection, genetic drift, and inbreeding (Ladizinsky, 1985; Clutton—Brock, 1992; Ollivier, 2002; Mignon-Grasteau et al., 2005; Andueza-Noh et al., 2015); however, the extent of their effects on the genome can vary. In general, the consequences of inbreeding and genetic drift are widespread and can be observed throughout the genome, whereas selection tends to act differentially across the genome depending on the genetic architecture of the trait (Burke et al., 2005). These micro-evolutionary processes need to be taken into consideration when trying to identify how an organism's genome is being affected by domestication.

One way to understand the genetic consequences of domestication and to identify signatures of selection is to compare population genetic metrics between captive and wild populations (Simmons et al., 2006; López et al., 2019). Recent advances in high-throughput whole genome sequencing has enabled the cost-effective development of genome-wide markers for many non-model species. Such technological developments have enabled researchers to not only harness increased power in identifying the extent to which genetic processes like selection, genetic drift, and inbreeding affect a genome, but also identify specific regions of the genome that have responded to such processes (Carter et al., 2008; Scandura et al., 2011; López et al., 2019). Therefore, evaluating the genetic differences between wild and domestic populations can also help identify genomic regions associated with domestication and desirable market traits, wild populations that exhibit these traits, and local adaptations in wild populations. Additionally, these differences can be used to detect escapees and help estimate their potential impact on local populations.

In 2002, the Abbassa Strain (AS) of Nile tilapia (*Oreochromis niloticus*) was initiated by the WorldFish Center in an effort to increase aquaculture production of this species in Egypt (Rezk et al., 2009; Ibrahim et al., 2013). Its purpose was to provide a genetically diverse population based on the

local strain of Nile tilapia that could be selectively improved for growth. Subsequently, the AS was established from four Egyptian populations (three wild: Zawia, Abbassa, and Aswan; one hatchery: Maryout). The production of AS is currently restricted to the Nile Delta; however, WorldFish and the Egyptian government plan to disseminate the AS line throughout Egypt.

To date, genetic diversity studies have found that wild Nile tilapia populations have evidence for sub-structuring in Egypt, particularly between populations in the Nile Delta in Upper Egypt compared with populations in the Lower Egyptian portion of the Nile River (Hassanien et al., 2004; Hassanien and Gilbey, 2005). However, due to the age of the studies, possible translocations and the availability of improved genetic technologies, updated investigations into the genetic structure of these populations using high density, genome-wide markers are warranted to determine the current status of wild population genetic structuring.

This study investigated the population genetic structure, evidence for signatures of selection, and genetic diversity related to domestication in the AS compared to wild Egyptian Nile River *O. niloticus* populations. This information can then be used to understand the impact disseminating the AS may have on wild stocks, as well as understand if targeted breeding in the AS has resulted in signatures that may be indicative of domestication.

MATERIALS AND METHODS

Sampling and DNA Extraction Wild Population Sampling

Fin clips from 400 Nile tilapia were collected from eight wild populations (Aswan, n = 50; Manzala Lagoon, n = 50; Kanata, n = 50; Lake Idku, n = 50; Damietta, n = 50; Lake Burullus, n = 50; Rosetta, n = 50; and Asyut, n = 50) along the Nile River, Egypt. Of these, Aswan was one of the four sites from which individuals were sampled to create the domesticated Abbassa Strain in 2002. Samples were obtained directly from commercial fishing boats, with fish for an individual location obtained over a distance of approximately 1 to 175 km. Samples were preserved in 70% ethanol and submitted to Diversity Arrays Technology (DArT) in Canberra, Australia, for DNA extraction and high throughput genotyping by sequencing using proprietary DArTseqTM technology¹. To obtain purified DNA, extractions were conducted using commercially available extraction kits (Promega, Qiagen; Lind et al., 2017).

¹https://www.diversityarrays.com

Abbassa Strain Population Sampling

Fin clips from 483 samples were collected from the three most recent generations of the AS at the time of this study [121 individuals from generation 9 (G9); 216 individuals from generation 10 (G10); and 146 individuals from generation 11 (G11)]. DNA extractions and genotyping were conducted by Diversity Arrays Technology (DArT) as described in Lind et al. (2017).

Library Preparation and Sequencing

DArTseqTM uses a combination of complexity reduction methods, which were originally optimized on the Jaccoud et al. (2001) microarray platform. These methods effectively select low copy sequences from a genome before sequencing them on next generation sequencing platforms (Kilian et al., 2012; Courtois et al., 2013; Von Mark et al., 2013; Raman et al., 2014; Lind et al., 2017). As this process uses both a rare and a more frequently cutting enzyme, it is similar to double digest RAD sequencing (ddRAD; Peterson et al., 2012; Lind et al., 2017).

DArTseqTM reduced-representation libraries were prepared as described by Sansaloni et al. (2011) and Kilian et al. (2012). In short, optimization of the complexity reduction process for Nile tilapia was achieved by using a combination of PstI and HpaII methylation-sensitive restriction enzymes for digestion and unique barcode sequences ligated onto the ends of each resulting fragment (Kilian et al., 2012; Lind et al., 2017; Kjeldsen et al., 2019). Bridge amplification was achieved by incorporating a PstI specific adaptor with an Illumina flow-cell attachment region, primer sequence, and unique barcode coupled with the reverse HpaII specific adaptor containing a second Illumina flowcell attachment sequence (Lind et al., 2017; Schultz et al., 2018; Kjeldsen et al., 2019). Therefore, only fragments containing both PstI and HpaII cut sites were amplified for sequencing. To ensure complete digestion and a uniform range of fragment sizes, all samples were checked using an agarose gel. Any samples which displayed downshifted bands after digestion during DArTseq library preparation were removed. These downshifted samples exhibited a lower amplicon range than expected when compared to other samples and are not ideal for a consistent genotype assay. A total of eight downshifted samples were not included within the sequencing effort. Additionally, a minimum of 15% random technical replicates were included in all genotyping batches for quality control.

Quality Control and Initial SNP Calling

DArT's proprietary marker calling algorithm DArTsoft14 was used to call SNPs (Lind et al., 2017), implemented in the KDCompute framework². Samples from wild locations were then co-analyzed by DArT alongside 483 samples from three generations of the AS, which had already been processed using DArTseqTM technology as part of a previous experiment (Nayfa et al., 2020).

A total of 19,505 SNP markers were identified across all 875 samples and were filtered using a custom Python script adapted

from DartQC³ and CD-HIT-EST (Li and Godzik, 2006). Briefly, samples with greater than 50% missing data were removed from the dataset and individual genotypes calls made with fewer than five reads were silenced. Genotypes with a count comparison, or the comparison of read counts between REF and SNP alleles, were silenced if they fell between 0.05 and 0.1, where <0.05 is considered to be homozygous and >0.1is considered to be heterozygous (see text footnote 3). SNPs were then filtered if they had an average replication statistic of less than 90%, a call rate less than 50%, and a minor allele frequency (MAF) of less than 1% in at least one population. The clone ID sequences from which SNPs were called and clustered together at 95% similarity using CD-HIT-EST (Li and Godzik, 2006). Within each cluster, the SNP with the highest MAF was retained to ensure a more even representation of the genome. A total of 9,827 high quality SNPs and 821 samples (90.9% of collected samples) were retained for all downstream analyses.

Population Genetic Structure Broad Scale Population Structure

To determine broad-scale population differentiation across the eight wild locations and three generations of the AS, two separate clustering models (the allele frequencies correlated model and the allele frequencies independent model) were utilized within a Bayesian cluster population structure analysis in STRUCTURE 2.3.4 (Pritchard et al., 2000; Falush et al., 2003, 2007; Hubisz et al., 2009). In order to avoid inappropriate clustering due to K being set too small, K was set from 1 to 12, so that the maximum clustering possible was larger than the number of putative populations (Kalinowski, 2011). Three repeat runs were performed for each K (1–12), with a burn-in period of 5,000 iterations followed by 50,000 final iterations using the admixture model and no prior probabilities for cluster membership. Both clustering models yielded near identical results. The optimal number of population clusters, K, was determined using an ad hoc statistic Delta K (ΔK). ΔK is the degree of change in the log probability of data between successive K values, and was calculated using Structure Harvester (Evanno et al., 2005; Earl and vonHoldt, 2012). To ensure that any structuring observed in the wild populations was not biased by the inclusion of individuals from a domesticated line, analyses with the same parameters were repeated on only the eight native sampling locations testing a *K* of 1 to 9.

Fine Scale Population Structure

Fine-scale population genetic structuring across all eight wild sampling locations and the three AS generations was assessed using pairwise relationships based on identityby-state (IBS) distance calculated in Plink v.1.9 (Purcell et al., 2007; Purcell, 2020). These relationships were then visualized using mutual k-nearest neighbor graphs in the NETVIEW pipeline v.1.1 at kNN values between 1 and 100 (Neuditschko et al., 2012; Steinig et al., 2016). To confirm

²http://www.kddart.org/kdcompute.html

³https://github.com/esteinig/dartqc

fine-scale population genetic structuring a principal coordinates analysis (PCoA) was also conducted using GenAlEx v. 6.51b2 (Peakall and Smouse, 2006, 2012). Genetic distance per population was first calculated in GenAlEx v. 6.51b2 and then a covariance-standardized PCoA method was applied (Peakall and Smouse, 2006, 2012). To identify the percentage of genetic variation that can be attributed to differences between and within populations, an analysis of molecular variance (AMOVA) based on genetic distance was then conducted using 9,999 permutations in GenAlEx v. 6.51b2 (Peakall and Smouse, 2006, 2012).

To test if any identified genetic structuring followed an isolation-by-distance model of population divergence, Mantel's test for correlation between genetic (F_{st}) and genetic distance (km) was conducted in the R package *adegenet* using 10,000 permutations in the *mantel.randtest()* function (Jombart, 2008; Jombart and Ahmed, 2011). Genetic distance was calculated using an Euclidean method based on Angular distance in the *adegenet* function *dist.genpop()* (Jombart, 2008; Jombart and Ahmed, 2011). Geographic distances were calculated based on the shortest distance between two points according to the "Vicenty (ellipsoid)" method calculated using the R package *geosphere* (Hijmans et al., 2017).

Signatures of Selection

Population Outlier Analysis

To identify outliers (including loci which are being influenced by selective processes), two independent software were utilized: Arlequin 3.5.2.2 (Excoffier and Lischer, 2010) and BayeScan 2.1 (Foll and Gaggiotti, 2008; Foll, 2012). For comparisons between the two groups (wild Nile tilapia and the domesticated AS Nile tilapia), only those candidate outliers that were jointly identified between programs were categorized as putative outliers. Outlier analyses within Arlequin 3.5.2.2 were based on a hierarchical island model with 20,000 simulations, 50 simulated groups, and 100 demes simulated per group (Excoffier and Lischer, 2010). AMOVA computations were conducted using a pairwise difference method with no Gamma correction (Excoffier and Lischer, 2010).

Outlier analyses within BayeScan 2.1 were based on a neutral model with 1:10 prior odds, 20 pilot runs consisting of 5,000 iterations each, followed by 100,000 iterations with a burn-in-length of 50,000 iterations as recommended by Foll (2012). To establish whether a neutral or selection model was in effect for each SNP the ratio of posterior probabilities, Bayes factors (BF) were calculated. A Jeffrey's interpretation of "strong" BF (*p*-value ≤ 0.05) to "decisive" BF (*p*-value- ≤ 0.01) was then utilized to identify outliers and ascertain which model the posterior odds favored (Foll, 2012). For markers which fell under a selection model, positive alpha values were then used to identify markers that were under diversifying or directional selection, whereas negative alpha values were used to identify those markers under background, or balancing, selection (Foll, 2012). For pairwise comparisons of populations, only BayeScan as (1) the hierarchical method utilized in Arlequin required the use of multiple populations per grouping,

this analysis did not support individual pairwise population comparisons and (2) the majority of outliers in genetic clusters identified by BayeScan were also identified by Arlequin 3.5.2.2 (approximately 70%).

To test for the normality of markers, quantile-quantile plots (QQ-plots) with a 95% confidence interval were constructed in the R package GWASTools v. 3.1 (Gogarten et al., 2012) for the full marker set, as well as the neutral marker sets (Gondro et al., 2013; Hayes, 2013). To validate the outlier selection criteria selected (i.e., markers jointly identified by both BayeScan and Arlequin), QQ-plots using the two different neutral marker sets (one with all identified outliers removed and one with only jointly identified outliers removed) were created. How well the data fitted the assumption of normality was then compared between both datasets, and only those jointly identified moved forward. Comparison of these datasets allowed the validity of identified outliers to be established.

Genomic Regions Under Selection

Raw clone sequences from which SNPs were identified during the DArTseq process were annotated to the available genome assembly for *Oreochromis niloticus* (GenBank Assembly Accession: GCA_00188235.2; Orenil1.1) using a custom Perl script based on NCBI CGI BLAST interface with a 70% minimum sequence identity (Heller-Uszynska et al., 2011; **Supplementary Material 1**). The Orenil1.1 genome assembly was used instead of the more recent O_niloticus_UMD_NMBU assembly as it was in greater agreement with the linkage maps created for the Abbassa Strain (unpublished data).

Genetic Diversity Statistics

To determine the genetic diversity available within each sampled population (wild and AS), observed (H_o) and expected (H_e) heterozygosity in addition to the number of polymorphic markers within a population were calculated in ARLEQUIN 3.5.2.2 (Excoffier and Lischer, 2010). Heterozygosity and the number of polymorphic markers were examined across scenarios with different amounts of missing data (all markers; 5% missing allowed per SNP within individual populations; and 50%, 25%, and 5% missing allowed per SNP across the entire dataset). Additionally, average multilocus heterozygosity (MLH) for each population was computed using the R package inbreedR (Stoffel et al., 2016). Private SNPs per population were calculated using the R package PopGenKit v.1.0 (Paquette and Paquette, 2011). To determine the level of differentiation amongst populations, pairwise and global Fst values were calculated in ARLEQUIN 3.5.2.2 (Excoffier and Lischer, 2010). Levels of inbreeding per sampling location and time point were examined using the inbreeding coefficient (F_{is}) calculated in ARLEQUIN 3.5 using 1,000 permutations (Excoffier and Lischer, 2010). Hardy-Weinberg equilibrium (HWE) was calculated in ARLEQIUN 3.5 using 1,000,000 Markov chain steps and 100,000 dememorization steps (Excoffier and Lischer, 2010; Waples, 2014). Effective population size in each native location was calculated using the linkage disequilibrium method (LDNe) in NeEstimator V2.01 (Do et al., 2014).

RESULTS

Population Genetic Structure Broad Scale Population Structure

The ad hoc ΔK statistic indicated evidence for two major genetic clusters within the dataset (Supplementary Material 2). This distinction was supported by STRUCTURE admixture analysis whereby the domesticated AS generations formed one genetic cluster and the eight wild populations comprised the second cluster (Figure 1 and Supplementary Material 2). With a K of two, the admixture model used in STRUCTURE assumes that each individual has ancestry from only one or both of these genetically distinct clusters (Lawson et al., 2018). Given this, every individual from the AS shares genetic material with the wild Nile tilapia. This is reflected in the minimal population structuring identified between AS and wild sampling locations identified by pairwise F_{st} values ($F_{st} = -0.008 - 0.058$; Supplementary Material 3). The largest genetic distance was observed between the two most southern wild sampling locations (Asyut and Aswan) and the AS ($F_{st} = 0.045 - 0.058$; Supplementary Material 3).

Within the wild sampling locations, one individual from Rosetta was more closely related to the AS than to the wild genetic cluster (**Figure 1**). There are two individuals from Damietta, two from Kanater, and one from Aswan which also had a higher proportion of shared ancestry with the AS than expected based on the other individuals in the wild genetic cluster (**Figure 1**).

When the eight wild locations were examined separately, the ΔK statistic identified a total of four weakly separated genetic clusters (**Figure 2** and **Supplementary Material 4**). While each sampled location showed evidence of all four genetic clusters within them, the proportion of these genetic clusters changed along the northern to southern gradient of the Nile River. The two most southern populations (Asyut and Aswan) exhibited the greatest difference in admixture ratios compared to Lake Idku, Rosetta, Lake Burullus, Damietta, and Manzala Lagoon (**Figure 2**). Kanater displayed the largest shift between the northern and southern sampling locations (**Figure 2**). This was supported by pairwise F_{st} values which revealed no subpopulation structuring amongst the wild populations. The greatest F_{st} was between the northernmost population (Lake Burullus) and the southernmost population (Aswan; $F_{st} = 0.021$; **Supplementary Material 3**).

Individuals showing an independent genetic cluster (green) in **Figure 2** were the same as individuals which displayed a greater association with the AS in **Figure 1**. This pattern suggests these individuals are possibly escapees (Rosetta) or subsequent offspring (Kanater, Damietta, and Aswan) of the AS.

Fine-Scale Population Structuring

Mutual k-nearest neighbor analyses conducted in NetView pipeline v.1.1 to determine fine-scale population structuring exhibited a similar pattern to the STRUCTURE admixture analysis. The three generations of the AS formed a distinct genetic cluster separate from the eight wild sampling locations, whilst the wild populations exhibited evidence of isolation-by-distance (Figure 3B and Supplementary Material 5). The two most southern populations further north and form a smaller, separate cluster (Supplementary Material 5). However, a few individuals from these southern locations intermingled with northern samples indicating gene flow between these populations (Figure 3B). There is a single sample from Damietta which







FIGURE 2 Broad-Scale Population Structure. Structure plot of the eight wild sampling locations along a geographical gradient down the Nile River, Egypt at *K* = 4. Colors (red, yellow, blue, and green) represent the four genetic clusters identified. Vertical bar colors are indicative of admixture of the four identified genetic clusters per individual.



clustered with the AS. Since Damietta is in close geographical proximity to the farm, it is conceivable that this individual is an escapee from the AS program (**Figure 3B**). Additionally, two individuals from the southernmost Aswan population formed a third clustering: indicating, that populations from further south in the Nile River and connecting waterways and lakes may exhibit greater variation amongst populations or include hybridized individuals (**Figure 3B**).

Similar patterns were observed in the PCoA analysis of population, the three AS generations formed an independent cluster along the first coordinate axis which accounted for 45.3% of the variation among clusters (Figure 3C). An AMOVA between the three generations of AS and the eight wild populations indicated that 13.8% of the molecular variance observed was between the AS and wild genetic clusters

(p = 0.0001). The Nile Delta populations were more distinct from the upstream Nile populations, Asyut and Awan (**Figure 3C**). Kanater is located both physically and genetically between the five Nile Delta populations and the two southernmost upstream populations (**Figures 3A,B**). An AMOVA of the eight wild populations demonstrated that 10.6% of the molecular variance observed is among populations (p = 0.0001).

Signatures of Selection

The QQ-plots examining the entire marker set revealed that the data violated the assumption of normality, indicating the presence of outliers (**Supplementary Material 6**). A total of 674 outliers were jointly identified by both BayeScan and Arlequin between wild and domestic genetic clusters (**Table 1**). These outliers were confirmed by re-examining normality of the data

TABLE 1	Pairwise outlier	analysis all	directional an	d balancing	outlier loci identified.
		a layoro an	an ootion an an	a bala long	

	Gen 10	Gen 11	Lake Idku	Rosetta	Lake Burullus	Damietta	Manzala Lagoon	Kanater	Asyut	Aswan	Wild
Gen 9	0	11	5	3	11	4	6	6	13	6	
Gen 10		14	6	6	13	3	5	4	11	5	
Gen 11			2	4	13	5	5	4	10	1	
Lake Idku				0	0	0	0	0	1	0	
Rosetta					2	0	0	0	11	2	
Lake Burullus						0	3	1	0	1	
Damietta							1	0	4	0	
Manzala Lagoon								0	3	0	
Kanater									1	1	
Asyut										3	
Domestic											674

Outliers detected between the identified broadscale populations (i.e., Wild and Domestic) are those that were jointly identified by both BayeScan and Arlequin. Outliers detected between pairwise comparisons of sampled locations and AS generations are those detected by BayeScan, the more conservative of the two programs utilized for analysis.

using QQ-plots when the identified outliers were removed. QQplots revealed that the data conformed more to the assumption of normality than previously; however, there were likely still unidentified outliers in the dataset (**Supplementary Material 6**). When all outliers identified by either BayeScan or Arlequin were removed from the dataset, they did not conform to the assumption of normality, indicating that those outliers identified by only one program were unlikely to be true outliers (**Supplementary Material 6**). This confirmed the decision to utilize only jointly identified markers by both BayeScan and Arlequin when multiple sampling sites constituted a population (i.e., domestic or wild genetic clusters).

The greatest number of outliers (674) was found between the two genetic clusters identified using broad-scale population structuring analysis (**Table 1** and **Figure 1**). Of those outliers, 187 had negative alpha values in BayeScan and are under balancing selective forces, whereas the remaining 487 outliers had positive alpha values indicating directional selection. On average, pairwise comparisons of either Asyut or Lake Burullus to domestic populations (G9-11) yielded the greatest number of outliers (10–13; **Table 1**). The five wild populations which are most closely located in the Nile Delta (Rosetta, Lake Burullus, Damietta, Manzala Lagoon, and Kanater) had the fewest identified outliers (zero-three) when compared pairwise amongst themselves (**Table 1**). Regarding the pairwise comparisons of wild populations, Asyut vs. Rosetta had the greatest number of outliers (11) followed by Asyut vs. Damietta (4; **Table 1**).

Outliers accounted for approximately 6.9% of the entire SNP marker set, with balancing outliers accounting for approximately 1.9% of the entire marker set and diversifying outliers accounting for approximately 5.0%. Diversifying outliers accounted for 72.3% of all identified outliers, whereas balancing outliers accounted for 27.7% (**Supplementary Material 7**). Of the 674 identified outliers, 493 mapped back to the Orenil1.1 genome (**Supplementary Material 7**). Every chromosome in *O. niloticus* had both directional and balancing outliers present, with the number of outliers per chromosome ranging from 9 to 61 (**Supplementary Material 7**).

Genetic Diversity

The estimated effective population size for the AS ranged between $N_e = 14.8-48.6$ per generation, with only approximately 20 - 81% of each generation's breeding population genotyped (**Table 4.2**; Nayfa et al., 2020). Estimated effective population sizes of wild populations ranged from 30.5 – infinite, with infinite being indicative of an infinite-sized ideal population and is taken to be an extremely high and positive value (**Table 4.2**; Jones et al., 2016). Despite these variations in effective population size, all F_{is} values were non-significant and negative in all AS generations and wild populations (**Supplementary Material 8**). The proportion of SNPs that deviated from HWE in domestic populations were 2.8 – 14.6 times more frequent than in wild populations (**Supplementary Material 8**).

Overall, the domestic population genetic cluster had higher expected heterozygosity (He), observed heterozygosity (Ho), multilocus heterozygosity (MLH), minor allele frequencies (MAF) and polymorphic loci than the wild genetic cluster when all, neutral, or directional markers were taken into consideration (Table 2). The greatest difference among populations and genetic clusters was observed when directional outlier markers were examined. When AS generations and wild population were individually considered, levels of both Ho and He for all and neutral markers were similar. In most instances, wild sampling locations (except Ho: Rosetta and Damietta and He: Rosetta, Damietta, and Kanater) had higher levels of heterozygosity than individual AS generations (Table 2). Rosetta had the lowest observed heterozygosity ($H_{o_All} = 0.181$, $H_{o_Neutral} = 0.180$, and H_{o_Directional} = 0.154) and expected heterozygosity $(H_{e_All} = 0.212, H_{e_Neutral} = 0.210, and H_{e_Directional} = 0.214)$ in these three marker sets (Table 2).

The domestic populations, considered as a whole genetic cluster and individually, had a higher MLH overall than wild populations across three marker subsets (All, Neutral, and Directional; **Table 2**). Manzala Lagoon had the lowest MLH in all three marker sets (MLH_{All} = 0.145, MLH_{Neutral} = 0.130, and MLH_{Directional} = 0.151; **Table 1**). However, when only balancing outlier markers were analyzed, genetic diversity indices

	Category n N _{eLD}				$He \pm SE$				MLH ± SE					MAF	\pm SE	Polymorphic Loci							
				All	Neutral	Direc- tional	Balan- cing	All	Neutral	Direc- tional	Balan- cing	All	Neutral	Direc- tional	Balan- cing	All	Neutral	Direc- tional	Balan- cing	All	Neutral	Direc- tional	Balan- cing
Gen 9	Domestic	121	34.1– 45.6	0.211 ± 0.013	0.207 ± 0.013	0.271 ± 0.013	0.284 ± 0.014	0.232 ± 0.014	0.226 ± 0.014	0.329 ± 0.012	0.291 ± 0.014	0.182 ± 0.002	0.178 ± 0.002	0.221 ± 0.005	0.271 ± 0.003	0.191 ± 0.006	0.126 ± 0.031	0.249 ± 0.014	0.203 ± 0.011	9,291	8,658	447	186
Gen10	Domestic	204	37.9– 48.6	0.212 ± 0.010	0.206 ± 0.010	0.283 ± 0.010	0.280 ± 0.011	0.231 ± 0.011	0.224 ± 0.011	0.338 ± 0.009	0.290 ± 0.011	0.170 ± 0.002	0.166 ± 0.002	0.214 ± 0.003	0.255 ± 0.003	0.178 ± 0.004	0.127 ± 0.028	0.242 ± 0.014	0.206 ± 0.013	8,671	8,084	406	181
Gen 11	Domestic	145	14.8– 21.9	0.211 ± 0.013	0.205 ± 0.012	0.283 ± 0.012	0.282 ± 0.013	0.229 ± 0.013	0.222 ± 0.013	0.339 ± 0.011	0.289 ± 0.012	0.175 ± 0.001	0.171 ± 0.001	0.222 ± 0.004	0.268 ± 0.003	0.185 ± 0.005	0.126 ± 0.022	0.247 ± 0.011	0.204 ± 0.010	8,934	8,338	414	185
Lake Idku	Wild	49	493.9- Infinite	0.214 ± 0.024	0.213 ± 0.023	0.200 ± 0.023	0.286 ± 0.024	0.232 ± 0.024	0.229 ± 0.024	0.247 ± 0.025	0.284 ± 0.022	0.133 ± 0.002	0.132 ± 0.002	0.102 ± 0.002	0.270 ± 0.005	0.156 ± 0.007	0.129 ± 0.022	0.153 ± 0.030	0.200 ± 0.020	6,404	5,970	251	183
Rosetta	Wild	48	30.5- Infinite	0.181 ± 0.029	0.180 ± 0.029	0.154 ± 0.029	0.274 ± 0.028	0.212 ± 0.029	0.210 ± 0.029	0.214 ± 0.031	0.286 ± 0.027	0.134 ± 0.002	0.133 ± 0.002	0.109 ± 0.002	0.261 ± 0.007	0.168 ± 0.009	0.125 ± 0.022	0.157 ± 0.028	0.199 ± 0.020	7,626	7,076	366	184
Lake Burullus	Wild	50	746.6- Infinite	0.221 ± 0.024	0.220 ± 0.024	0.208 ± 0.024	0.286 ± 0.023	0.236 ± 0.024	0.233 ± 0.024	0.251 ± 0.025	0.287 ± 0.022	0.142 ± 0.001	0.141 ± 0.001	0.110 ± 0.003	0.276 ± 0.004	0.169 ± 0.007	0.127 ± 0.022	0.144 ± 0.029	0.200 ± 0.020	6,754	6,285	285	184
Damietta	Wild	50	590- Infinite	0.202 ± 0.023	0.201 ± 0.023	0.180 ± 0.023	0.272 ± 0.023	0.225 ± 0.024	0.223 ± 0.024	0.226 ± 0.025	0.285 ± 0.022	0.136 ± 0.002	0.135 ± 0.002	0.107 ± 0.003	0.257 ± 0.005	0.166 ± 0.007	0.127 ± 0.024	0.149 ± 0.029	0.203 ± 0.021	7,041	6,551	307	183
Manzala Lagoon	Wild	43	175.4- Infinite	0.222 ± 0.025	0.221 ± 0.025	0.205 ± 0.025	0.271 ± 0.025	0.240 ± 0.025	0.238 ± 0.025	0.252 ± 0.026	0.284 ± 0.023	0.126 ± 0.004	0.125 ± 0.004	0.097 ± 0.004	0.247 ± 0.006	0.149 ± 0.007	0.130 ± 0.022	0.151 ± 0.033	0.200 ± 0.021	5,942	5,521	244	177
Kanater	Wild	50	297.3- Infinite	0.216 ± 0.025	0.214 ± 0.025	0.215 ± 0.027	0.305 ± 0.025	0.221 ± 0.024	0.218 ± 0.024	0.241 ± 0.026	0.290 ± 0.022	0.153 ± 0.006	0.152 ± 0.006	0.130 ± 0.007	0.292 ± 0.007	0.172 ± 0.008	0.152 ± 0.009	0.153 ± 0.028	0.205 ± 0.020	7,627	7,125	316	186
Asyut	Wild	33	149.3- Infinite	0.268 ± 0.042	0.266 ± 0.042	0.270 ± 0.042	0.314 ± 0.044	0.259 ± 0.036	0.257 ± 0.036	0.292 ± 0.037	0.287 ± 0.034	0.164 ± 0.005	0.163 ± 0.005	0.133 ± 0.006	0.301 ± 0.008	0.182 ± 0.013	0.149 ± 0.012	0.147 ± 0.036	0.205 ± 0.026	6,553	6,106	260	187
Aswan	Wild	28	55.7- Infinite	0.247 ± 0.032	0.245 ± 0.042	0.251 ± 0.034	0.290 ± 0.033	0.265 ± 0.031	0.264 ± 0.031	0.288 ± 0.031	0.290 ± 0.029	0.137 ± 0.003	0.136 ± 0.003	0.107 ± 0.004	0.277 ± 0.006	0.169 ± 0.009	0.152 ± 0.012	0.147 ± 0.041	0.203 0.203	5,995	5,581	228	186
Domestic	0	470	470	0.208 ± 0.007	0.203 ± 0.007	0.271 ± 0.006	0.281 ± 0.007	0.228 ± 0.007	0.222 ± 0.007	0.330 ± 0.006	0.289 ± 0.007	0.175 ± 0.001	0.171 ± 0.001	0.218 ± 0.002	0.263 ± 0.002	0.184 ± 0.003	0.151 ± 0.006	0.246 ± 0.007	0.204 ± 0.006	9,234	8,609	439	186
Wild		351	351	0.177 ± 0.009	0.176 ± 0.009	0.147 ± 0.009	0.286 ± 0.008	0.165 ± 0.009	0.191 ± 0.009	0.185 ± 0.010	0.285 ± 0.008	0.035 ± 0.001	0.140 ± 0.001	0.111 ± 0.002	0.272 ± 0.002	0.166 ± 0.003	0.127 ± 0.008	0.150 ± 0.011	0.202 ± 0.008	8,577	7,979	412	186

TABLE 2 | Genetic diversity indices calculated using all SNPs and subsets of SNPs (neutral markers, directional outlier markers, and balancing outlier markers).

The estimated effective population size for each sampling location and/or timepoint calculated using linkage disequilibrium. Reported lower bound and upper bound numbers reflect a 95% confidence interval calculated using the jackknife method, a non-parametric method at a minimum allele frequency of 0.05. The average observed heterozygosity (H_o), expected heterozygosity (H_e), multilocus heterozygosity (MLH), minor allele frequency (MAF), and the number of polymorphic loci per sampling location and per STRUCTURE population designation (K = 2; domestic genetic cluster, wild genetic cluster). Bold text represents results for the two genetic clusters identified in structure.

for all populations and genetic clusters were similar to one another (Table 2).

The number of polymorphic loci per population ranged between 5,995 and 9,291 loci (61.0 - 94.5%), with domestic populations having 24.8% more polymorphic loci on average than the wild populations when all markers were considered (**Table 2**). A total of 565 private SNPs were identified within the domestic genetic cluster, while no private SNPs were identified within the wild genetic cluster.

As the number of polymorphic loci varied greatly between domestic and wild populations, the effect of missing data on genetic diversity indices was also examined (**Supplementary Material 9**). Markers with less than 50%, 25%, and 5% missingness in all samples were tested, as well as markers with a maximum of 5% missingness within a single population. As the percentage of missingness allowed per SNP decreased, the number of markers that passed this quality control measure also decreased. The number of polymorphic markers decreased from 61 to 95% when all markers were included to 44 – 67% when 50% missing data was allowed (**Supplementary Material 9**). The percentage of polymorphic markers was similar between 25% missing data (29–44%) and 5% missing data per population (27 – 48%; **Supplementary Material 9**).

In general, as the proportion of missing data allowed decreased, the number of polymorphic loci also decreased and estimates of observed and expected heterozygosity remained similar (± 0.01) or decreased, with the exception of Rosetta at 25% missing data (**Supplementary Material 9**). The marker set with only a total of 5% missingness per population allowed had the lowest number of polymorphic markers (8.5 – 16.4%), H_o, and H_e (**Supplementary Material 9**). In population groupings with a larger number of individuals sampled (121 – 470 samples), heterozygosity estimates were less affected and patterns remained more consistent than in groupings with fewer sampled individuals (28 - 50 samples; **Supplementary Material 9**). Rosetta (48 samples), Asyut (33 samples), and Aswan (20 samples) showed the greatest variability among marker subsets (**Table 2** and **Supplementary Material 9**).

DISCUSSION

This study used genome-wide SNP markers to (1) investigate population genetic structure, (2) detect signatures of selection in three generations of the AS and eight wild populations of Nile tilapia (*O. niloticus*; Aswan, Manzala Lagoon, Kanater, Lake Idku, Damietta, Lake Burullus, Rosetta, and Asyut) throughout the Nile River, Egypt, and (3) audit genetic diversity in the AS and wild populations.

Clear population genetic structuring was observed indicating that the domesticated AS genetic cluster has become genetically distinct from the wild genetic cluster in Egypt. The genetic distinction between the AS and wild populations is likely due to the initial bottleneck created by a small founding population, genetic drift and the subsequent selection for faster growth rates, larger sizes, and domestication within this limited population. This clear separation between wild and domestic populations has also been observed in Atlantic Salmon, *Salmo salar* (Gutierrez et al., 2016) and gilthead sea bream, *Sparus aurata* (Cossu et al., 2019). The effects of the bottleneck created by the small founding population for AS can be observed in the smaller effective population size (max. 48.6) of the domesticated AS in comparison to the wild effective population size (max "infinite"). Similar results have been seen in other aquaculture species, like Atlantic Salmon, *Salmo salar* (Domestic Ne 33–125, Wild Ne = 50->20,000; Bentsen and Thodesen, 2005), Pacific oyster, *Crassostrea gigas* (Domestic Ne = 47.6–58.5, Wild Ne = 527.9-infinite; Zhong et al., 2017), and gilthead sea bream, *Sparus aurata* (Domestic Ne = 21–111, Wild Ne = 133-infinity with the exception of one domestic population; Cossu et al., 2019).

The genetic difference between the AS and Aswan, one of the strain's founding populations was one of the largest observed. This is surprising, but not entirely unexpected given the AS's management history. A previous study found that the AS was created by two founding events with *O. niloticus* (Nayfa et al., 2020) in addition to the hybridization events with *O. aureus* (Grobler, 2017). Of the original founders, which included individuals from Aswan, only 53 of the original 201 founder genomes are present in Generations 9, 10, and 11 of the AS. From those 53 founder genomes, only 34 account for over 84% of the AS' genetic composition (Nayfa et al., 2020). Thus, it is likely that the Aswan founder genomes have been bred out of the AS.

Despite evidence of gene flow among the eight wild populations, isolation-by-distance was detected with the two most southern populations (Asyut and Aswan) being more distinct from the Nile Delta populations to the north than the geographically intermediate Kanater population. In addition to the effects of physical distance to gene flow and population structure, environmental factors may have also influenced this distinction between Delta and upstream riverine populations. Individuals within Delta populations, particularly Lake Idku, Lake Burullus, and Manzala Lagoon, which have a direct connection to the sea, live in brackish to freshwater conditions whilst the individuals within the upstream populations live in freshwater conditions (Hassanien et al., 2004; Balah, 2012).

These results are similar to those observed in 2004 and 2005 in two separate studies using microsatellites and randomly amplified polymorphic DNA (RAPD) where evidence of population sub-structuring was identified (Hassanien et al., 2004; Hassanien and Gilbey, 2005). Structuring in these studies was not only identified between geographically distant Nile Delta populations and upstream Egyptian Nile populations, but also amongst lake and river base populations in the Delta (Hassanien et al., 2004; Hassanien and Gilbey, 2005). However, unlike those studies, the present study observed no significant population structuring among Nile Delta populations. This disparity may be attributed to the difference in molecular technologies utilized between studies and the dramatic rise in aquaculture in Egypt (Soliman and Yacout, 2016).

Differences in molecular technologies have likely contributed to the disparities in population structure. For instance, Hassanien and Gilbey (2005) inferred the presence of null alleles based on lower levels of observed vs. expected heterozygosity levels in their microsatellite dataset. Null alleles in microsatellite studies can result in the overestimation of F_{st} and genetic distance (Chapuis and Estoup, 2006). Whereas, the RAPDs used in Hassanien et al. (2004) are limited by the fact that the majority of RAPD markers are dominant, making it impossible to determine whether a DNA segment is amplified from a homozygous or heterozygous locus (Kumar and Gurusubramanian, 2011). This can result in uncertain estimates to genetic structure (Fritsch and Rieseberg, 1996). Additionally, the molecular criteria which determine what constitutes population structure are flexible and can vary based on the organism, study question, and genetic markers used (Waples and Gaggiotti, 2006; Putman and Carbone, 2014).

Genetic technologies are not the only factor to have changed over the years. Since 2005, Egypt has experienced a considerable increase in extensive, semi-intensive, and intensive farming systems for Nile tilapia (Soliman and Yacout, 2016). The vast majority of these farms are located in the Nile Delta region and concentrated in the Northern Lakes (Maruti, Idku, Brulus, and Manzala Lagoon; Soliman and Yacout, 2016). As a result, increased movement of fish among hatcheries and farms has occurred in the region in that time. In addition, the number of fish escaping from farms has likely increased due to a combination of local weather conditions, including flash flooding events (Moawad et al., 2016), and farm practices. With five of the eight sampled locations in the Nile Delta regions, and farming occurring at or near the remaining three sampling locations (Soliman and Yacout, 2016), the genetic diversity of the wild populations may have been affected by exchange with farmed stocks.

A comparison of wild and domestic genetic clusters identified 674 outlier markers, with a higher proportion of markers deviating from HWE in domestic populations than wild populations. This is indicative of a finite population size and selective forces, such as artificial selection for marketable traits and domestication (Waples, 2014). The large amount of outliers detected concurs with other genetic studies of domestic vs. wild aquatic populations, including brown trout Salmo trutta L., (431 SNP outliers; Linløkken et al., 2017) and Atlantic salmon, Salmo salar L. (337 and 270 SNP outliers; López et al., 2019). Both balancing and diversifying outliers identified between domestic and native populations were found in every chromosome. Unlike other studies which found specific regions of the genome under selection when comparing domestic and wild populations (Marrano et al., 2018; López et al., 2019), there was a lack of localized clustering of outliers.

A limited number of outliers (0 – 11) detected in pairwise comparisons of wild populations is consistent with the limited genetic differentiation observed among the wild populations. The fact that the number of outliers detected increased with geographic distance from the upstream (Asyut, Kanater, and Aswan) to Nile Delta populations (Lake Idku, Rosetta, Lake Burullus, Damietta, and Manzala Lagoon) also reflects the isolation-by-distance determined using the whole data set. These results suggest that despite known differences in salinity levels in delta and upstream populations, there appears to be little or no effect on selection. This is not entirely surprising as Nile tilapia are known for their tolerance to a wide range of environmental conditions (Balarin, 1982; Avella et al., 1993; Shelton and Popma, 2006; Rebouças et al., 2016). Alternatively gene flow may be high enough between geographic regions to combat the forces of natural selection (Lenormand, 2002). Consequently, few outliers amongst wild populations indicate that the AS would be expected to perform similarity in different locations once disseminated throughout Egypt.

Differences in genetic diversity resulted in the domesticated AS being clearly distinguishable from wild populations. In general, genetic diversity indices indicate that AS populations have higher levels of heterozygosity than wild populations. This held true regardless of the number of SNPs and levels of missing data allowed. These results differ from what is traditionally seen in domesticated and/or selectively bred populations vs. wild populations where wild populations exhibit either higher levels of genetic diversity (Makino et al., 2018; Zamani et al., 2018), or similar levels of heterozygosity (Gutierrez et al., 2016). This may be explained by (1) hybridization with another tilapia species, (2) the isolation-by-distance observed in this study among current wild populations and (3) the historical development of fishing and aquaculture in Egypt.

The AS had a higher number of polymorphic markers and private SNPs (5.7% of all SNPs) than wild populations. While this may be a result of domestication or founder effects, it is suspected that introgression has occurred with blue tilapia (O. aureus). Blue tilapia from a population maintained at the Abbassa Station, Egypt have been observed in earthen ponds in AS facilities (Benzie, 2019; pers. comm.). This population of blue tilapia has now been removed from the Abbassa Station. Unpublished research by the WorldFish Center and affiliated researchers found that the AS is comprised of 10% O. aureus (blue tilapia; Grobler, 2017). This interpretation is further supported by the large number of outliers detected, as hybridization has been interpreted to explain the detection of outliers in other species (Cullingham et al., 2014) and species-specific SNPs are often picked up when developing SNPs from samples that include multiple species or hybrids (Liu et al., 2011; Silva-Junior et al., 2015). Thus, the incorporation of O. aureus in the AS genome may account for the high number of private SNPs identified in the AS, as well as the higher number of polymorphic markers and heterozygosity observed in the AS genetic cluster over the wild genetic cluster as these markers may have been speciesspecific SNPs. While the AS showed the greatest number of polymorphic loci, the wild populations all exhibited different subsets of polymorphic loci per sampling location, indicating that hybridization with O. aureus may have also occurred in the wild. Given that tilapia species are well known for hybridizing in both aquaculture and wild environments, this is unsurprising (Lovshin, 1982; D'Amato et al., 2007; Deines et al., 2014; Meier et al., 2019).

Despite the low level of genetic distinction between wild and domestic populations of Nile tilapia detected in the present research, putative AS escapees were easily identified, with suggested evidence of first and later generation escapees in Rosetta, Kanater, and Damietta detected. Escapees in other locales, particularly from selectively bred individuals, have been shown to lower the fitness of wild populations (Yang et al., 2019) as demonstrated in Atlantic salmon, *Salmo salar* (McGinnity et al., 2003; Glover et al., 2013); European sea bass, *Dicentrarchus labrax* (Toledo-Guedes et al., 2014); and Turbot, *Scophthalmus maximus* (Prado et al., 2018). It is not clear to what extent this may be a concern for tilapia, because while there was evidence in the present study of AS genetic material in wild Egyptian populations, to date, there is no information on fitness differentials between the domesticated AS and wild tilapia populations.

High levels of genetic diversity were still observed within the AS, suggesting that the potential detrimental effects on diversity of any AS escapees that do survive in wild populations may be minimal. This is particularly true as the AS was founded from both Nile Delta and upstream populations of Nile tilapia in Egypt. Thus, the genetic diversity observed in the AS is a subset of what is already available in wild populations. This in addition to the relatively low number of escapees detected when considering all wild populations, suggests that escapees may either be a rare occurrence or may have low survival within wild populations. This has been demonstrated previously in domestic rainbow trout (Oncorhynchus mykiss) who experience lower survival rates in the wild due to their increased size and bolder foraging habits exposing them to higher predation (Biro et al., 2004). Regardless, continued monitoring of escapees from the AS and other domestic lines is important as many wild Nile tilapia populations are at risk of an altered population structure and genetic diversity due to anthropogenic changes such as habitat disturbance, overfishing, and indiscriminate fish transfers of tilapia species throughout Africa (Eknath and Hulata, 2009).

CONCLUSION

The present study has highlighted the valuable information for improved management of aquaculture species by investigating population genetic structure, genetic diversity, and signatures of selection between domestic and wild populations. In the case of Nile tilapia in Egypt, domestic and wild populations were found easily distinguishable from one another using SNP markers, even when compared to founding populations. In turn, this distinct clustering allowed for easy detection of putative escapees. Although the wild genetic cluster was not panmictic, with wild populations displaying evidence of isolation-bydistance, levels of genetic differentiation were relatively low and no evidence of significant signatures of selection among wild populations were observed. After 11 years of selective breeding, the AS displayed high levels of genetic diversity. These data suggest that the AS could be disseminated throughout Egypt with negligible differences in performance expected and minimal disruption to wild populations. The genetic diversity comparisons also helped better understand how the effects of selection, founder effect, inbreeding, and genetic drift have affected this domestic line. The introgression with O. aureus may explain the large number of outliers detected between

wild and captive genetic clusters. While both balancing and diversifying outliers were traced back to all 22 *O. niloticus* chromosomes, additional research is required to determine the nature of these signatures and their direct relevance to biological or evolutionary processes within domestic and wild populations.

DATA AVAILABILITY STATEMENT

The datasets presented in this study can be found in online repositories. The names of the repository/repositories and accession number(s) can be found below: https://doi.org/10.7910/DVN/IFSGQF, https://dataverse.harvard.edu/dataset.xhtml?persistentId=doi:10.7910/DVN/IFSGQF.

ETHICS STATEMENT

Ethical review and approval was not required for the animal study because finclips for genetic sampling were acquired from a commercial farm during harvesting and from fishing boat catches. Written informed consent was obtained from the owners for the participation of their animals in this study.

AUTHOR CONTRIBUTIONS

MN conceptualized the project, acquired partial project funding, led the acquisition and formal analysis of the data as well as led the literature research and writing of the manuscript. DBJ supervised the data analysis and contributed to the review and editing of the manuscript. JB helped to conceptualize the project, acquired the majority of project funding, supervised the data acquisition, and contributed to the review and editing of the manuscript. KZ helped to conceptualize the project, supervised the execution of the project and analysis of data as well as contributed to the review and editing of the manuscript. DRJ helped to conceptualize the project, supervised the research project, acquired partial project funding, and contributed to editing the manuscript. All authors contributed to the article and approved the submitted version.

FUNDING

This publication was made possible through support provided by the European Commission-IFAD Grant Number 2000001539, the International Fund for Agricultural Development (IFAD), the United States Agency for International Development (USAID), the CGIAR Research Program on Livestock and Fish Agri-Food Systems (LIVESTOCK AND FISH), the CGIAR Research Program on Fish Agri-Food Systems (FISH) led by WorldFish, and a small grant from the Fisheries Society of the British Isles. This program was supported by contributors to the CGIAR Trust Fund.

ACKNOWLEDGMENTS

The authors wish to thank Khairul Rizal Abu Bakar for his aid in processing samples and attaining data, the breeding team in Egypt for collection of samples, Prof. Mohamed Said for collecting wild Nile tilapia samples for the project, Hooi Ling Khaw and Wagdy Mekkawy for their assistance in obtaining and interpreting data, Jarrod Guppy, Shannon Kjeldsen, Natalia Andrade, Katie Sambrook, Alyssa Budd, and Melissa Joyce for

REFERENCES

- Andueza-Noh, R. H., Martínez-Castillo, J., and Chacón-Sánchez, M. I. (2015). Domestication of small-seeded lima bean (*Phaseolus lunatus* L.) landraces in Mesoamerica: evidence from microsatellite markers. *Genetica* 143, 657–669. doi: 10.1007/s10709-015-9863-0
- Argue, B. J., Arce, S. M., Lotz, J. M., and Moss, S. M. (2002). Selective breeding of Pacific white shrimp (Litopenaeus vannamei) for growth and resistance to Taura Syndrome Virus. *Aquaculture* 204, 447–460. doi: 10.1016/s0044-8486(01)00830-4
- Avella, M., Berhaut, J., and Bornancin, M. (1993). Salinity tolerance of two tropical fishes, *Oreochromis aureus* and *O. niloticus*. I. Biochemical and morphological changes in the gill epithelium. *J. Fish Biol.* 42, 243–254. doi: 10.1111/j.1095-8649.1993.tb00325.x
- Balah, M. I. (2012). "North Delta Lakes, Egypt," in *Encyclopedia of Lakes and Reservoirs*, eds L. Bengtsson, R. W. Herschy, and R. W. Fairbridge (Dordrecht: Springer).
- Balarin, J. A. (1982). "The intensive culture of tilapia in tanks, raceways and cages," in *Recent Advances in Aquaculture*, eds J. F. Muir and R. J. Roberts (London: Helm Publishers), 265–356.
- Bentsen, H. B., and Thodesen, J. (2005). Genetic Interactions Between Farmed and Wild Fish, with Examples from the Atlantic Salmon Case in Norway. Selection and Breeding Programs in Aquaculture. Cham: Springer.
- Benzie, J. (2019). RE: Program Leader, Sustainable Aquaculture, Research Lead, Kenya at the WorldFish Center.
- Biro, P. A., Abrahams, M. V., Post, J. R., and Parkinson, E. A. (2004). Predators select against high growth rates and risk-taking behaviour in domestic trout populations. *Proc. R. Soc. Lond. Ser. B Biol. Sci.* 271, 2233–2237. doi: 10.1098/ rspb.2004.2861
- Burke, J. M., Knapp, S. J., and Rieseberg, L. H. (2005). Genetic consequences of selection during the evolution of cultivated sunflower. *Genetics* 171, 1933–1940. doi: 10.1534/genetics.104.039057
- Carter, D., Liti, G., Moses, A., Parts, L., James, S., Davey, R., et al. (2008). Population genomics of domestic and wild yeasts. *Nature* 458, 337–341.
- Chapuis, M.-P., and Estoup, A. (2006). Microsatellite null alleles and estimation of population differentiation. *Mol. Biol. Evol.* 24, 621–631. doi: 10.1093/molbev/ msl191
- Clutton-Brock, J. (1992). The process of domestication. Mammal Rev. 22, 79-85.
- Cossu, P., Scarpa, F., Sanna, D., Lai, T., Dedola, G. L., Curini–Galletti, M., et al. (2019). Influence of genetic drift on patterns of genetic variation: the footprint of aquaculture practices in *Sparus aurata* (Teleostei: Sparidae). *Mol. Ecol.* 28, 3012–3024. doi: 10.1111/mec.15134
- Courtois, B., Audebert, A., Dardou, A., Roques, S., Ghneim-Herrera, T., Droc, G., et al. (2013). Genome-wide association mapping of root traits in a japonica rice panel. *PLoS one* 8:e78037. doi: 10.1371/journal.pone.0078037
- Cullingham, C. I., Cooke, J. E., and Coltman, D. W. (2014). Cross-species outlier detection reveals different evolutionary pressures between sister species. *New Phytol.* 204, 215–229. doi: 10.1111/nph.12896
- D'Amato, M. E., Esterhuyse, M. M., Van Der Waal, B. C., Brink, D., and Volckaert, F. A. (2007). Hybridization and phylogeography of the Mozambique tilapia *Oreochromis mossambicus* in southern Africa evidenced by mitochondrial and microsatellite DNA genotyping. *Conserv. Genet.* 8, 475–488. doi: 10.1007/ s10592-006-9186-x

their advice and support, and Jacqueline Muliro for her help in data management.

SUPPLEMENTARY MATERIAL

The Supplementary Material for this article can be found online at: https://www.frontiersin.org/articles/10.3389/fgene. 2020.567969/full#supplementary-material

- Deines, A., Bbole, I., Katongo, C., Feder, J., and Lodge, D. (2014). Hybridisation between native Oreochromis species and introduced Nile tilapia O. niloticus in the Kafue River, Zambia. Afr. J. Aquat. Sci. 39, 23–34.
- Do, C., Waples, R. S., Peel, D., Macbeth, G., Tillett, B. J., and Ovenden, J. R. (2014). NeEstimator v2: re-implementation of software for the estimation of contemporary effective population size (Ne) from genetic data. *Mol. Ecol. Resour.* 14, 209–214. doi: 10.1111/1755-0998.12157
- Earl, D. A., and vonHoldt, B. M. (2012). STRUCTURE HARVESTER: a website and program for visualizing STRUCTURE output and implementing the Evanno method. *Conserv. Genet. Resour.* 4, 359–361. doi: 10.1007/s12686-011-9548-7
- Eknath, A. E., and Hulata, G. (2009). Use and exchange of genetic resources of Nile tilapia (Oreochromis niloticus). Rev. Aquacult. 1, 197–213. doi: 10.1111/j.1753-5131.2009.01017.x
- Evanno, G., Regnaut, S., and Goudet, J. (2005). Detecting the number of clusters of individuals using the software STRUCTURE: a simulation study. *Mol. Ecol.* 14, 2611–2620. doi: 10.1111/j.1365-294x.2005.02553.x
- Excoffier, L., and Lischer, H. E. (2010). Arlequin suite ver 3.5: a new series of programs to perform population genetics analyses under Linux and Windows. *Mol. Ecol. Resour.* 10, 564–567.
- Falush, D., Stephens, M., and Pritchard, J. K. (2003). Inference of population structure using multilocus genotype data: linked loci and correlated allele frequencies. *Genetics* 164, 1567–1587.
- Falush, D., Stephens, M., and Pritchard, J. K. (2007). Inference of population structure using multilocus genotype data: dominant markers and null alleles. *Mol. Ecol. Notes* 7, 574–578.
- FAO (2018). The State of World Fisheries and Aquaculture 2018 Meeting the Sustainable Development Goals. Rome: FAO.
- Foll, M. (2012). BayeScan v2. 1 user manual. Ecology 20, 1450-1462.
- Foll, M., and Gaggiotti, O. (2008). A genome-scan method to identify selected loci appropriate for both dominant and codominant markers: a Bayesian perspective. *Genetics* 180, 977–993.
- Fritsch, P., and Rieseberg, L. H. (1996). "The use of random amplified polymorphic DNA (RAPD) in conservation genetics," in *Molecular Genetic Approaches in Conservation*, Vol. 56, eds T. B. Smith and R. K. Wayne (London: Oxford University Press), 54–73.
- Glover, K. A., Pertoldi, C., Besnier, F., Wennevik, V., Kent, M., and Skaala, Ø (2013). Atlantic salmon populations invaded by farmed escapees: quantifying genetic introgression with a Bayesian approach and SNPs. *BMC Genet.* 14:74. doi: 10.1186/1471-2156-14-74
- Gogarten, S. M., Bhangale, T., Conomos, M. P., Laurie, C. A., Mchugh, C. P., Painter, I., et al. (2012). Gwastools: an R/Bioconductor package for quality control and analysis of genome-wide association studies. *Bioinformatics* 28, 3329–3331.
- Gondro, C., Van Der Werf, J., and Hayes, B. J. (2013). Genome-Wide Association Studies and Genomic Prediction. Cham: Springer.
- Grobler, M.-T. (2017). Sex Determination in the WorldFish Abbassa Strain of Nile Tilapia (Oreochromis niloticus L.). Ph. D. Thesis, University of Stirling, Stirling.
- Gutierrez, A., Yáñez, J., and Davidson, W. (2016). Evidence of recent signatures of selection during domestication in an Atlantic salmon population. *Mar. Genomics* 26, 41–50.
- Hassanien, H. A., Elnady, M., Obeida, A., and Itriby, H. (2004). Genetic diversity of *Nile tilapia* populations revealed by randomly amplified polymorphic DNA (RAPD). *Aquacult. Res.* 35, 587–593. doi: 10.1111/j.1365-2109.2004.01057.x

- Hassanien, H. A., and Gilbey, J. (2005). Genetic diversity and differentiation of Nile tilapia (Oreochromis niloticus) revealed by DNA microsatellites. Aquacult. Res. 36, 1450–1457. doi: 10.1111/j.1365-2109.2005.01368.x
- Hayes, B. (2013). "Overview of statistical methods for genome-wide association studies (GWAS)," in *Genome-Wide Association Studies and Genomic Prediction*, eds C. Gondro, J. van der Werf, and B. Hayes (Totowa, NJ: Humana Press). doi: 10.1007/978-1-62703-447-0_6
- Heller-Uszynska, K., Uszynski, G., Huttner, E., Evers, M., Carlig, J., Caig, V., et al. (2011). Diversity arrays technology effectively reveals DNA polymorphism in a large and complex genome of sugarcane. *Mol. Breed.* 28, 37–55. doi: 10.1007/ s11032-010-9460-y
- Hijmans, R. J., Williams, E., Vennes, C., and Hijmans, M. R. J. (2017). Package 'geosphere'. R Package Version, 3. Available online at: https://cran.r-project.org/ web/packages/geosphere/geosphere.pdf
- Hossain, S., Panozzo, J., Pittock, C., and Ford, R. (2011). Quantitative trait loci analysis of seed coat color components for selective breeding in chickpea (*Cicer* arietinum L.). Can. J. Plant Sci. 91, 49–55. doi: 10.4141/cjps10112
- Hubisz, M. J., Falush, D., Stephens, M., and Pritchard, J. K. (2009). Inferring weak population structure with the assistance of sample group information. *Mol. Ecol. Resour.* 9, 1322–1332. doi: 10.1111/j.1755-0998.2009.02591.x
- Ibrahim, N. A., Zaid, M. Y. A., Khaw, H. L., El–Naggar, G. O., and Ponzoni, R. W. (2013). Relative performance of two *Nile tilapia (Oreochromis niloticus* L innaeus) strains in Egypt: the Abbassa selection line and the K afr E l S heikh commercial strain. *Aquacult. Res.* 44, 508–517. doi: 10.1111/j.1365-2109.2012. 03240.x
- Jaccoud, D., Peng, K., Feinstein, D., and Kilian, A. (2001). Diversity arrays: a solid state technology for sequence information independent genotyping. *Nucleic Acids Res.* 29:e25. doi: 10.1093/nar/29.4.e25
- Johnsson, J. I., Petersson, E., Jönsson, E., Björnsson, B. T., and Järvi, T. (1996). Domestication and growth hormone alter antipredator behaviour and growth patterns in juvenile brown trout, *Salmo trutta. Can. J. Fish. Aquat. Sci.* 53, 1546–1554. doi: 10.1139/f96-090
- Jombart, T. (2008). adegenet: a R package for the multivariate analysis of genetic markers. *Bioinformatics* 24, 1403–1405. doi: 10.1093/bioinformatics/btn129
- Jombart, T., and Ahmed, I. (2011). adegenet 1.3-1: new tools for the analysis of genome-wide SNP data. *Bioinformatics* 27, 3070–3071. doi: 10.1093/ bioinformatics/btr521
- Jones, A. T., Ovenden, J. R., and Wang, Y. G. (2016). Improved confidence intervals for the linkage disequilibrium method for estimating effective population size. *Heredity* 117, 217–223. doi: 10.1038/hdy.2016.19
- Kalinowski, S. T. (2011). The computer program STRUCTURE does not reliably identify the main genetic clusters within species: simulations and implications for human population structure. *Heredity* 106:625. doi: 10.1038/hdy.2010.95
- Kilian, A., Wenzl, P., Huttner, E., Carling, J., Xia, L., Blois, H., et al. (2012).
 "Diversity arrays technology: a generic genome profiling technology on open platforms," in *Data Production and Analysis in Population Genomics*, eds F. Pompanon and A. Bonin (Totowa, NJ: Humana Press). doi: 10.1007/978-1-61779-870-2_5
- Kjeldsen, S. R., Raadsma, H. W., Leigh, K. A., Tobey, J. R., Phalen, D., Krockenberger, A., et al. (2019). Genomic comparisons reveal biogeographic and anthropogenic impacts in the koala (*Phascolarctos cinereus*): a dietaryspecialist species distributed across heterogeneous environments. *Heredity* 122:525. doi: 10.1038/s41437-018-0144-4
- Kumar, N. S., and Gurusubramanian, G. (2011). Random amplified polymorphic DNA (RAPD) markers and its applications. *Sci. Vis.* 11, 116–124.
- Ladizinsky, G. (1985). Founder effect in crop-plant evolution. *Econ. Bot.* 39, 191–199. doi: 10.1007/BF02907844
- Lawson, D. J., Van Dorp, L., and Falush, D. (2018). A tutorial on how not to overinterpret STRUCTURE and ADMIXTURE bar plots. *Nat. Commun.* 9:3258. doi: 10.1038/s41467-018-05257-7
- Lenormand, T. (2002). Gene flow and the limits to natural selection. *Trends Ecol. Evol.* 17, 183–189. doi: 10.1016/S0169-5347(02)02497-7
- Li, W., and Godzik, A. (2006). Cd-hit: a fast program for clustering and comparing large sets of protein or nucleotide sequences. *Bioinformatics* 22, 1658–1659. doi: 10.1093/bioinformatics/btl158
- Lind, C., Kilian, A., and Benzie, J. (2017). Development of diversity arrays technology markers as a tool for rapid genomic assessment in *Nile tilapia*, *Oreochromis niloticus. Anim. Genet.* 48, 362–364. doi: 10.1111/age.12536

- Linløkken, A. N., Haugen, T. O., Kent, M. P., and Lien, S. (2017). Genetic differences between wild and hatchery-bred brown trout (*Salmo trutta* L.) in single nucleotide polymorphisms linked to selective traits. *Ecol. Evol.* 7, 4963–4972. doi: 10.1002/ece3.3070
- Liu, S., Zhou, Z., Lu, J., Sun, F., Wang, S., Liu, H., et al. (2011). Generation of genome-scale gene-associated SNPs in catfish for the construction of a highdensity SNP array. *BMC Genomics* 12:53. doi: 10.1186/1471-2164-12-53 doi: 10.1186/1471-2164-12-53
- López, M. E., Benestan, L., Moore, J. S., Perrier, C., Gilbey, J., Di Genova, A., et al. (2019). Comparing genomic signatures of domestication in two Atlantic salmon (*Salmo salar* L.) populations with different geographical origins. *Evol. Appl.* 12, 137–156. doi: 10.1111/eva.12689
- Lovshin, L. (1982). "Tilapia hybridization," in Proceedings of the International Conference on the Biology and Culture of Tilapias, Bellagio (Italy), 2-5 Sep 1980, (Manila: International Center for Living Aquatic Resources Management).
- Makino, T., Rubin, C.-J., Carneiro, M., Axelsson, E., Andersson, L., and Webster, M. T. (2018). Elevated proportions of deleterious genetic variation in domestic animals and plants. *Genome Biol. Evol.* 10, 276–290. doi: 10.1093/gbe/evy004
- Marrano, A., Micheletti, D., Lorenzi, S., Neale, D., and Grando, M. S. (2018). Genomic signatures of different adaptations to environmental stimuli between wild and cultivated *Vitis vinifera* L. *Horticult. Res.* 5:34. doi: 10.1038/s41438-018-0041-2
- McGinnity, P., Prodöhl, P., Ferguson, A., Hynes, R., Maoiléidigh, N. Ó, Baker, N., et al. (2003). Fitness reduction and potential extinction of wild populations of Atlantic salmon, Salmo salar, as a result of interactions with escaped farm salmon. *Proc. R. Soc. Lond. Ser. B Biol. Sci.* 270, 2443–2450. doi: 10.1098/rspb. 2003.2520
- Meier, J. I., Stelkens, R. B., Joyce, D. A., Mwaiko, S., Phiri, N., Schliewen, U. K., et al. (2019). The coincidence of ecological opportunity with hybridization explains rapid adaptive radiation in Lake Mweru cichlid fishes. *Nat. Commun.* 10, 1–11. doi: 10.1038/s41467-019-13278-z
- Mignon-Grasteau, S., Boissy, A., Bouix, J., Faure, J.-M., Fisher, A. D., Hinch, G. N., et al. (2005). Genetics of adaptation and domestication in livestock. *Livestock Prod. Sci.* 93, 3–14. doi: 10.1016/j.livprodsci.2004.11.001
- Moawad, M. B., Abdel Aziz, A. O., and Mamtimin, B. (2016). Flash floods in the Sahara: a case study for the 28 January 2013 flood in Qena, Egypt. *Geomat. Nat. Hazard. Risk* 7, 215–236. doi: 10.1080/19475705.2014.885467
- Moss, S. M., Moss, D. R., Arce, S. M., Lightner, D. V., and Lotz, J. M. (2012). The role of selective breeding and biosecurity in the prevention of disease in penaeid shrimp aquaculture. *J. Invertebr. Pathol.* 110, 247–250. doi: 10.1016/j.jip.2012. 01.013
- Nayfa, M. G., Jones, D. B., Lind, C. E., Benzie, J. A. H., Jerry, D. R., and Zenger, K. R. (2020). Pipette and paper: combining molecular and genealogical methods to assess a *Nile tilapia (Oreochromis niloticus)* breeding program. *Aquaculture* 523:735171.
- Neuditschko, M., Khatkar, M. S., and Raadsma, H. W. (2012). NetView: a high-definition network-visualization approach to detect fine-scale population structures from genome-wide patterns of variation. *PLoS One* 7:e48375. doi: 10.1371/journal.pone.0048375
- Ollivier, L. (2002). Eléments de Génétique Quantitative: 2e Édition Revue et Augmentée. Versailles: Editions Quae.
- Paquette, S. R., and Paquette, M. S. R. (2011). Package 'PopGenKit'.
- Peakall, R., and Smouse, P. E. (2006). Genalex 6: genetic analysis in Excel. Population genetic software for teaching and research. *Mol. Ecol. Notes* 6, 288–295. doi: 10.1111/j.1471-8286.2005.01155.x
- Peakall, R., and Smouse, P. E. (2012). GenAlEx 6.5: genetic analysis in Excel. Population genetic software for teaching and research—an update. *Bioinformatics* 28, 2537–2539. doi: 10.1093/bioinformatics/bts460
- Peterson, B. K., Weber, J. N., Kay, E. H., Fisher, H. S., and Hoekstra, H. E. (2012). Double digest RADseq: an inexpensive method for de novo SNP discovery and genotyping in model and non-model species. *PLoS One* 7:e37135. doi: 10.1371/journal.pone.0037135 doi: 10.1371/journal.pone.0037135
- Prado, F., Vera, M., Hermida, M., Blanco, A., Bouza, C., Maes, G., et al. (2018). Tracing the genetic impact of farmed turbot *Scophthalmus maximus* on wild populations. *Aquacult. Environ. Interact.* 10, 447–463. doi: 10.3354/aei0 0282
- Price, E. O. (1984). Behavioral aspects of animal domestication. Q. Rev. Biol. 59, 1–32. doi: 10.1086/519795

- Pritchard, J. K., Stephens, M., and Donnelly, P. (2000). Inference of population structure using multilocus genotype data. *Genetics* 155, 945–959.
- Purcell, S. (2020). Plink v.1.9. Available online at: http://pngu.mgh.harvard.edu/ purcell/plink/ (accessed January 8, 2018).
- Purcell, S., Neale, B., Todd-Brown, K., Thomas, L., Ferreira, M. A., Bender, D., et al. (2007). PLINK: a tool set for whole-genome association and population-based linkage analyses. *Am. J. Hum. Genet.* 81, 559–575. doi: 10.1086/519795
- Putman, A. I., and Carbone, I. (2014). Challenges in analysis and interpretation of microsatellite data for population genetic studies. *Ecol. Evol.* 4, 4399–4428. doi: 10.1002/ece3.1305
- Raman, H., Raman, R., Kilian, A., Detering, F., Carling, J., Coombes, N., et al. (2014). Genome-wide delineation of natural variation for pod shatter resistance in *Brassica napus*. *PLoS One* 9:e101673. doi: 10.1371/journal.pone.0101673 doi: 10.1371/journal.pone.0101673
- Rebouças, V. T., Lima, F. R. D. S., and Cavalcante, D. D. H. (2016). Reassessment of the suitable range of water pH for culture of Nile tilapia Oreochromis niloticus L. in eutrophic water. Acta Sci. Anim. Sci. 38, 361–368. doi: 10.4025/ actascianimsci.v38i4.32051
- Rezk, M. A., Ponzoni, R. W., Khaw, H. L., Kamel, E., Dawood, T., and John, G. (2009). Selective breeding for increased body weight in a synthetic breed of Egyptian Nile tilapia, *Oreochromis niloticus*: response to selection and genetic parameters. *Aquaculture* 293, 187–194. doi: 10.1016/j.aquaculture.2009.03.019
- Robinson, N., and Hayes, B. (2008). Modelling the use of gene expression profiles with selective breeding for improved disease resistance in Atlantic salmon (Salmo salar). Aquaculture 285, 38–46. doi: 10.1016/j.aquaculture.2008.08.016
- Sansaloni, C., Petroli, C., Jaccoud, D., Carling, J., Detering, F., Grattapaglia, D., et al. (2011). Diversity arrays technology (DArT) and next-generation sequencing combined: genome-wide, high throughput, highly informative genotyping for molecular breeding of Eucalyptus. *BMC Proc.* 5:54. doi: 10.1186/1753-6561-5-S7-P54
- Scandura, M., Iacolina, L., and Apollonio, M. (2011). Genetic diversity in the European wild boar Sus scrofa: phylogeography, population structure and wild x domestic hybridization. Mammal Rev. 41, 125–137. doi: 10.1111/j.1365-2907. 2010.00182.x
- Schultz, A., Cristescu, R., Littleford-Colquhoun, B., Jaccoud, D., and Frère, C. (2018). Fresh is best: accurate SNP genotyping from koala scats. *Ecol Evol.* 8, 3139–3151. doi: 10.1002/ece3.3765
- Shelton, W. L., and Popma, T. J. (2006). "Biology," in *Tilapia: Biology, Culture, and Nutrition*, eds C. D. Webster and C. Lim (Boca Raton, FL: CRC Press).
- Silva–Junior, O. B., Faria, D. A., and Grattapaglia, D. (2015). A flexible multi–species genome–wide 60K SNP chip developed from pooled resequencing of 240 Eucalyptus tree genomes across 12 species. *New Phytol.* 206, 1527–1540. doi: 10.1111/nph.13322
- Simmons, M., Mickett, K., Kucuktas, H., Li, P., Dunham, R., and Liu, Z. (2006). Comparison of domestic and wild channel catfish (Ictalurus punctatus) populations provides no evidence for genetic impact. *Aquaculture* 252, 133– 146. doi: 10.1016/j.aquaculture.2005.11.006

- Soliman, N. F., and Yacout, D. M. (2016). Aquaculture in Egypt: status, constraints and potentials. *Aquacult. Int.* 24, 1201–1227. doi: 10.1007/s10499-016-9989-9
- Steinig, E. J., Neuditschko, M., Khatkar, M. S., Raadsma, H. W., and Zenger, K. R. (2016). netview p: a network visualization tool to unravel complex population structure using genome–wide SNPs. *Mol. Ecol. Resour.* 16, 216–227. doi: 10. 1111/1755-0998.12442
- Stoffel, M. A., Esser, M., Kardos, M., Humble, E., Nichols, H., David, P., et al. (2016). inbreedR: an R package for the analysis of inbreeding based on genetic markers. *Methods Ecol. Evol.* 7, 1331–1339. doi: 10.1111/2041-210X.12588
- Toledo-Guedes, K., Sanchez-Jerez, P., Benjumea, M. E., and Brito, A. (2014). Farming-up coastal fish assemblages through a massive aquaculture escape event. *Mar. Environ. Res.* 98, 86–95. doi: 10.1016/j.marenvres.2014. 03.009
- Von Mark, V. C., Kilian, A., and Dierig, D. A. (2013). Development of DArT marker platforms and genetic diversity assessment of the US collection of the new oilseed crop lesquerella and related species. *PLoS One* 8:e64062. doi: 10.1371/journal.pone.0064062 doi: 10.1371/journal.pone.0064062
- Waples, R. S. (2014). Testing for Hardy–Weinberg proportions: have we lost the plot? J. Heredity 106, 1–19. doi: 10.1093/jhered/esu062
- Waples, R. S., and Gaggiotti, O. (2006). Invited review: what is a population? An empirical evaluation of some genetic methods for identifying the number of gene pools and their degree of connectivity. *Mol. Ecol.* 15, 1419–1439. doi: 10.1111/j.1365-294X.2006.02890.x
- Yang, L., Waples, R. S., and Baskett, M. L. (2019). Life history and temporal variability of escape events interactively determine the fitness consequences of aquaculture escapees on wild populations. *Theor. Populat. Biol.* 129, 93–102. doi: 10.1016/j.tpb.2018.12.006
- Zamani, W., Ghasempouri, S. M., Rezaei, H. R., Naderi, S., Hesari, A. R. E., and Ouhrouch, A. (2018). Comparing polymorphism of 86 candidate genes putatively involved in domestication of sheep, between wild and domestic Iranian sheep. *Meta Gene* 17, 223–231. doi: 10.1016/j.mgene.2018.06.015
- Zhong, X., Li, Q., Kong, L., and Yu, H. (2017). Estimates of linkage disequilibrium and effective population size in wild and selected populations of the pacific oyster using single–nucleotide polymorphism markers. J. World Aquacult. Soc. 48, 791–801. doi: 10.1111/jwas.12393

Conflict of Interest: The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest.

Copyright © 2020 Nayfa, Jones, Benzie, Jerry and Zenger. This is an open-access article distributed under the terms of the Creative Commons Attribution License (CC BY). The use, distribution or reproduction in other forums is permitted, provided the original author(s) and the copyright owner(s) are credited and that the original publication in this journal is cited, in accordance with accepted academic practice. No use, distribution or reproduction is permitted which does not comply with these terms.