Title: Temperature is the evil twin: Effects of increased temperature and ocean acidification on reproduction in a reef fish.

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Abstract

Reproduction in many organisms can be disrupted by changes to the physical environment, such as those predicted to occur during climate change. Marine organisms face the dual climate change threats of increasing temperature and ocean acidification, yet no studies have examined the potential interactive effects of these stressors on reproduction in marine fishes. We used a long-term experiment to test the interactive effects of increased temperature and CO_2 on the reproductive performance of the anemonefish, *Amphiprion melanopus*. Adult breeding pairs were kept for 10 months at three temperatures, 28.5°C (+0.0°C), 30.0°C (+1.5°C) and 31.5°C (+3.0°C), cross-factored with 3 CO₂ levels, a current day control (417µatm) and moderate (644µatm) and high (1134µatm) treatments consistent with the range of CO_2 projections for the year 2100 under RCP8.5. We recorded each egg clutch produced during the breeding season, the number of eggs laid per clutch, average egg size,

fertilization success, survival to hatching, hatchling length and yolk provisioning. Adult body condition, hepatosomatic index, gonadosomatic index, and plasma 17β-estradiol concentrations were measured at the end of the breeding season to determine the effect of prolonged exposure to increased temperature and elevated CO₂ on adults, and to examine potential physiological mechanisms for changes in reproduction. Temperature had by far the stronger influence on reproduction, with clear declines in reproduction occurring in the +1.5°C treatment and ceasing altogether in the +3.0°C treatment. In contrast, CO₂ had a minimal effect on the majority of reproductive traits measured, but caused a decline in offspring quality in combination with elevated temperature. We detected no significant effect of temperature or CO_2 on adult body condition or hepatosomatic index. Elevated temperature had a significant negative effect on plasma 17β -estradiol concentrations, suggesting that declines in reproduction with increasing temperature were due to the thermal sensitivity of reproductive hormones rather than a reduction in energy available for reproduction. Our results show that elevated temperature exerts a stronger influence than high CO_2 on reproduction in A. melanopus. Understanding how these two environmental variables interact to affect the reproductive performance of marine organisms will be important for predicting the future impacts of climate change.

Keywords: Reproduction, ocean acidification, carbon dioxide, ocean warming, *Amphiprion melanopus*, climate change

1 Introduction

2	Reproduction is critical to individual fitness and the persistence of populations. Reproduction
3	in most organisms is also sensitive to changes in the physical environment. For example, the
4	timing of reproduction can be influenced by variation in temperature (Kjesbu, 1994; Visser et
5	al., 2009), photoperiod (Duston & Bromage, 1986; Dawson et al., 2001), rainfall (Donnelly
6	& Guyer, 1994; Hau et al., 2004) and flow regimes (Schlosser, 1982; Bunn & Arthington,
7	2002). Similarly, reproductive output is affected by temperature (King et al., 2003; Saino et
8	al., 2004) and food availability (Brown & Shine, 2007; Donelson et al., 2010). Consequently,
9	anthropogenic climate change is predicted to affect reproductive success of many species
10	(Parmesan, 2006; Poloczanska et al., 2013) and could be the primary driver of population
11	declines due to climate change (Van Der Kraak & Pankhurst, 1997; Zeh et al., 2012).
12	
13	For marine organisms, increasing temperature and ocean acidification are the most serious
14	climate change threats (Hoegh-Guldberg et al., 2007; Doney et al., 2009) and they are
15	predicted to be additive or synergistic in their effect on performance, potentially leading to
16	greater effects in combination than in isolation (Pörtner & Farrel, 2008). Many studies have
17	examined the effects of one or other of these two stressors on reproduction in marine
18	organisms (including Miller et al., 2013; Donelson et al., 2010), but few have examined the
19	potential interactive effects on reproduction. While an increasing number of studies are
20	testing the interacting effects of ocean warming and ocean acidification on invertebrates
21	(Byrne et al., 2009, 2010; Parker et al., 2009; Albright & Mason 2013, Cohen-Rengifo et al.,
22	2013) relatively few studies have tested these two stressors in combination for fish (but see
23	Munday et al., 2009a; Nowicki et al., 2012; Grans et al., 2014) and none have tested the
24	interacting effects on reproduction in fishes. In fishes, increased temperature has been shown
25	to negatively affect reproduction in many species (see reviews, Pankhurst & Porter, 2003;

26	Pankhurst & King, 2010). In contrast, ocean acidification, while predicted to have negative
27	impacts (Pörtner & Farrel, 2008; Ishimatsu et al., 2008) has been found to have little impact,
28	or even positive effects, on reproduction in multiple species of fish (Frommel et al., 2010;
29	Sundin et al., 2012, Miller et al., 2013; Fosgren et al., 2013). Yet, whether elevated
30	temperature and ocean acidification will interact to affect reproduction in fishes is not known.
31	
32	For many species of fish, temperature is one of the main cues for reproduction, signaling the
33	beginning and the end of the breeding season (Van Der Kraak & Pankhurst, 1997; Pankhurst
34	& Munday, 2011). For spring-summer spawner's, the increase in water temperatures
35	following the winter minimum, elicits physiological changes, production of sex steroids,
36	maturation of gonads, and spawning (Kjesbu, 1994; Pankhurst et al., 1996). Nevertheless,
37	reproduction only occurs within a narrow range of temperatures that the population normally
38	experiences (Van der Kraak & Pankhurst, 1997). If temperatures exceed this thermal
39	window, reproduction can quickly decline and may cease altogether (Donelson et al., 2010;
40	Dorts et al., 2011). Tropical fishes may be especially sensitive to changes in temperature, as
41	they inhabit a more thermally stable environment than higher latitude species (Tewksbury et
42	al., 2008; Rummer et al., 2014). This means that even a relatively small increase in average
43	temperature, such as predicted by climate change, could have serious effects on reproductive
44	performance in tropical species (Donelson et al., 2010; Zeh et al., 2012).
45	
46	Reproduction may decline at elevated temperatures as a result of energetic constraints or
47	through the effects of temperature on hormonal pathways. As temperatures increase past the
48	thermal optimum, individuals need to expend more energy maintaining cellular function
49	(Pörtner & Farrell, 2008). Organisms have a finite amount of energy available and as more of

50 the energy is used for homeostasis less is available for other activities, such as reproduction

51	(Somero, 2002; Sokolova et al., 2012). Under energy constraints, adults could opt to produce
52	the same number of offspring as under normal conditions, but at a cost to offspring
53	provisioning. Alternatively, individuals may produce fewer offspring that have adequate
54	levels of provisioning in an attempt to ensure offspring survival (Stearns, 1992).

56 Reproduction may also decline with increasing temperature due to the thermal sensitivity of 57 reproductive hormones. Reproduction in fish is tightly controlled through the interplay of 58 multiple hormones and steroids created by the hypothalamus, the pituitary and the gonads 59 (Hypothalamic-Pituitary-Gonadal axis (HPG axis)) (Yaron & Levavi-Sivan, 2011). Elevated 60 temperatures have the ability to inhibit the HPG axis at multiple sites, through changes in 61 hormone synthesis, action and structures (Pankhurst & Munday, 2011). The inhibitory effects 62 of temperature can occur through changes in protein and hormone structures, resulting in a 63 reduced uptake or insolubility of the hormones. These changes can then lead to the hormones 64 failing to reach the correct receptor, or passing straight through the kidneys and being 65 excreted, thereby impairing the particular reproductive process (Van Der Kraak & Pankhurst, 66 1997; Pankhurst & Munday, 2011). Ultimately, wherever the disruption to the hormonal 67 cascade occurs, elevated temperatures result in declines in reproductive activity, egg size and 68 offspring survival.

69

In addition to increasing temperatures, marine fishes will have to cope with increasing partial pressure of carbon dioxide (pCO_2) in the ocean. Increasing pCO_2 has been documented to negatively impact reproduction in a number of invertebrates (see Ross *et al.*, 2011). Fishes, however, have well-developed mechanisms for acid-base regulation and are able to maintain their internal pH against an elevated CO₂ gradient, through active transport of ions across the gills and in their blood and tissues (Brauner & Baker, 2009; Esbaugh *et al.*, 2012). This

76	process is not cost free and it has been predicted that the increase in energy required to
77	maintain acid-base balance should result in a decline in energy available for reproduction and
78	other activities (Pörtner et al., 2004; Ishimatsu et al., 2008). However, only one study (Inaba
79	et al., 2003) has documented a negative impact of increasing pCO_2 on a reproductive trait,
80	with sperm motility being reduced in some flatfishes, but not in a range of other species.
81	Other studies have reported little to no effects of increased CO ₂ on reproduction. For
82	example, Frommel et al. (2010) found no effect on sperm motility in Baltic cod (Gadhus
83	morhua), Sundin et al. (2012) found no difference in reproductive propensity in pipefish
84	(Syngnathus typhle) and Fosgren et al. (2013) found no differences in clutch size but did see
85	a significant decline in egg survival with increasing CO ₂ in a temperate goby (Gobiusculus
86	flavescens). Interestingly, several studies have documented increases in reproduction or
87	reproductive related traits in response to elevated CO ₂ (Miller et al., 2013; Schade et al.,
88	2014) Furthermore, Preus-Olsen et al. (2014) documented increased levels of sex steroid
89	hormones in Atlantic cod at high CO ₂ which is consistent with greater rates of reproduction
90	in the other studies. Neither Miller et al. (2013) or Schade et al. (2014) found negative
91	consequences of the increased reproductive activity in high CO ₂ on the condition of the
92	adults or the resulting offspring. Instead these studies show transgenerational acclimation of
93	the offspring to elevated CO ₂ due to parental exposure to high CO ₂ (Miller <i>et al.</i> , 2012:
94	Schade et al., 2014).

96 Increases in temperature and pCO_2 will not occur in isolation from each other, and for that 97 reason, it is important to understand how they may interact to affect reproduction. As these 98 two variables have been documented to have contrasting effects on reproduction in fish, it is 99 especially important to understand how they might interact to affect this critical process. The 100 aim of this study was to document the effect of elevated CO_2 and increased temperature on

101	reproductive activity, offspring quality, and any effect on adult condition (physical and
102	reproductive) in a tropical reef fish. Adult pairs of Amphiprion melanopus were kept in
103	current-day control CO ₂ or elevated CO ₂ treatments (moderate and high). CO ₂ treatments
104	were fully cross-factored with 3 temperature treatments, current-day summer average water
105	temperature, 28.5°C (+0.0°C), or two elevated temperatures, 30.0°C (+1.5°C) and 31.5°C
106	(+3.0°C). Adult pairs were placed in CO_2 treatment during winter, slowly brought up to the
107	required temperature treatments through spring and then allowed to reproduce naturally
108	during the summer breeding season. Throughout the breeding season we assessed the effect
109	of elevated temperature and increased CO ₂ on key reproductive traits related to breeding and
110	spawning, egg production and survival, and offspring provisioning (Fig. 1). At the end of the
111	reproductive season we assessed adult physiological (Fulton's K body condition index,
112	hepatosomatic index) and reproductive condition (gonadosomatic index, plasma hormone
113	concentration) (Fig. 1) to determine if difference in reproductive performance were
114	potentially associated with the energetic cost of reproduction or effects on reproductive
115	hormones. 17 β -estradiol (E ₂) was chosen as the focal sex steroid due to its well-defined role
116	in vitellogenesis and oocyte maturation in female fish (Lubzen et al., 2010; Yaron & Levavi-
117	Sivan, 2011).

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119 Methods

120 *Study species and husbandry*

121 The cinnamon anemonefish, *Amphiprion melanopus* (Pomacentridae) inhabits coral reefs

- 122 throughout the Indo-Pacific region, including the Great Barrier Reef, Australia (Drew et al.,
- 123 2008). *Amphiprion melanopus* occur in large social groups containing multiple breeding pairs
- that reproduce repeatedly during the summer. Eggs are laid in clutches attached to hard
- 125 substratum near their host anemone (Michael, 2008). Embryonic duration is 7-9 days, during

- 126 which time the males tend the eggs. Larvae hatch after dark and are pelagic for
- 127 approximately 11 days, at which point they metamorphose and become competent to settle to
- 128 reef habitat (Bay et al., 2006). Adults reach a maximum length of 12cm (Lieske & Myers,
- 129 1994) and have a reported maximum age of 5 years (Allen, 1975).
- 130
- 131 Adult breeding pairs of *A. melanopus* were collected between June 2009 and June 2011 from
- 132 4 reefs in the central Great Barrier Reef: Orpheus Island (18.6183°S, 146.4936°E), Bramble
- 133 Reef (18.417°S, 146.700°E), Davies Reef (18.83°S, 147.63°E) and Slasher's Reef (18.467°S,
- 134 147.083°E) and transferred to James Cook University, Townsville. Pairs were housed
- individually in 45L aquaria and provided with a half terracotta post as a nest site and shelter.
- 136 Aquaria were provided with continuous flow of seawater at 1.5Lmin⁻¹. Pairs were fed 0.1g of
- 137 commercial fish feed (INVE NRD 12/24) three times a day, equivalent to 1.21% of the
- 138 average body weight (Donelson *et al.*, 2010).
- 139

140 Experimental design

141 Between seven and eight adult pairs of A. melanopus were assigned to each of the nine 142 treatment groups and held individually, i.e. one pair per aquaria. Three CO₂ groups were 143 cross-factored with three temperature groups reflective of pCO_2 and temperatures projected 144 to occur in the ocean by 2100 under RCP6 and RCP8.5 (Collins *et al.*, 2013). The three CO₂ 145 levels used were a current-day control (\sim 417µatm), a moderate (\sim 644 µatm) and a high 146 (\sim 1134 µatm) CO₂ treatment. The temperatures were the current-day summer average of the 147 collection region, 28.5°C ($\pm 0.0^{\circ}$ C), a moderate 30.0°C ($\pm 1.5^{\circ}$ C) and a high, 31.5°C ($\pm 3.0^{\circ}$ C) 148 temperature treatments (See Table 1 for full experimental parameters). These temperatures 149 reflect the current summer average water temperature for the Orpheus Island region where 150 the adults were collected, and the 1.5 to 3.0°C warming predicted to occur in the tropical

151 oceans over the coming century due to climate change (Poloczanska *et al.*, 2007; Ganachaud
152 *et al.*, 2011).

153

Three 8000L liter aquarium systems were utilized, each dosed to the desired pCO_2 . Due to the scale of the study it was not possible to have separate CO_2 and temperature treatments applied to each aquarium. Pairs were in their respective CO_2 treatments by June 2011 at current day winter water temperatures (22.5°C). Temperatures were subsequently increased over a two month period to achieve the required temperature separation and then increased by 0.5°C weekly to reach experimental breeding temperatures in the second week of November 2011.

161

162 Data collection of key reproductive traits

163 Throughout the breeding season, key reproductive traits related to reproductive activity, egg 164 survival, offspring provisioning and adult physiological and reproductive condition were 165 collected (Fig. 1). Terracotta pots were checked daily between 0900 and 1100 for the 166 presence of egg clutches. A digital photograph (Canon G12) was taken of each new clutch. A 167 sample of 10-20 eggs was then taken from the clutch and preserved in 6% formalin. Daily 168 photographs of each clutch were taken until the eggs hatched or there were no more eggs due 169 to mortality. Parents often eat the eggs if development is abnormal or if eggs are diseased. 170 Clutches were considered successful if any eggs survived to 6 - 8 days post-spawning. 171 Surviving egg clutches were hatched into 70L aquaria at the same pCO_2 and temperature 172 treatment as their parents. A sample of between 10-20 larvae was taken the morning after 173 hatching between 0730 and 0830, within 12 hours of hatching. Larvae were euthanized with 174 an overdose of clove oil before being preserved in 4% phosphate buffered formaldehyde. A

digital photograph (Canon G12) of each larva was taken on a 5mm grid in a horizontalposition within 3 days of sampling.

177

178 The number of eggs laid in each clutch and the number of eggs remaining at hatching was 179 counted from the digital image with the aid of ImageJ. The percentage of surviving eggs was 180 then determined from these two counts. Fertilization success was determined by counting the 181 number of unfertilized eggs in the initial clutch photograph. Unfertilized eggs were identified 182 by their white colouration, whereas fertilized eggs had an orange colouration. To determine 183 egg size the eggs sampled from each clutch were photographed (Canon G12) while placed 184 horizontally on a 5mm grid so that the longest axis was visible. The image was viewed on a 185 computer screen and ImageJ was used to trace the outside of 5 eggs from each sample and the average egg area (mm²) was determined. Reproductive output for each clutch was estimated 186 187 by multiplying the total number of eggs by the average egg area for that clutch providing a relative estimate of investment for each clutch (mm²). Hatchling standard length (SL) was 188 189 measured to the nearest 0.1mm from a photograph using ImageJ, by drawing a line from the 190 tip of the mouth to the beginning of the tail. Yolk area was determined to the nearest 0.1mm² 191 by tracing the yolk sac in ImageJ from photographs viewed on a computer screen.

192

193 Data collection of adult physiological and reproductive traits

Adults were euthanized at the end of the breeding season to examine the effects of increased

195 temperature and pCO_2 on body condition, liver condition, oocyte production and gonadal

- 196 steroidogenesis (17β-estradiol, E₂). Each fish was weighed (wet weight W, nearest 0.01g) and
- 197 measured (standard length SL, to the nearest 0.01mm). Fulton's K body condition factor
- 198 (body condition) was then calculated using the formula $K=100^{*}(W/SL^{3})$, where W is wet
- 199 weight in grams and SL is standard length in centimeters. To maintain genetic material and

- allow for potential genetic analysis, liver and gonads were dissected and snap frozen in liquid
- 201 nitrogen. After freezing they were weighed to the nearest 0.0001g and then fixed in 4%
- 202 phosphate buffered formaldehyde for several days before storage in 100% ethanol.
- 203 Hepatosomatic index and gonadosomatic index were determined by the formula HSI/GSI=
- 204 (liver weight or gonad weight (g)/fish weight (g))*100.
- 205
- 206 *Plasma 17β-estradiol quantification*

207 Blood samples were taken from females prior to euthanasia to estimate 17β -estradiol (E₂)

208 concentrations. 17β-estradiol was chosen due to its role in vitellogenesis and oocyte

- 209 maturation in female fish. Thus, changes in E₂ concentration could result in changes yolk
- 210 provisioning, egg size, the number of eggs per clutch, the number of clutches, hatchling
- survival and hatchling length (Lubzens et al., 2010; Yaron & Levavi-Sivan, 2011). Changes
- in E₂ concentration can therefore provide a direct endocrine pathway to any changes in
- 213 reproductive output.
- 214

Blood samples were taken towards the end of the breeding season to allow for a direct comparison of plasma hormone concentrations to the reproductive activity of the individual female. Each fish was caught using a hand-net and a cranial concussion was delivered to render the fish unconscious. Blood samples were taken immediately via caudal puncture using pre-heparinized syringes and centrifuged immediately after collection. Plasma was aspirated from the top of the sample and snap frozen in liquid nitrogen before being stored at -80°C until steroid measurement.

222

To determine the E_2 concentration in the plasma samples, a known volume of plasma was combined with a solvent (containing 1:1 ethyl acetate and N-hexane) in a glass test tube (1:4

plasma:solvent). The mixture was vortexed twice for 10 seconds, with the layers allowed to separate between each round. The upper clear liquid was then transferred into a clean glass

test vial, and the extraction steps were repeated three times for each sample. The residual

solvent mix was evaporated by heating to 37°C under a gentle stream of nitrogen. The

extracted samples were then capped and frozen at -20°C until analysis.

230

231 Plasma concentrations of E₂ were measured using enzyme immunoassay (EIA) kits obtained

from Cayman Chemicals, Sapphire Bioscience (No 582251), and validated for A. melanopus

233 (S. Metcalfe, unpublished data). The manufacturers instructions were followed, except that

extracted plasma samples were measured in duplicate. Absorbance at 405nm was measured

235 on a spectophotomere (Thermo Multiskan Ascent). Average extraction efficiency was

236 114±5.1% (standard error) (n=4), intra-assay variability was 5.3% (n=7) and 13.7% (n=6),

and inter-assay variability was 9.6% (n=13).

238

239 Liver and gonad samples

240 Fixed ovaries were embedded in histoparaffin and 5-µm sections were taken at 3 points along 241 the longest axis. Sections were mounted on a glass slide and stained with Mayer's alum 242 haematoxylin and Young's eosin-erythrosine. To determine the reproductive status of 243 individuals, a transect was run along each representative section and the type of sex cell 244 under 100-graticules marked on an eyepiece micrometer was recorded at a 10 times 245 magnification. Female cells were categorized into: oogonia (Stage 1), perinucleolus (Stage 246 2), cortical alveolus (Stage 3), early vitellogenic oocytes (Stage 4) and late vitellogenic 247 oocytes (Stage 5), following Genten et al. (2009). The relative abundance of each cell type in 248 each section was calculated.

249

250 Aquarium systems and seawater analysis

251 The pCO_2 in each system was controlled by an AquaMedic AT-controller that dosed a 3000L

sump with CO_2 to maintain the pH at the appropriate level for the desired pCO_2 . The control

- temperature (+0.0°C) was maintained by circulating seawater through a SolarWise
- 254 heater/chiller on each system. The +0.0°C temperature seawater was either delivered directly
- to the aquaria, or was sent through Toyesi inline 2.5kW heaters to raise the temperature
- 256 +1.5°C or +3.0°C, prior to delivery to the aquaria.
- 257
- 258 pH_{NBS} (Hach HQ40d) and temperature (Comark C26 thermometer) were recorded daily from
- 259 replicate aquariums for each treatment. Total alkalinity was estimated weekly by Gran

260 Titration (Metrohm 888 Titrando titrator) and validated against certified reference material

261 (A.G. Dickson Scipss Institute of Oceanography). Salinity (Hach HQ15d) was measured

- 262 weekly. The Aqua Medic pH set points were adjusted as needed to maintain the desired pCO_2
- on each system.
- 264
- 265 *p*CO₂ was calculated in CO2SYS (<u>http://cdiac.ornl.gov/oceans/co2rprt.html</u>) using the daily
- 266 pH_{NBS} and temperature (°C) readings and the weekly total alkalinity and salinity
- 267 measurements. As temperature affects seawater pCO_2 , CO_2 levels were not exactly the same
- among temperature treatments within each CO₂ treatment group. Nevertheless, CO₂
- treatments remained well separated and for simplicity, the average CO₂ levels are reported
- 270 (417µatm, 644µatm and 1134µatm) and are referred to as control, moderate or high CO₂

treatments.

272

273 Data analysis

ANCOVA was used to compare the number of clutches produced per pair in each treatment
group. The number of clutches produced was the dependent variable, CO₂ and temperature
treatments were the fixed variables, and female weight the covariate.

277

278 Linear mixed effects models (LME) (Pinheiro & Bates, 2000) were used to analyze the 279 reproductive characteristics: interclutch interval, average number of eggs laid per clutch, 280 average fertilization success, average egg area and reproductive output per clutch. As the 281 experiment aimed to determine the interactive effect of increased temperature and CO_2 on 282 reproduction, only clutches that were produced after experimental temperatures were attained 283 were included in the analysis. The simplest LME constructed included the reproductive 284 characteristic of interest as the dependent variable, CO₂ and temperature treatments as the 285 fixed variables, and female weight was included as a random variable because reproductive 286 traits can be strongly weight-dependent in fishes (Model A). Model B was constructed as 287 above, but also grouped the data according to the breeding pair. Model C added the step of 288 allowing for heterogeneity of variance within each pair, as pairs will have naturally 289 fluctuating reproductive effort. The model that best represented the data set was determined 290 by comparing the Akaike Information Criterion. Hatchling length and yolk area were also 291 analyzed using linear mixed effects models. These analyses were constructed in the same 292 order as described above (Model A, Model B and Model C). For hatchling length and yolk 293 area, data was grouped according to clutch ID and heterogeneity of variance was allowed to 294 occur within each clutch, as this was the level of replication.

295

296 The proportion of clutches that survived to hatching, and the proportion of eggs within each

297 clutch that survived to hatching, were analysed with a penalized quasi-likelihood general

298 linear mixed model (Splus Mass library). Temperature and CO₂ treatment were fixed

variables, and breeding pair was a random variable in each model. For clutch survival, the proportion of the number of clutches that survived was weighted against the number of clutches that were produced by each pair. All mixed effects models were constructed and compared in Splus.

303

304 Fixed factor ANOVA (Type III) was used to compare Fulton's K, hepatosomatic index,

305 gonadosomatic index and plasma E₂ concentration of females. The physiological trait was the

306 dependent variable and CO₂ and temperature treatment were the fixed variables. Where a

307 significant difference was detected a Fisher's LSD test was used to determine which

- 308 treatment groups were significantly different.
- 309

310 The relationship between plasma E₂ concentration and the stage of gonadal development in

311 individual fish was examined using factor analysis (Manly 1994, Kroon *et al.*, 2003). The

312 data was transformed using a varimax raw rotation to differentiate the original variables by

313 extracted factor. Initial analysis identified two factors, and these two factors were used as the

 $314 \qquad independent \ variables \ in \ a \ multiple \ regression \ analysis, \ where \ the \ plasma \ concentration \ of \ E_2$

315 was the dependent variable.

316

317 **Results**

318 Reproductive characteristics

319 Reproduction in all treatment groups began in early October 2011, a month prior to summer

320 breeding temperatures being achieved. Reproduction continued throughout the breeding

321 season at +0.0°C for all CO₂ treatments (Fig. 2a). In contrast, reproduction in the +3.0°C

322 treatment groups, irrespective of CO_2 level, and in moderate $CO_2 + 1.5$ °C, effectively ceased

323 within a month of experimental temperatures being attained (Fig. 2b,c). Reproduction in all

324	other groups continued from late September 2011 through to mid April 2012 with no obvious
325	peaks in reproductive activity (Fig. 2a,b). An unequal number of pairs reproduced in the
326	treatment groups, with 5 pairs reproducing in the moderate +0.0°C but only 2 reproducing in
327	the moderate +3.0°C (Fig 2). The total number of egg clutches produced was greatest at
328	+0.0°C (N=152) and declined markedly with increasing temperature (N=53 at +1.5°C and
329	N=8 at +3.0°C) (Fig. 2). At +0.0°C the moderate and high CO ₂ groups produced more
330	clutches in total compared to the control CO_2 group (n= 55, 58 and 39 respectively), but this
331	trend was not apparent at higher temperatures. Temperature appeared to have a stronger
332	effect on the moderate CO ₂ breeding pairs, as the moderate +1.5°C did not reproduce
333	successfully once temperatures were attained, whereas, the control and high CO_2 at the same
334	temperature continued to reproduce successfully (Fig. 2b).
335	
336	The average number of clutches produced per pair declined with increasing temperature
337	(Table 2, Fig. 3a). At +3.0°C there was a decline of between 78% to 87% in the average
338	number of clutches produced per pair in all CO ₂ treatments compared to the respective
339	+0.0°C clutches produced per pair (Fig. 3a). In contrast, elevated CO ₂ had no significant
340	effect on the number of clutches produced per pair (Table 2, Fig. 3a). There was no effect of
341	female weight on the average number of clutches produced per pair (Table 2.)
342	
343	Unsurprisingly, given the differences in the number of clutches produced among treatments,
344	both temperature and elevated CO_2 significantly increased the interclutch interval (Table 3).
345	This was most marked in the moderate CO ₂ group, where interclutch interval increased from
346	15±0.8 (SE) days at +0.0°C to 96.5±115 (SE) days at +3.0°C. A similar, though less marked
347	effect, was seen in the control CO_2 group where interclutch interval increased from 18±0.8 at
348	+0.0°C to 35±4 at +3.0°C. In contrast, interclutch interval decreased from 17±2 days at

+0.0°C to 14±1 days at +1.5°C in the high CO₂ group before increasing to 19±6 days at
+3.0°C.



- 374 similar pattern to the number of eggs produced, with temperature having a negative effect on
- 375 output, however only the control +3.0°C was different from the control +0.0°C (Table 3, Fig.
- 376 3d). There was no effect of CO_2 on reproductive output (Table 3).
- 377
- 378 More than 85% of the clutches produced in +0.0°C survived to hatching regardless of the
- 379 CO₂ treatment. However, the number of clutches that survived to hatching markedly declined
- 380 with increasing temperature (Table 4, Fig. 3e). No clutches survived to hatching at moderate
- 381 or high +3.0°C and only one clutch survived to hatching in control +3.0°C (Fig. 3e). Egg
- 382 survival to hatching was quite low, the highest average survival being 47% in the control CO₂
- $+0.0^{\circ}$ C. Egg survival decreased with increasing CO₂ down to 29% survival in high $+0.0^{\circ}$ C
- 384 (Table 4, Fig. 3f). In addition, temperature increase also decreased egg survival to 7% in the
- control and high $CO_2 + 1.5^{\circ}C$ and no survival in the moderate $CO_2 + 1.5^{\circ}C$ or the $+3.0^{\circ}C$
- 386 groups (Table 4, Fig. 3f). There was insufficient reproduction in the moderate 1.5°C and the
- 387 +3.0°C treatment groups for significant differences to be detected.
- 388
- 389 Offspring characteristics
- 390 No clutches survived to hatchling in moderate $CO_2 + 1.5$ °C or the +3.0 °C groups,
- 391 consequently only the +0.0°C and control and high +1.5°C were analyzed. Both temperature
- and CO₂ had a significant effect on hatchling length (Table 5, Fig. 4a). Newly hatched larvae
- in the high +0.0°C and in the control +1.5°C groups were significantly shorter than the
- 394 control +0.0°C (Fig. 4a). Elevated CO₂ but not temperature, had a significant negative impact
- 395 on yolk area, with both the moderate and high +0.0°C treatment group larvae having smaller
- 396 yolk reserves compared to control +0.0°C (Table 5, Fig. 4b).
- 397
- 398 Adult body and reproductive condition

399	Neither Fulton's K body condition factor or hepatosomatic index (HSI) were significantly
400	affected by either temperature or CO ₂ , for reproductive females (Table 6, Fig. 5a,b). Fulton's
401	K and HSI levels were generally high and only reduced at the most extreme treatment, high
402	CO ₂ +3.0°C. Temperature significantly effected gonadosomatic index (GSI) and there was an
403	interaction between temperature and CO ₂ treatment (Table 6). At +0.0°C, moderate CO ₂ was
404	significantly different from all other treatment groups (Fig. 5c). GSI was unaffected by
405	temperature in control CO ₂ . At moderate CO ₂ , GSI was highest at +0.0°C and declined at the
406	higher temperature. In contrast, GSI was highest at +1.5°C before decreasing at +3.0°C in the
407	high CO ₂ treatment (Fig. 5c).
408	
409	Plasma E_2 concentrations
410	Plasma E ₂ concentrations were significantly negatively affected by increasing temperature
411	(Table 6). Concentrations in the +3.0°C were significantly lower than the +0.0°C E_2
412	concentrations (Fig. 5d). Factor analysis identified two main trends in the variance of plasma
413	E ₂ concentrations and gonadal development. First, that there was a negative relationship
414	between the presence of stage 2 oocytes and a positive relationship between the presence of
415	stage 4 and 5 oocytes with plasma E_2 concentration (Table 7) accounting for nearly 50% of
416	the variation. The second factor, accounting for 25% of the variation, was influenced by the
417	presence of stage 3 oocytes (Table 7). Multiple regression analysis showed significant
418	relationship between factor 1 and plasma E_2 concentration ($F_{2,47}=25.63$, p<0.0001) which

419 included a significant relationship to factor 1 ($F_{1,47}$ =7.11, p<0.000001) but not factor 2

420 (F_{1,47}=-0.88, p=0.38).

421

422 Discussion

423	Higher temperatures and elevated CO ₂ can act additively or synergistically to reduce
424	individual performance (Pörtner & Farrell, 2008). Previous studies have found that
425	reproduction declines at elevated temperatures in a range of marine fishes (Donelson et al.,
426	2010; Hilder & Pankhurst 2003; Van Der Kraak & Pankhurst 1997). Similarly, increased CO ₂
427	is predicted to increase the energy required for maintaining homeostasis, and therefore reduce
428	the amount of energy available for reproduction (Ishimatsu et al., 2008, Melzner et al., 2009;
429	Pörtner, 2012). Despite this prediction, recent studies have found that exposure to increased
430	CO ₂ does not, on its own, cause reproduction to decline in fish (Sundin et al., 2012; Frommel
431	et al., 2010; Miller et al., 2013; Schade et al., 2014). This study is the first to examine the
432	interaction between temperature and CO ₂ on reproduction in reef fish for an entire
433	reproductive season. We found that the interaction between CO ₂ and temperature was
434	complex, but that overall, elevated temperature had a much greater effect on reproduction
435	than did projected future CO_2 levels (Table 8). At control temperatures there was an apparent
436	decline in reproductive output and offspring quality with increasing CO ₂ . At +1.5°C above
437	current-day temperatures, breeding pairs in the moderate CO ₂ treatment didn't produce
438	successful clutches; however the control and high CO ₂ pairs at this temperature continued to
439	reproduce, though at a reduced rate compared to the same CO ₂ treatments at control
440	temperatures. By far the most obvious result from this study, was the complete cessation of
441	reproduction at $+3.0^{\circ}$ C above current-day summer average temperature, irrespective of CO ₂
442	level.

444 *Reproductive and offspring characteristics: CO*₂

445 In a previous experiment, Miller et al. (2013) observed an increase in reproduction at

446 elevated CO₂ levels similar to those used in this experiment. A similar increase in

reproduction has since been documented in the Three-spine stickleback (Schade *et al.*, 2014)

448 and elevated CO₂ resulted in increased levels of reproductive hormones in Atlantic cod

449 (Preus-Olsen et al., 2014). Together, these results suggest that stimulation of reproduction by

450 elevated CO_2 could occur in a variety of fish species from a range of families.

451

452 In this study, however, reproduction in the high CO₂ group was not significantly increased 453 compared to the control or moderate groups (at control temperatures matching Miller et al. 454 2013). In this instance, the control and moderate groups doubled their reproductive activity 455 compared to Miller *et al.* (2013), while the high group maintained reproductive levels in 456 comparison to the previous study. The high CO_2 group did produce a similar number of eggs 457 per clutch, and more clutches over the season, compared to Miller et al. (2013). The reason 458 for the increase in reproductive performance of the control and moderate CO_2 groups 459 compared to reproductive performance of control and moderate pairs in Miller et al. (2013) is 460 unclear, but may be related to difference in the time required to acclimate to laboratory 461 conditions. Breeding of wild-caught fish can improve with time in captivity, which could 462 explain why the control and moderate group performed better in their second year of 463 captivity (this study) compared with the earlier study (Miller et al., 2013). Behavioural 464 studies show that reef fish exposed to $CO_2 > 700\mu$ atm tend to be bolder and more active 465 (reviewed Munday et al., 2012) which may compensate for the stress response to captivity 466 (Pankhurst & Van Der Kraak, 1997), leading to greater breeding in the first year in this 467 group. Whatever the mechanism, our results suggest that examining just one reproductive 468 season may not provide a full picture of the effect of elevated CO_2 on reproduction. 469 470 Unlike our previous study (Miller *et al.*, 2013) we also detected significant negative impacts 471 on reproduction, with clear declines in egg survival and yolk provisioning. A similar decline

472 in embryonic survival has recently been seen in a temperate goby (Forsgren *et al.*, 2013), but

473 was not detected in a closely related species, *Amphiprion percula*, (Munday *et al.*, 2009b).

474 The decline detected here equates to the high $CO_2 + 0.0^{\circ}C$ group having less than half the

475 number of surviving eggs per clutch (~250 eggs) compared to the control group (~540 eggs).

- 476 A decline in reproductive output of this magnitude, if it occurs in wild populations, could
- 477 potentially have a significant effect on population replenishment.
- 478

479 Further to the decline in embryonic survival, the larvae that were produced under high CO₂ 480 were both shorter and had less yolk compared to the control and moderate CO_2 larvae. A 481 reduction in yolk area was also detected in *A. percula* larvae reared at similar CO₂ levels 482 (Munday et al., 2009b). Yolk reserves provide the energy for growth until the larvae are able 483 to feed, therefore a reduction in yolk provisioning could lead to reduced somatic growth at 484 least in the early larval stage. Yolk reserve is also a good indicator for future growth and 485 performance (Hoey & McCormick, 2004; Grorud-Colvert & Sponaugle, 2006). In addition, 486 the high CO₂ offspring were significantly shorter at hatching compared to control offspring. 487 Hatchling length is a key fitness-related trait (Miller *et al.*, 1998). Reductions in both yolk 488 reserve and hatchling length could reduce juvenile performance, potentially increasing 489 mortality.

490

491 Reproductive characteristics: Temperature and interaction

492 The most obvious result from our data was the negative impact of increasing temperature on

493 every reproductive characteristic investigated, except fertilization success, regardless of CO₂

- 494 level. This was particularly obvious in the decline in number of eggs produced per clutch,
- 495 with an increase of +3.0°C reducing the egg output of the control group by 75%. Even more
- 496 startling was the decline in the number of eggs that survived to hatching. An increase of
- 497 + 1.5°C reduced survival to hatching from ~49% to ~7% in the control group. The same

498	temperature increase resulted in no surviving eggs in the moderate CO ₂ group, and at +3.0°C
499	there were no surviving eggs regardless of CO ₂ treatment. This trend of declining
500	reproduction with increasing temperature has been shown in a number of tropical and
501	temperate fishes (Donelson et al., 2010; Lansteiner & Kletzl, 2012; Warren et al., 2012) and
502	in other ectothermic animals (Snell 1986; Lee et al., 2003). Given this trend, there could be
503	serious declines in fish populations by 2100, through reduced reproduction, unless there is
504	sufficient scope for thermal acclimation or adaptation of reproduction over the next few
505	decades.
506	

No studies have yet examined the potential for genetic adaptation of reproduction in fishes to

507

508 ocean warming. However, one study has tested the potential for acclimation of reproduction 509 to projected future warming in a reef fish. Donelson et al. (2014) found that reproductive 510 traits in Acanthochromis polyacanthus were restored to control levels when fish complete 511 development and are reaed their entire life at +1.5°C, but there was no reproductive 512 acclimation when fish were reared at $+3.0^{\circ}$ C for their entire lives. Consequently, there appear 513 to be constraints on the potential for acclimation, at least for some reef fishes, particularly at 514 the higher temperatures (+3.0°C) that caused the greatest declines in reproduction in our 515 study. Whether there is potential for transgenerational acclimation of reproduction is 516 currently unknown. 517 518 As with elevated CO₂, we detected a significant negative effect of increased temperature on 519 hatchling length in the control CO_2 group, and a similar, though non-significant, trend in the

520 high CO₂ group. There may be a minimum size for hatchling length, similar to the minimum

521 or optimal length required for metamorphosis of juveniles in fish and other species

522 (Chambers & Leggett, 1987; Altwegg & Reyer, 2007). If so, the effect of elevated CO₂

treatment may have already reduced hatchling length close to the minimum viable length, to a point that increased temperature did not have a further significant impact. This hypothesis is supported by the declines in yolk provisioning that occurred in $+1.5^{\circ}$ C control offspring, not being present in the $+1.5^{\circ}$ C high CO₂ offspring. A minimum energy requirement may be needed for embryo's to survive to hatching.

528

529 Potentially the most surprising result in this study was the cessation of reproduction in the 530 moderate $CO_2 + 1.5$ °C group. This was not due a delay in reproduction, as there was 531 reproduction in this group prior to experimental temperatures being attained. The fact that the 532 moderate CO₂ +0.0°C and the control +1.5°C groups both reproduced suggests that, on their 533 own, neither stressor is enough to restrict reproduction. However, when the two occur in 534 combination they cause sufficient stress on the organism, causing reproduction to cease. Interestingly, despite the clear decline in reproduction in the moderate $CO_2 + 1.5$ °C, there was 535 536 not a significant decline in plasma E₂ concentrations in this group. This suggests that changes 537 in concentrations of this particular sex steroid are not responsible for the decline in 538 reproduction at moderate CO_2 +1.5°C. One possible explanation for the cessation of 539 reproduction in this group, but not the high $CO_2 + 1.5$ °C, is that the moderate CO_2 level lies 540 below the threshold at which physiological acclimation occurs. In Miller et al. (2013) we 541 suggested that the increased reproduction in the high CO_2 group could be a hormetic 542 response. It is possible that at the high CO_2 level, a change, caused by the increased CO_2 543 occurs that "switches on" reproduction. This switch could explain why the high CO₂, but not 544 the moderate CO_2 group reproduced at +1.5°C. 545 546 There is a CO_2 threshold at which the behavioural impacts of elevated CO_2 begin to occur,

547 somewhere between 600 and 700µatm in most reef fishes studied to date (Munday *et al.*,

548 2010). It is possible that whatever causes the change in behaviour (hypothesized to be 549 disrupted neural activity, Nilsson et al., 2012; Hamilton et al., 2014) could also cause the 550 stimulation of reproduction. At a more practical level, the cessation of reproduction at 551 projected mid-century CO_2 and temperature levels is quite disturbing. The synergistic effects 552 of these stressors could result in reproductive failures for tropical fishes within the next 40 553 years potentially impacting on commercial and non-commercial species alike. While there is 554 evidence for transgenerational acclimation of life history-traits to ocean acidification (Miller 555 et al., 2012), and the at least potential for reproductive acclimation to moderate warming 556 $(+1.5^{\circ}C)$ (Donelson *et al.*, 2014), as yet, there is no evidence for acclimation to both these 557 stressors in combination.

558

559 Hormonal and physiological impacts

560 The results suggest that reproductive, but not physiological condition, of females caused the 561 changes in reproductive and offspring characteristics. First, elevated temperature resulted ina 562 decrease in E_2 concentrations. Moreover, there was a strong correlation between E_2 563 concentration and the average number of clutches produced per pair. In contrast, there was 564 little effect of elevated CO₂ or temperature on the physiological condition of either 565 reproductive or non-reproductive females. Both Fulton's K body condition index and 566 hepatosomatic index were high for all treatment groups, indicating no decline in body 567 condition or energy stores. This suggests that the dramatic declines in reproductive 568 outputobserved at higher temperature were not due to energetic constraints in the females. 569 Previous studies have shown that under elevated temperatures the enzyme CYP19 aromatase, 570 which catalyses the irreversible conversion of testosterone into E_2 in inhibited (Watts *et al.*, 571 2004). Inhibition of CYP19 aromatase will result in a decline in E_2 synthesis, and a 572 subsequent reduction in vitellogenesis and oocyte maturation (Piferrer & Blázquez, 2005;

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574 E₂ concetrations, likely through inhibition of CYP19 aromatase may be the cause for

575 decrease reproduction at higher temperature in *A. melanopus*.

576

577 Conclusions

578 This is the first time that the potential interactive effects of projected future CO_2 and

temperature conditions have been tested in regards to fish reproduction. As with similar

580 studies conducted on invertebrate reproduction (Chua *et al.*, 2013; Byrne *et al.*, 2009), we

581 found that temperature had a much stronger impact that CO₂. Our data, similar to other

582 studies on tropical fish, showed that there was complete reproductive failure at +3.0°C above

583 the current-day average temperature. Given that sea surface temperatures within the tropics

are projected to rise up to +3.0°C (Ganachaud *et al.*, 2011) by 2100, there could be

585 significant consequences for reproduction in tropical fish populations. Nevertheless, we did

586 detect interactions between temperature and CO₂ at the combined moderate levels. Previous

587 studies have shown that when temperature has increased +1.5°C that reproduction is reduced,

as we saw in the control CO_2 group. Yet when the extra stress of increased CO_2 was added,

and without any compensatory mechanisms, there appears to be a major reproductive failure.

590 Our data suggest that, without reproductive acclimation or adaptation, there could be

reproductive failure for this species as early as the middle of the century. These results

reinforce the importance of examining how multiple stressors will interact, so that accurate

593 climate change predictions can be made.

594

595 Our results also show that, for this species, there is no direct correlation between

596 physiological condition and reproductive response and that one will not necessarily predict

the other in future warmer and more acid conditions. Reproduction involves a complex series

- 598 of interactions between environmental conditions and hormonal pathways. Further research
- 599 will be required to determine the mechanisms responsible for declining reproduction at
- 600 higher temperature, but it will likely involve thermal sensitivity of hormonal pathways. As
- 601 yet there is no easy way to predict how reproduction will respond to climate change scenarios
- 602 other than to experimentally test the population.

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612	
613	References
614	Albright R, Mason B (2013) Projected near-future levels of temperature and pCO ₂ reduce
615	coral fertilization success. Plos One, doi:10.1371/journal.pone.0056468.
616	
617	Allen GR (1975) The anemone fishes. Their classification and biology. Second Edition.
618	T.F.H Publications, Inc. Neptune City, New Jersey.
619	
620	Altwegg R, Reyer H-U (2007) Patterns of natural selection on size at metamorphosis in water
621	frogs. Evolution, 57, 872-882.
622	
623	Bay LK, Crozier RH, Caley MJ (2006) The relationship between population genetic structure
624	and pelagic larval duration in coral reef fishes on the Great Barrier Reef. Marine Biology,
625	149 , 1247-1256.
626	

627	Brauner CJ, Baker DW	(2009)) Patterns of acid-base regulation during exposure	to
		· · · ·		

- 628 hypercarbia in fishes. In Cardio-Respiratory Control in Vertebrates (eds Glass ML, Woods
- 629 SC), pp43-63. Springer Berlin, Germany.
- 630
- Brown GP, Shine R (2007) Rain, prey and predators: climatically driven shifts in frog
- abundance modify reproductive allometry in a tropical snake. *Oecologia*, **154**, 361-368.

- Byrne M, Ho M, Selvakumaraswamy P, Nguyen HD, Dwojanyn SA, Davis AR (2009)
- 635 Temperature, but not pH, compromises sea urchin fertilization and early development under
- 636 near future climate change scenarios. *Proceedings of the Royal Society B: Biological*
- 637 Sciences, 276, 1883-1888.
- 638
- 639 Byrne M, Soars NA, Ho MA, Wong E, McElroy D, Selvakumaraswamy P, Dworjanyn SA,
- 640 Davis AR (2010) Fertilization in a suite of coastal marine invertebrates from SE Australia is
- robust to near-future ocean warming and acidification. *Marine Biology*, 157, 2061-2069.

642

- 643 Bunn SE, Arthington AH (2002) Basic principles and ecological consequences of altered
- flow regimes for aquatic biodiversity. *Environmental Management*, **30**, 492-507.
- 645
- 646 Chambers RC, Legget WC (1987) Size and age at metamorphosis in marine fishes: an
- 647 analysis of laboratory-reared winter flouder (*Pseudopleutonectes americanus*) with a review
- 648 of variation in other species. Canadian Journal of Fisheries and Aquatic Sciences, 44, 1936-

649 1947.

651	Chua CM, Leggat	W, Mova A	Baird AH	(2013) Ten	nperature affects	the early	life history
			,	() -			

stages of corals more than near future ocean acidification. *Marine Ecology Progress Series*,
475, 85-92.

654

655 Clutton-Brock TH (1984) Reproductive effort and terminal investment. *The American*656 *Naturalist*, **123**, 219-229.

657

658 Cohen-Rengifo M, García E, Hernández CA, Hernández JC, Clemente S (2013) Global

659 warming and ocean acidification affect fertilization and early development of the sea urchin

660 *Paracentrotus lividus. Cahiers de biologie marine*, **54**, 667-675.

661

662 Collins M, Knuttie R, Arblaster J, Dufresne J-L, Fichefet T, Friedlingstein P, Gao X,

663 Gutowski WJ, Johns T, Krinner G, Shongwe M, Tebaldi C, Weaver AJ, Wehner M. 2013:

664 Long-term Climate Change: Projects, Commitments and Irreversability. In: *Climate Change*

665 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment

666 Report of the Intergovernmental Panel on Climte Change [Stocker TF, Qin D, Plattner G-K,

667 Tignor M, Allen SK, Boschung J, Nauels A, Xia Y, Bex V and Midgley PM (eds)].

668 Cambridge University Press, Cambridge United Kingdom, and New York NY, USA.

669

670 Costantini D, Metcalfe NB, Monaghan P (2010) Ecological processes in a hermetic

671 framework. *Ecology Letters*, **13**, 1435-1447.

672

673 Dawson A, King VM, Bentley GE, Ball GF (2001) Photoperiodic control of seasonality in

674 birds. Journal of Biological Rhythms, 16, 365-380.

675

676	Donelson JM, Munday	v PL, McCormick MI, I	Pankhurst NW, Pankhurst PM (2010) Effects	; of
					-

- 677 elevated water temperature and food availability on the reproductive performance of a coral
- 678 reef fish. *Marine Ecology Progress Series*, **401**, 233-243.
- 679
- 680 Donelson JM, Munday PL, McCormick MI, Pitcher CR (2011) Rapid transgenerational
- acclimation of a tropical reef fish to climate change. *Nature Climate Change*, **2**, 30-32.
- 682
- 683 Donelson JM, McCormick, MI, Booth DJ, Munday PL (2014) Reproductive acclimation to

684 increased water temperature in a tropical reef fish. *PlosONE*, 9, e97223.

- 685
- 686 Doney SC, Fabry VJ, Feely RA, Kleypas JA (2009) Ocean acidification: The Other CO₂
- 687 problem. *Annual Review of Marine Science*, **1**, 169-192.
- 688
- Donnelly MA, Guyer C (1994) Patterns of reproduction and habitat use in an assemblage of
- 690 Neotripcal hylid frogs. *Oecologia*, **98**, 291-302.
- 691
- 692 Dorts J, Grenouillet G, Douxfils J, Mandiki SNM, Milla S, Silvestre F, Kestemont P (2011)
- 693 Evidence that elevated water temperature affects the reproductive physiology of the European
- 694 bullhead Cottus gobio. Fish Physiology and Biochemistry, **38**, 389-399.
- 695
- 696 Drew J, Allen GR, Kaufman L, Barber PH (2008) Endemism and regional colour and genetic
- difference in five putatively cosmopolitan reef fishes. *Conservation Biology*, 22, 965-975.
- 698
- 699 Duston J, Bromage N (1986) Photoperiodic mechanisms and rhythms of reproduction in the
- female rainbow trout. *Fish Physiology and Biochemistry*, **2**, 35-51.

702	Esbaugh AJ, Heuer R, Grossel M (2012) Impacts of ocean acidification on respiratory gas
703	exchange in a marine teleost, Ospanus beta. Journal of Comparative Physiology B, 182, 921-
704	934.
705	
706	Fosgren E, Dupont S, Jutfelt F, Amundsen T (2013) Elevated CO ₂ affects embryonic
707	development and larval phototasix in a temperature marine fish. Ecology and Evolution,
708	doi:10.1002/ece3.709.
709	
710	Frommel AY, Stiebens V, Clemmesen C, Havenhand J (2010) Effect of ocean acidification
711	on marine fish sperm (Baltic cod: Gadus morhua). Biogeosciences Discussions, 7, 5859-
712	5872.
713	
714	Ganachaud AS, Gupta AS, Orr JC, Wijffels SE, Ridgway KR, Memer MA, Maes C,
715	Steinberg CR, Tribollet AD, Qiu B, Kruger JC (2011) Observed and expected changes to the
716	tropical Pacific Ocean. In: Vulnerability of Tropical Pacific Fisheries and Aquaculture to
717	Climate Change. (ed Bell JD, Johnson JE, Hobday AJ). Secretariat of the Pacific
718	Community, Noumea, New Caledonia.
719	
720	Genten F, Terwinghe E, Danguy A (2009) Atlas of Fish Histology. BIOS Scientific
721	Publishers, UK.
722	
723	Grans A, Jutfelt F, Sandblom E, Jonsson E, Wiklander K, Seth H, Olsson C, Dupont S,
724	Ortega-Martinez O, Einarsdottir I, Bjornsson BT, Sundell K, Axelsson M (2014) Aerobic

- scope fails to explain the detrimental effects on growth resulting from warming and elevated
- 726 CO₂ in Atlantic halibut. *Journal of Experimental Biology*, **217**, 711-717.
- 727
- 728 Grorud-Colvert K, Sponaugle S (2006) Influence of condition on behaviour and survival
- potential of a newly settled coral reef fish, the bluehead wrasse *Thalassoma bifasciatum*.
- 730 Marine Ecology Progress Series, 327, 279-288.
- 731
- 732 Hamilton TJ, Holcombe A, Treguerres M (2014) CO₂-induced ocean acidification increased
- anxiety in Rockfish via alteration of GABAA receptor functioning. Proceedings of the Royal
- 734 *society B: Biological Science*, **281**, doi:10.1098/rspb.2013.2509.
- 735
- Hau M, Wikelski M, Gwinner J, Gwinner E (2004) Timing of reproduction in a Darwin's
- finch: temporal opportunism under spatrial constraints. *Oikos*, **106**, 489-500.
- 738
- Hilder ML, Pankhurst NW (2003) Evidence that temperature change cues reproductive
- 740 development in the spiny damselfish, Acanthochromis polyacanthus, Environmental Biology
- 741 *of Fishes*, **66**, 187-196.
- 742
- 743 Hoegh-Guldber O, Mumby PJ, Hootn AJ, Steneck RS, Greenfield P, Gomez E, Harvell CD,
- 744 Sale PF, Edwards AJ, Caldeira K, Knowlton N, Eakin CM, Iglesias-Prieto R, Muthiga N,
- 745 Bradbury RH, Dubi A, Hatziolos ME (2007) Coral reefs under rapid climate change and
- 746 ocean acidification. *Science*, **318**, 1737-1742.

- 748 Hoey AS, McCormick MI (2004) Selective predation for low body condition at the larval-
- juvenile transition of a coral reef fish. *Oecologia*, **139**, 23-29.

751	Inaba K, Dreanno C, Cosson J (2003) Control of flatfish sperm motility by CO ₂ and carbonic
752	anhydrase. Cell Motility and the Cytoskeleton, 55, 174-187.
753	
754	Ishimatsu A, Hayashi M, Kikkawa T (2008) Fishes in a high-CO ₂ , acidified oceans. Marine
755	Ecology Progress Series, 373, 295-302.
756	
757	King HR, Pankhurst NW, Watts M, Pankhurst PM (2003) Effect of elevated summer
758	temperature on gonadal steroid production, vitellogenesis and egg quality in female Atlantic
759	salmon. Journal of Fish Biology, 63, 153-167.
760	
761	Kjesbu OS (1994) Time of start of spawning of Atlantic cod (Gadus morhua) females in
762	relation to vitellogenic oocyte diameter, temperature, fish length and condition. Journal of
763	<i>Fish Biology</i> , 45 , 719-735.
764	
765	Kroon FJ, Munday PL, Pankhurst NW (2003) Steroid hormone levels and bi-directional sex
766	change in Gobiodon histrio, Journal of Fish Biology, 62, 153-167.
767	
768	Lahnsteiner F, Kletzl M (2012) The effect of water temperature on gamete maturation and
769	gamete quality in the European grayling (Thymalus thymallus) based on experimental data
770	and on data from wild populations. Fish Physiology and Biochemistry, 38, 455-467.
771	
772	Lee H, Ban S, Ikefa T, Matshuishi T (2003) Effect of temperature on development, growth
773	and nonnaduction in the manine command Draude calcuus a numericat esticting food condition
	and reproduction in the marine copepod <i>Pseudocalanus newmani</i> at satiating food condition.
774	Journal of Plankton Research, 25, 261-271.

- T76 Lieske E, Myers R (1994) Collins Pocket Guide, Coral Reef Fishes, Indo-pacific &
- 777 *Caribbean including the Red Sea.* Collins Publishers, Australia.
- 778
- Lubzens E, Young G, Bobe J, Cerdà J (2010) Oogenesis in teleosts: How fish eggs are
- formed. General and Comparative Endocrinology, 165, 367-389.
- 781
- 782 Manly BFJ (1994) Multivariate statistical methods. Chapman & Hall, New York, New York,
- 783 USA.
- 784
- 785 Melzner F, Gobel S, Langenbuch M, Gutowska MA, Pörtner H-O, Lucassen M (2009)
- 786 Swimming performance in Atlantic cod (*Gadus morhua*) following long-term (4-12months)
- acclimation to elevated seawater *p*CO₂. *Aquatic Toxicology*, **92**, 30-37.
- 788
- 789 Michael SC (2008) Damselfishes and Anemonefishes: The Complete Illustrated Guide to
- 790 Their Identification, Behaviours and Captive Care. T.F.H. Publications, New Jersey, USA.
- 791
- 792 Miller GM, Watson S-A, McCormick MI, Munday PL (2013) Increased CO₂ stimulates
- reproduction in a coral reef fish. *Global Change Biology*, **19**, 3037-3045.
- 794
- 795 Miller TJ, Crowder LB, Rice JA, Marshall EA (1998) Larval size and recruitment
- 796 mechanisms in fishes toward a conceptual framework. Canadian Journal of Fisheries and
- 797 Aquatic Sciences, 45, 1657-1670.

799	Munday PL, Crawley NE, Nilsson GE (2009a) Interacting effects of ocean acidification on
800	the aerobic performance of coral reef fishes. <i>Marine Ecology Progress Series</i> , 388 , 235-242.
801	
802	Munday PL, Donelson JM, Dixson DL, Endo GGK (2009b) Effects of ocean acidification on
803	the early life history of a tropical marine fish. Proceedings of the Royal Society B, 276, 3275-
804	3283.
805	
806	Munday PL, Dixson DL, McCormick MI, Meekan M, Ferrari MCO, Chivers DP (2010)
807	Replenishment of fish populations is threatened by ocean acidification. Proceedings of the
808	National Academy of Sciences USA, 107, 12930-12934.
809	
810	Munday PL, McCormick MI, Nilsson GE (2012) Impact of global warming and rising CO ₂
811	levels on coral reef fishes: what hope for the future? The Journal of Experimental Biology,
812	215 , 3865-3873.
813	
814	Nilsson GE, Dixson DI, Domenici P, McCormick MI, Sorensen C, Watson S-A, Munday PL
815	(2012) Near-future carbon dioxide levels alter fish behaviour by interfering with
816	neurotransmitter function. Nature Climate Change, 2, 201-204.
817	
818	Nowicki JP, Miller GM, Munday PL (2012) Interactive effects of elevated temperature and
819	CO ₂ on foraging behaviour of juvenile coral reef fish. Journal of Experimental Marine
820	Biology and Ecology, 412, 46-51.
821	
822	Pankhurst NW, King HR (2010) Temperature and salmonid reproduction: implications for
823	aquaculture. Journal of Fish Biology, 76, 69-85.

825	Pankhurst NW, Munday PL (2011) Effects of climate change on fish reproduction and early
826	life history stages. Marine and Freshwater Research, 62, 1015-1026.
827	
828	Pankhurst NW, Porter MJR (2003) Cold and dark or warm and light: variations on the theme
829	of environmental control of reproduction. Fish Physiology and Biochemistry, 28, 385-389.
830	
831	Pankhurst NW, Purser GJ, Van der Kraak G, Thomas PM, Forteath GNR (1996) Effect of
832	holding temperature on ovulation, egg fertility, plasma levels of reproductive hormones and
833	in vitro ovarian steroidogenesis in the rainbow trout, Oncorhynchus mykiss. Aquaculture,
834	146, 277-290.
835	
836	Pankhurst NW, Van Der Kraak (1997) Effects of stress on reproduction and growth of fish,
837	In: Fish Stress and Health in Aquaculture. (eds: Iwama QK, Pickering AD, Sumpter JP,
838	Schreck CB). Cambridge University Press, Cambridge UK.
839	
840	Parker LM, Ross PM, O'Conner WA (2009) The effect of ocean acidification and
841	temperature on the fertilization and embryonic development of the Sydney rock oyster
842	Saccostrea glomerata, (Gould 1850). Global Change Biology, 15, 2123-2136.
843	
844	Parmesan C (2006) Ecological and evolutionary responses to recent climate change. Annual
845	Review of Ecology, Evolution and Systematics, 37, 637-669.
846	
847	Piferrer F, Blázquez M (2005) Aromatase distribution and regulation in fish. Fish Physiology
848	and Biochemistry, 31 , 215-226.

850	Pinheiro JC, Bates DM (2000) Mixed effects models in S and S-Plus. Springer-Verlag, New
851	York.

852

853	Poloczanska ES.	Babcock RC.	Butletr A.	Hobday A.	Hoegh-C	Guldberg O	, Kunz TJ,	Matear R.
					,0 -		,,	

- 854 Milton DA, Okey TA, Richardson AJ (2007) Climate change and Australian marine life.
- 855 Oceanography and Marine Biology: An Annual Review, 45, 407-478.

856

- 857 Pörtner HO (2012) Integrating climate-related stressor effects on marine organisms: unifying
- 858 principles linking molecule to ecosystem level changes. Marine Ecology Progress Series,
- **470**, 273-290.
- 860
- 861 Pörtner HO, Langenbuch M, Reipschlarger A (2004) Biological impact of elevated ocean
- 862 CO₂ concentrations: lessons from animal physiology and earth history. Journal of
- 863 *Oceanography*, **60**, 705-718.
- 864

```
865 Pörtner HO, Farrel AP (2008) Physiology and Climate Change. Science, 322, 690-692.
```

866

- 867 Pradelles P, Grassi J, Maclouf J (1985) Enzyme immunoassays of eicosanoids using
- acetylcholine esterase as label: an alternative to radioimmunassay. Analytical chemistry, 57,
- 869 1170-1173.

- 871 Preus-Olsen G, Olufsen MO, Pedersen SA, Letcher RJ, Arukwe A (2014) Effects of elevated
- 872 dissolved carbon dioxide and perfluoroctane sulfonic acid, given singly and in combination,

873 on steroidogenic and biotransformation pathways of Atlantic cod. Aquatic Toxicology,

doi:10.1016/j.aquatox.2014.06.017.

875

876	Ross PM, Parker L,	O'Conner WA	, Bailey EA ((2011) The	e impact of oc	cean acidification on
-----	--------------------	-------------	---------------	------------	----------------	-----------------------

- reproduction, early development and settlement of marine organisms. *Water*, **3**, 1005-1030.
- 878

879 Rummer JL, Couturies CS, Stecyk JAW, Gardiner NM, Kinch JP, Nilsson GE, Munday PL

880 (2014) Life on the edge: Thermal optima for aerobic scope of equatorial fishes are close to

current day temperatures. *Global Change Biology*, 20, 1055-1066.

882

883 Saino N, Romano M, Ambrosini R, Ferrari RP, Moller AP (2004) Timing of reproduction

and egg quality covary with temperature in the insectivorous Barn Swallow, *Hirundo rustica*.

885 *Functional Ecology*, **18**, 50-57.

886

887 Schade, FM, Clemmesen C, Wegner KM (2014 Within-and transgenerational effects of

888 ocean acidification on life history of marine three-spined stickleback (Gasterosteus

aculeatus). Marine Biology, doi:10.1007/s00227-014-2450-6.

890

- 891 Schlosser IJ (1982) Trophic structure, reproductive success, and growth rate of fishes in a
- 892 natural and modified headwater stream. Canadian Journal of Fisheries and Aquatic Sciences,
- **39**, 968-978.

894

895 Shama LNS, Strobel A, Mark FC, Wegner KM (2014) Transgenerational plasticity in marine

sticklebacks: maternal effects mediate impacts of a warming ocean. Functional Ecology,

897 doi:10.1111/1365-2435.12280.

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- 901
- 902 Sokolova IM, Frederick M, Bagwe R, Lannig G, Sukhotin AA (2012) Energy homeostasis as
- 903 an integrative tool for assessing limits of environmental stress tolerance in aquatic
- 904 invertebrates. Marine Environmental Research, 29, 1-15.
- 905
- 906 Somero GN (2002) Thermal physiology and vertical zonation of intertidal animals: Optime,

907 limits, and costs of living. *Integrative & Comparative Biology*, **42**, 780-789.

- 908
- 909 Stearns SC (1992) The Evolution of Life Histories. Oxford University Press, New York,
- 910 USA.
- 911
- Sundin J, Rosenqvist G, Berglund A (2012) Altered ocean pH impairs mating propensity in a
 pipefish. *Ethology*, **119**, 86-93.
- 914
- 915 Tewksbury JJ, Huey RB, Deutsch CA (2008) Putting the heat on tropical animals. *Science*,
- 916 **320**, 1296-1297.
- 917
- 918 Van Der Kraak G, Pankhurst NW (1997) Temperature effects on the reproductive
- 919 performance of fish. In: *Global Warming: Implications for Freshwater and Marine Fish.*
- 920 (eds. Wood CM, McDonald DG). Cambridge University Press, UK.

⁹⁰⁰ reproduction in *Brachionus plicatilis* (Rotifera). *Marine Biology*, **92**, 157-162.

- 922 Visser ME, Holleman LJM, Caro SP (2009) Temperature has a causal effect on avian timing
- 923 of reproduction. *Proceedings of the Royal Society B: Biological Sciences*, **276**, 2323-2331.

- 925 Warren DR, Robinson JM, Josephson DC, Sheldon DR, Kraft CE (2012) Elevated summer
- 926 temperatures delay spawning and reduce red construction for resident brook trout (Salvelinus
- 927 fontinalis). Global Change Biology, 18, 1804-1811.
- 928
- 929 Watts M, Pankhurst NW, King HR (2004) Maintenance of Atlantic Salmon (Salmo salar) at
- elevated temperature inhibits cytochrome P450 aromatase activity in isolated ovarian
- 931 follicles. *General and Comparative Endocrinology*, **135**, 381-390.
- 932
- 933 Yaron Z, Levavi-Sivan B (2011) Endocrine Regulation of Fish Reproduction. In:
- 934 Encyclopedia of Fish Physiology: From Genome to Environment (ed: Farrell AP) Academic
- 935 Press, San Diego, USA.
- 936
- 2012) Zeh JA, Bonilla MM, Su EJ, Padua MV, Anderson RV, Kaur D, Yang D, Zeh DW (2012)
- 938 Degrees of disruption: projected temperature increase has catastrophic consequences for
- reproduction in a tropical ectotherm. *Global Change Biology*, **18**, 1833-1842.
- 940

- 941 Table 1: Seawater parameters for adult *Amphiprion melanopus* held under control, moderate
- 942 or high CO₂ cross-factored with +0.0°C (28.5°C), +1.5°C (30.0°C) or +3.0°C (31.5°C) water
- 943 temperature treatments. Temperature, salinity, total alkalinity and pH_{NBS} were measured in

944	situ,	while	pCO_2	was	calculated	using	CO2SYS.	
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Treatment	Temp	Salinity	Total	pH_{NBS}	pCO ₂
	(°C)	(ppt)	alkalinity		(µatm)
			(µmol kg ⁻¹		
			SW)		
Control +0.0°C	28.4±0.01	33.32±0.12	2058±16	8.15±0.005	400±6
Control +1.5°C	29.8±0.02	33.32±0.12	2064±16	8.14±0.005	411±6
Control +3.0°C	31.4±0.02	33.32±0.12	2077±16	8.12±0.005	441±7
Moderate +0.0°C	28.5±0.01	32.7±0.12	2152±10	8.00±0.007	634±13
Moderate +1.5°C	30.1±0.01	32.7±0.12	2117±7	8.00±0.006	642±12
Moderate +3.0°C	31.5±0.01	32.7±0.12	2130±8	8.00±0.007	658±13
High +0.0°C	28.5±0.01	33.62±0.09	2168±7	7.81±0.008	1087±25
High +1.5°C	29.8±0.02	33.62±0.09	2167±7	7.79±0.008	1126±24
High +3.0°C	31.5±0.01	33.62±0.09	2169±7	7.78±0.008	1191±27

947 Table 2. ANCOVA (type III) results for the average number of clutches produced by each

adult pair kept at control, moderate or high CO₂ cross-factored with +0.0°C (28.5°C), +1.5°C

Treatment	DF	Sums of	Mean	F-value	p-value
		Squares	Squares		
Weight	1	1.252	1.252	0.095	0.7619
Temperature	2	207.932	103.9663	7.856	< 0.01*
CO ₂ :Temperature	4	16.702	5.176	0.316	0.8629
Residuals	18	238.1976	13.233		
Residuals	18	238.1976	13.233		

 $(30.0^{\circ}C)$ or $+3.0^{\circ}C$ $(31.5^{\circ}C)$ water temperature treatments.

Table 3. Linear mixed effects model tables for the reproductive characteristics from adults kept at control, moderate or high CO₂ cross-factored with +0.0°C (28.5°C), +1.5°C (30.0°C) or +3.0°C (31.5°C) water temperature treatments. Treatment groups that are significantly different to Control +0.0°C are marked by *.

<0.0001*
0.4905
<0.0001*
<0.001*
0.2371
0.3982
<0.0001*
0.8582
0.2989
0.2974
<0.05*
0.2566
0.1397
0.7034
0.8094
0.4416
0.5239
0.5139
0.6477

	~ `					
Fertilisation success	(Intercept)	90.160	3.791	157	23.783	<0.0001*
	Control +1.5°C	9.409	6.405	18	1.469	0.1591
	Control +3.0°C	7.322	8.223	18	0.890	0.3850
	Moderate +0.0°C	9.362	6.161	18	1.520	0.1460
	Moderate +1.5°C	-8.931	13.081	18	-0.682	0.5035
	Moderate +3.0°C	-55.415	12.019	18	-4.611	<0.001*
	High +0.0°C	4.527	3.791	18	1.194	0.2479
	High +1.5°C	-3.439	8.528	18	-0.403	0.6915
	High +3.0°C	-2.009	13.196	18	-0.152	0.8807
Egg Area	(Intercept)	2.563	0.061	1627	41.684	<0.0001*
	Female Weight	-0.014	0.002	174	-7.943	<0.0001*
	Control +1.5°C	-0.277	0.034	174	-8.194	<0.0001*
	Control +3.0°C	-0.794	0.076	174	-10.428	<0.0001*
	Moderate +0.0°C	-0.095	0.032	174	-3.011	<0.01*
	Moderate +1.5°C	-0.197	0.075	174	-2.621	<0.01*
	Moderate +3.0°C	0.338	0.117	174	2.897	<0.01*
	High +0.0°C	-0.128	0.036	174	-3.571	<0.001*
	High +1.5°C	0.018	0.043	174	0.413	0.6804
	High +3.0°C	0.184	0.157	174	1.252	0.2122
Reproductive output	(Intercept)	1161.624	741.883	157	1.566	0.1194
	Female Weight	39.682	20.922	17	1.897	0.0750
	Control +1.5°C	-701.157	413.854	17	-1.694	0.1085
	Control +3.0°C	-1462.128	589.445	17	-2.481	<0.05*
	Moderate +0.0°C	104.453	411.792	17	0.254	0.8028
	Moderate +1.5°C	64.732	802.208	17	0.081	0.9366

Moderate +3.0°C	-398.727	793.082	17	-0.503	0.6216
High +0.0°C	-405.597	452.430	17	-0.896	0.3825
High +1.5°C	412.802	547.022	17	0.755	0.4608
High +3.0°C	-49.727	887.772	17	-0.056	0.9560

959	Table 4. Quasi-likelihood general linear model results for the average proportion of clutches
960	that survived to hatching and average proportion of eggs that survived to hatching for adults
961	kept at control, moderate or high CO ₂ cross-factored with +0.0°C (28.5°C), +1.5°C (30.0°C)
962	or +3.0°C (31.5°C) water temperature treatments. Treatment averages that are significantly
963	different from Control +0.0°C are marked with an asterix (*).

Treatment	Value	SE	DF	t-value	p value
Successful clutches (%)					
Control +0.0°C	-0.211	0.2	64	-1.125	0.2646
Control +1.5°C	-2.220	0.4	19	-5.217	<0.0001*
Control +3.0°C	-1.706	0.8	19	-2.069	0.0525
Moderate +0.0°C	0.020	0.3	19	0.079	0.9378
Moderate +1.5°C	-24.661	116809.1	19	0.000	0.9998
Moderate +3.0°C	-24.54947	110522.5	19	0.000	0.9998
High +0.0°C	-0.444	0.3	19	-1.481	0.1549
High +1.5°C	0.348	0.6	19	0.555	0.585
High +3.0°C	-23.584	54075.0	19	0.000	0.999
Egg survival					
Control +0.0°C	-0.156	0.15	179	-1.050	0.2953
Control +1.5°C	-2.250	0.36	19	-6.230	<0.0001*
Control +3.0°C	-1.762	0.87	19	-2.018	0.0580
Moderate +0.0°C	-0.292	0.2	19	-1.451	0.1630
Moderate +1.5°C	-23.271	65787.90	19	0.000	0.9997
Moderate +3.0°C	-23.237	71691.82	19	0.000	0.9997
High +0.0°C	-0.659	0.23	19	-2.828	<0.05*
High +1.5°C	0.661	0.49	19	1.258	0.1904

High +3.0°C	-22.369	35076.46	19	0.000	0.995	
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966	Table 5. Linear mixed effects ANOVA results for physical characteristics (hatchling length
967	and yolk area) of offspring resulting from adults kept at control, moderate or high CO ₂ cross-
968	factored with +0.0°C (28.5°C) or +1.5°C (30.0°C) water temperature treatments. No clutches
969	survived to hatching in the Moderate +1.5°C or +3.0°C treatment groups. Significant effects
970	(p<0.05) are denoted by *.

Characteristic	Variable	Value	Standard	DF	t-value	p-value
			Error			
Hatchling length	(Intercept)	3.245	0.128	557	25.358	<0.0001*
	Female Weight	-0.000	0.004	64	-0.012	0.9901
	Control +1.5°C	-0.083	0.036	64	-2.322	<0.05*
	Moderate +0.0°C	0.047	0.057	64	0.820	0.4152
	High +0.0°C	-0.111	0.055	64	-2.004	<0.05*
Yolk area	(Intercept)	0.486	0.036	557	13.685	<0.0001*
	Female Weight	-0.002	0.001	64	-1.903	0.0615
	Control +1.5°C	-0.006	0.010	64	-0.623	0.5356
	Moderate +0.0°C	-0.032	0.016	64	-2.001	<0.05*
	High +0.0°C	-0.046	0.015	64	-3.078	<0.01*

Table 6. Fixed factor type III ANOVA table for adult physiological parameters. Adults were

975	kept at Control,	Moderate or High (CO ₂ cross-factored	with +0.0°C	(28.5°C), +1.5	°C (30.0°C)
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Characteristic	Variable	DF	MS	F-value	p-value
Fulton's K	CO ₂	2	0.079	1.184	0.313
	Temperature	2	0.211	3.157	0.051
	CO ₂ :Temperature	4	0.024	0.356	0.839
	Residuals	54	0.067		
H.S.I.	CO ₂	2	0.016	0.374	0.689
	Temperature	2	0.041	0.971	0.385
	CO ₂ :Temperature	4	0.038	0.887	0.478
	Residuals	54	0.042		
G.S.I	CO ₂	2	3.365	2.266	0.114
	Temperature	2	11.655	7.848	<0.01*
	CO ₂ :Temperature	4	4.157	2.800	<0.05*
	Residuals	54	1.486		
17β-estradiol	CO ₂	2	1765838	1.044	0.361
	Temperature	2	10229356	6.050	<0.01*
	CO ₂ :Temperature	4	1584219	0.937	0.452
	Residuals	43	1690826		

976 or $+3.0^{\circ}C$ (31.5°C) water temperatures. Significant effects (p<0.05) are denoted by *.

977

- Table 7. Eigenvector values from factor analysis examining the trends in variance of gonadal
- 980 development in relation to plasma 17β-estradiol concentration in female *A. melanopus*. The
- 981 percentage of variation in plasma 17β-estradiol concentration explained by the first two
- 982 factors in given. Gamete stages that contributed >70% to the factors are bolded.

	Factor 1	Factor 2
Gonadal cell stage	48%	25%
Stage 1 Oogonia	-0.691	0.499
Stage 2 Perinucleolus	-0.793	0.170
Stage 3 Cortical alveolous	0.007	0.975
Stage 4 Early vitellogenic	0.891	0.011
Stage 5 Late vitellogenic	0.719	0.204

Table 8. Summary table of the reproductive and physiological characteristics measured, the predicted response of the stressors (elevated temperature and CO_2) and the observed effect of the stressors. + symbols represent positive effects of the treatment, - symbol represent a negative effect of treatment, = symbol represents no difference between treatment and control.

	Predicted response		Observed respo	onse
	Temperature	CO ₂	Temperature	CO ₂
Physical				
Characteristics				
Fulton's K	_	_	=	=
Hepatosomatic Index	-/+	_	=	=
Gonadosomatic Index	_	+	_	+
17β-estradiol	_	+	_	+
concentration				
Offspring				
Hatchling length	-	=	_	_
Yolk area	_	=	_	_
Reproductive				
characteristics				
Clutches produced	_	+	_	=
Interclutch interval	+	_	+	_

Eggs per clutch	—	+	=	=
Fertilization success	_	=	=	=
Egg area	—/+	=	-	-
Reproductive output	—/+	+	_	_
Successful clutches	-	=	=	=
Egg survival	_	=	=	=

992 Figure 1. The data collected at each stage of reproduction in *A. melanopus*, (a) reproductive 993 adults (b) clutches produced, (c) eggs collected form reproductive adults and (d) the resulting 994 hatchlings. Breeding pairs were kept at control, moderate or high CO₂ cross-factored with 995 either +0.0°C (28.5°C), +1.5°C (30.0°C) or +3.0°C (31.5°C) temperature treatments. 996 997 Figure 2. The total egg clutches produced per month across the breeding season for each CO_2 998 treatment at (a) +0.0°C (28.5°C), (b) +1.5°C (30.0°C) and (c) +3.0°C (31.5°C). Shown is the 999 total number of clutches produced in each treatment for that month. The total number of 1000 clutches produced in each temperature by CO₂ treatment group and the number of pairs that 1001 reproduced in each treatment group are shown on the figure. 1002 1003 Figure 3. The reproductive characteristics of breeding pairs kept at control, moderate or high 1004 CO_2 cross factored with either +0.0°C (28.5°C), +1.5°C (30.0°C) or +3.0°C (31.5°C) 1005 temperature treatments. The reproductive characteristics are: (a) the average number of 1006 clutches produced per pair for the breeding season, (b) the average number of eggs produced 1007 per clutch, (c) the average egg area (mm^2) , (d) the average reproductive output, being the 1008 average egg area per clutch multiplied by the number of eggs in the clutch to give an estimate 1009 of energy (mm^2) , (e) the average survival rate of the clutches produced and (f) the average egg survival. * denotes a treatment group that is significantly different from Control CO₂ 1010 1011 +0.0°C. 1012 1013 Figure 4. Offspring characteristics from parents kept at control, moderate or high CO₂ cross-1014 factored with either +0.0°C (28.5°C) or +1.5°C (30.0°C). No eggs survived to hatching in the

1015 moderate +1.5°C or the +3.0°C treatment groups, consequently they are not shown. The

- 1016 offspring characteristics were (a) hatchling standard length (mm) and (b) yolk area (mm²). *
- 1017 denotes a treatment group that is significantly different from control $CO_2 + 0.0^{\circ}C$.
- 1018
- 1019 Figure 5. Adult physiological condition and hormone concentrations at the end of the
- 1020 breeding season, (a) Fulton's K body condition index, (b) hepatosomatic index, (c)
- 1021 gonadosomatic index and (d) plasma 17β -E₂ concentrations of females. * represent groups
- that are significantly different.

Reproductive Measurements





Total clutches produced per month





Hatchling length (mm)

Yolk area (mm²)



Temperature treatment

Temperature treatment